

Runoff Water Quality from Highway Cut Slopes in Pyritic Rock Formations: Characterization and Potential Treatment



Tennessee Department of Transportation



Final Report

June 30, 2018

Cover page photo: Road cut through Chattanooga shale east bound I-640, Nashville, Tennessee

Photo by J. Schwartz, July 2014

DISCLAIMER

This research was funded through the State Research and Planning (SPR) Program by the Tennessee Department of Transportation and the Federal Highway Administration under RES# 2013-26, Research Project Titled: *Runoff Water Quality from Highway Cut Slopes in Pyritic Rock Formations: Characterization and Potential Treatment*.

This document is disseminated under the sponsorship of the Tennessee Department of Transportation and the United States Department of Transportation in the interest of information exchange. The State of Tennessee and the United States Government assume no liability of its contents or use thereof.

The contents of this report reflect the views of the author(s) who are solely responsible for the facts and accuracy of the material presented. The contents do not necessarily reflect the official views of the Tennessee Department of Transportation or the United States Department of Transportation.

Technical Report Documentation Page

| | | |
|---|--|---------------------------------|
| 1. Report No. RES2013-26 | 2. Government Accession No. | 3. Recipient's Catalog No. |
| 4. Title and Subtitle Runoff Water Quality from Highway Cut Slopes in Pyritic Rock Formations: Characterization and Potential Treatment | 5. Report Date June 30, 2018 | 6. Performing Organization Code |
| | 8. Performing Organization Report No. | |
| 7. Author(s) John S. Schwartz, Angle Palomino, and Qiang He | 10. Work Unit No. (TRAIS) | |
| 9. Performing Organization Name and Address University of Tennessee, Dept. of Civil & Environmental Engineering 851 Neyland Drive Knoxville, Tennessee 37996 | 11. Contract or Grant No. RES#2013-26 | |
| | 13. Type of Report and Period Covered Final Report June 1, 2013-May 31, 2018 | |
| 12. Sponsoring Agency Name and Address Tennessee Department of Transportation Long Range Planning Division James K. Polk Building, Suite 900 505 Deaderick Street Nashville, TN 37243-0334 | 14. Sponsoring Agency Code | |
| | 15. Supplementary Notes | |
| 16. Abstract <p>Road construction through pyritic (sulfidic) rock formations has the potential to cause episodic runoff acidification and impairment to aquatic biota when acid producing materials (APM) are exposed to precipitation. State highways departments (DOTs) across eastern United States in the Appalachia region have recognized the need to handle APM during road construction projects, including Pennsylvania, Virginia, and Tennessee. These state DOTs have developed testing, material handling, and disposal guidance. Although much is known about acid material drainage (AMD) chemistry, treatment options, and potential harmful impacts from it, little is known about the water quality generated from post-construction road cuts through APM, and what hydrogeological conditions may produce harmful levels of acidic waters containing sulfates and dissolved metals. The project objectives were to: 1) characterize the water quality from APM highway cut slopes to assess whether long-term environmental issues occur to receiving streams; 2) assess the potential effectiveness of on-site treatment applications at road cuts through a physical model experiment; and 3) by a review of AMD treatment literature assess the possibility of using passive treatment solutions for the level of acid pollution generated from road cuts in Tennessee. Characterization of runoff water quality from highway cut slopes through pyrite geology was conducted at ten monitoring stations in Middle and East Tennessee. The ten stations were located among five different surficial geologic formations; they were the Chattanooga Shale, Fentress Formation, Sandsuck Formation, Great Smoky Group (Anakeesta Formation), and the Snowbird Group: Roaring Fork Sandstone. The influence of various environmental factors was investigated on their effect on runoff water quality, which included: rock formation type and age of cut, cut slope and aspect, vegetation, and rainfall magnitude and intensity. Runoff pH and other chemical parameters varied widely between sites and within sites among different runoff events. Site mean pH ranged from 4.29 to 6.27, and the full sample event range was between 2.97 and 7.70. Site means for acid neutralizing capacity (ANC) ranged from -0.82 to 29.85 meq/L, and did not correlated per site with mean pH values. The fact that pH and ANC did not correlate suggests complex and unique biogeochemical processes are occurring among the study sites. Availability of base cations from the pyritic formation or an adjacent over- or under-burden sedimentary formation that contribute to the runoff greatly controls the ANC. Site means for sum of base cations (Ca²⁺, Mg²⁺, Na⁺, K⁺) ranged from 1.15 to 40.24 meq/L. Decreased pH can result from the displacement of base cations from the rock with Al complexes, where the base cations act as counter ions to electrochemically balance leached sulfate and other acid anions. Overall, rock formation type was the dominant controlling factor for water chemistry of road cut runoff. The findings</p> | | |

suggested that there are minimal effects from seasons, storm event magnitude, and road cut characteristics of slope and aspect. Also, it suggests the older the road cut the likelihood of acidic runoff is small and that planting vegetative cover at runoffs can be used as an effective management strategy to minimize acidic runoff. Useful information was obtained from an outdoor experimental study which consisted of constructing four container panels to place pyrite rock, and for three of the four panels install treatments to limit or prevent exposure to the rock, and subjected to both simulated and natural rainfall. The treatment types were: 1) soil/vegetation cover, 2) shotcrete, and 3) a geosynthetic membrane liner. One panel with exposed pyrite rock (no treatment) served as the experimental design control, and provided unique data on runoff chemistry and ion export from the bare rock surface. For the simulated rainfall experiment, pH increased from approximately 3.6 to above 6.0 within four hours. However, rapid oxidation occurred where the next day the initial pH was near 3.6 however reached pH values above 6.0 within two hours. Runoff ANC from the pyrite rock remained low during the simulations due to the lack of base cations to neutralize the acid anions. Mineral ions from the rock surface are quickly washed off as observed by the rapid decline in conductivity within the first hour of rainfall and basically diminished of ion source within two hours of rainfall. Shotcrete and a geosynthetic membrane liner were found to be effective to protect runoff water quality. The treatment consisting of soil/vegetation was more complex, and runoff quality needs further investigation. From the experiments, several roadside treatment options are available; they include: limestone beds and oxic/open channels, aerobic wetlands, settling ponds, bio-reactors, and pebble quicklime. Design of these treatment options are described in the guidance document “*Guidelines for Acid Producing Rock Investigation, Testing, Monitoring and Mitigation*” by Golder Associates.

| | | | |
|--|--|---|---------------------|
| 17. Key Words ACID MATERIAL DRAINAGE; ROAD CUT RUNOFF; WATER QUALITY; ACID PRODUCING MATERIALS | | 18. Distribution Statement No restriction. This document is available to the public at https://www.tn.gov/tdot/topic/longrange-research . | |
| 19. Security Classif. (of this report) Unclassified | 20. Security Classif. (of this page) Unclassified | 21. No. of Pages 175 | 22. Price \$0.00 |

Runoff Water Quality from Highway Cut Slopes in Pyritic Rock Formations: Characterization and Potential Treatment

Submitted by:

University of Tennessee, Knoxville
Department of Civil and Environmental Engineering

John S. Schwartz, Ph.D., P.E.
Angel Palomino, Ph.D.
Qiang He, Ph.D.



Prepared for:

Tennessee Department of Transportation
Materials & Tests Division, Geotechnical Engineering Section
Environmental Division, Ecology & Permits Office

Robert Jowers, P.E.
Daniel Reeves
Matt Richards



Final Report: June 30, 2018

Table of Contents

| | <u>Page</u> |
|---|-------------|
| Executive Summary | ix |
| Acknowledgements..... | xii |
| 1. Introduction..... | 1 |
| 1.1 Overview..... | 1 |
| 1.2 Project Objectives | 2 |
| 1.3 Scope of Work | 3 |
| 2. Characterization of Runoff Water Quality from Road Cuts through APM..... | 4 |
| 2.1 Introduction..... | 4 |
| 2.2 Methods..... | 6 |
| 2.2.1 Study Area | 6 |
| 2.2.2 Rainfall and Runoff Hydrology | 7 |
| 2.2.3 Water Sample Collections..... | 10 |
| 2.2.4 Water Quality Chemical Analysis | 11 |
| 2.2.5 Site Water Quality Characterization | 11 |
| 2.2.6 Statistical Analysis..... | 12 |
| 2.3 Results..... | 13 |
| 2.3.1 Road Cut Site Chemistry | 13 |
| 2.3.2 Environmental Factors Associated with Road Cut Site Chemistry | 18 |
| 2.3.3 Mass Export of Chemistry Parameters in Runoff from Road Cuts | 26 |
| 2.3.4 Potential Water Quality Impairment from Road Cut Site Chemistry | 26 |
| 2.4 Discussion and Conclusion..... | 29 |
| 3. Experimental Testing of Treatments for APM Highway Cut Slopes | 33 |
| 3.1 Introduction..... | 33 |
| 3.2 Methods..... | 34 |
| 3.2.1 Experimental Set-up..... | 34 |
| 3.2.2 Water Quality Chemical Analysis | 37 |
| 3.2.3 Data Analysis | 37 |
| 3.3 Results..... | 37 |
| 3.3.1 Simulated Rainfall Experiments | 37 |
| 3.3.2 Natural Rainfall Monitoring | 43 |
| 3.3.3 Relationships between Runoff Chemistry and Natural Rainfall Volumes | 45 |
| 3.4 Discussion and Conclusion..... | 50 |
| 4. Passive Water Quality Treatment Options at Road Cuts through APM..... | 54 |
| 4.1 Introduction..... | 54 |
| 4.2 Potential Treatment Options | 54 |
| 4.3 Cited Literature for Acid Rock Drainage Treatment | 57 |
| References: Chapters 2, 3 and 4..... | 61 |
| Appendices..... | 68 |

List of Figures

| | | |
|-------------|--|----|
| Figure 2.1 | Tennessee state map showing locations of the ten runoff monitoring stations (shown as white squares). Image Credits: <i>Google and Google Earth image overlay</i> | 7 |
| Figure 2.2 | Photos of Pinson et al. (2003) flow divider buckets for passive runoff collection at the monitoring sites: a) Nashville 2 and b) Ocoee 1 | 10 |
| Figure 2.3 | Scatterplots of sum of acid anions (Sum AA) and base cations (Sum BC) with storm event average intensity (RFIA). RFIA vs Sum AA ($p = 0.04$, $R^2 = 0.20$), RFIA vs Sum BC ($p = 0.04$, $R^2 = 0.19$)..... | 19 |
| Figure 2.4 | Boxplot of runoff chemistry for pH and ANC (meq/L) for the five different road-cut formations. Formations are: Great Smoky Group: Anakeesta Formation (Anakee) Chattanooga Shale (Chat), Fentress Formation (Fentr), Sandsuck Formation (Sand), and Snowbird Group: Roaring Fork (Snowb). | 21 |
| Figure 2.5 | Boxplots of runoff chemistry for conductivity (S/cm) and total hardness (mg/L) for the five different road-cut formations. Formations are Great Smoky Group: Anakeesta Formation (Anakee), Chattanooga Shale (Chat), Fentress: Formation (Fentr), Sandsuck Formation (Sand), and Snowbird Group: Roaring Fork Sandstone (Snowb)..... | 21 |
| Figure 2.6 | Boxplots of runoff chemistry for dissolved metals (mg/L) including Al, Cu, Mn, Fe, Si, Zn, Ba, Cd, and Ni for the five different road-cut formations. Formations are: Great Smoky Group: Anakeesta Formation (Anakee) Chattanooga Shale (Chat), Fentress Formation (Fentr), Sandsuck Formation (Sand), and Snowbird Group: Roaring Fork Sandstone (Snowb). | 22 |
| Figure 2.7 | Boxplots of runoff chemistry for conductivity ($\mu\text{S}/\text{cm}$) and total hardness (mg/L) for four different road-cut vegetative covers: none (None), grass (Grass), dense forest (Forest) and sparse forest (Stree). | 23 |
| Figure 2.8 | Boxplots of runoff chemistry for pH, ANC (meq/L), Sum AA (meq/L), and Sum BC (meq/L) for four different road-cut vegetative covers: none (None), grass (Grass), dense forest (Forest) and sparse forest (Stree)..... | 24 |
| Figure 2.9 | Boxplots of runoff chemistry for conductivity ($\mu\text{S}/\text{cm}$), total hardness (mg/L), pH, ANC (meq/L), Sum AA (meq/L), and Sum BC (meq/L) for three different road-cut ages: young, mid, and old..... | 25 |
| Figure 2.10 | Boxplots of runoff chemistry for total hardness (mg/L), pH, and ANC (meq/L), for four different road-cut aspects: east, west, north, and south. | 26 |
| Figure 3.1 | Schematic of the experimental test panel with pyritic rock. | 35 |
| Figure 3.2 | Photos of the experiment set-up on Middlebrook Pike, Knoxville, Tennessee showing Anakeesa Formation rock treatment trays (<i>upper right photo</i> – exposed rock and <i>lower right photo</i> – soil/grass), and the rainfall simulator (<i>lower left photo</i>). | 36 |
| Figure 3.3. | Runoff pH from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2). | 39 |
| Figure 3.4. | Runoff ANC (meq/L) from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2)..... | 39 |
| Figure 3.5. | Runoff sum of acid anions (meq/L) from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2).... | 40 |

| | | |
|--------------|--|----|
| Figure 3.6. | Runoff sum of base cations (meq/L) from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2).... | 40 |
| Figure 3.7. | Runoff sulfate (mg/L) from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2)..... | 41 |
| Figure 3.8. | Runoff conductivity ($\mu\text{S}/\text{cm}$) from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2)..... | 41 |
| Figure 3.9. | Runoff dissolved iron (mg/L) from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2)..... | 42 |
| Figure 3.10. | Runoff dissolved aluminum (mg/L) from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2).... | 42 |
| Figure 3.11. | Runoff pH from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain)..... | 46 |
| Figure 3.12. | Runoff conductivity ($\mu\text{S}/\text{cm}$) from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain)..... | 46 |
| Figure 3.13. | Runoff ANC (meq/L) from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain)..... | 47 |
| Figure 3.14. | Runoff sum of acid anions (meq/L) from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain)..... | 47 |
| Figure 3.15. | Runoff sum of base cations (meq/L) from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain)..... | 48 |
| Figure 3.16. | Runoff sulfate (mg/L) from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: shotcrete (Conc), and geoliner (Liner), and rain water (Rain)..... | 48 |
| Figure 3.17. | Runoff sulfate (mg/L) from natural rainfall (June – December 2016) for the treatment soil/vegetation (Grass),and rain water (Rain)..... | 49 |
| Figure 3.18. | Runoff dissolved iron (mg/L) from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain)..... | 49 |
| Figure 3.19. | Runoff dissolved aluminum (mg/L) from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain)..... | 50 |
| Figure 3.20. | Plots of storm event magnitude (depth, inch) from the pyrite rock control panel and chemistry for pH, conductivity ($\mu\text{S}/\text{cm}$), sulfate (mg/L), and dissolved iron (mg/L)..... | 52 |
| Figure 4.1. | Example photos of an open limestone channel (top) and a vertical flow wetlands (bottom). Photos from Paul Ziemkiewicz, West Virginia University, 2002..... | 56 |

List of Tables

| | | |
|------------|---|----|
| Table 2.1 | Study site highway locations in Tennessee, rock formations, and general site notes..... | 8 |
| Table 2.2 | Road cut site information on type, age of cut, cut slope, facing aspect, and vegetation. | 8 |
| Table 2.3 | Weather station locations used to estimate storm event volumes at road cut monitoring stations. Weather stations maintained by the Tennessee Valley Authority. | 9 |
| Table 2.4 | Road cut monitoring stations and the nearest TVA weather stations used to estimates storm event volumes. TVA Weather Station latitude and longitude locations defined in Table 2.3. Lat., long, and absolute (Δy , Δx) deviations summarized per road cut monitoring station. | 9 |
| Table 2.5 | NRCS rainfall-runoff equation parameters curve number (CN) and potential maximum retention (S) or each monitoring station for metric precipitation depths..... | 11 |
| Table 2.6 | Summary of water quality analyses, analytical chemical procedures, equipment, and method references. | 12 |
| Table 2.7 | Summary of toxic effects of pH, and aluminum and zinc concentrations on fish and benthic macroinvertebrates from published literature. | 13 |
| Table 2.8 | Summary of site chemistry by parameter. Site identifiers are: G1 is Grainger; J1, J2, & J3 are Jamestown; N1 & N2 are Nashville; O1, O2, & O3 are Ocoee; and S1 is Sevierville. Latitude and longitude coordinates are listed in Table 2.4. Site means and (standard deviations) and range. Appendix C contains the details of all site and storm event chemistry. | 14 |
| Table 2.9 | Summary of storm event characteristics for estimated volumes, and average and maximum intensities, and runoff volumes. Site means and (standard deviations) and range. Appendix B contains the details of all site and storm event data..... | 19 |
| Table 2.10 | Summary of runoff chemistry for the different formations along road-cut exposed to rainfall. Formations are Chattanooga Shale (Chat), Fentress Formation (Fentr), Sandstuck Formation (Sand), Great Smoky Group: Anakeesta Formation (Anakee), and Snowbird Group: Roaring Fork Sandstone (Snowb). Means and (standard deviations) provided per parameter and geologic formation. Letters ^{A, B, C, etc} indicate significant differences per chemical ($p \leq 0.05$). | 20 |
| Table 2.11 | Site mass loading export (g/day) of acid anions, base cations, and dissolved metals per storm event, and summarized by means, standard deviations (StDev), and range. Site identifiers are: G1 is Grainger; J1, J2, & J3 are Jamestown; N1 & N2 are Nashville; O1, O2, & O3 are Ocoee; and S1 is Sevierville. Latitude and longitude coordinates are listed in Table 2.4. | 27 |
| Table 3.1. | Linear regression relationships for the four 12-ft ² area panels, one control and three treatments. Runoff volume (∇) is in ft ³ , and rainfall depth (RD) is in inches. | 37 |

| | | |
|------------|--|----|
| Table 3.2. | Summary of runoff chemistry from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner). Water chemistry from the tank source was reported (Pool). Summary reported as the range (maximums and minimums)..... | 43 |
| Table 3.3. | Summary of runoff chemistry from natural rainfall from June to December 2016 for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner). Water chemistry of the rain water was also reported (Rain). Summary data reported as the average, standard deviation (StDev), and range (maximums and minimums)..... | 44 |
| Table 3.4. | Statistical results for runoff treatment/control/rain pairs using a t-test for chemical parameters from the natural rainfall events from June – December 2016. Control = pyrite rock (Rock), Treatments = soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain). | 45 |
| Table 3.5. | A summary of storm events for the natural rainfall monitoring between June – December 2016 consisting of storm event depth (inch) and estimated runoff volumes for the pyrite rock control panel (Rock) and three treatments soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner). | 51 |

Appendices

| | |
|---|-----|
| Appendix A: Photo Record of Road Cut Study Sites..... | 69 |
| Appendix B: Surficial Pyritic Rock Formations in Tennessee..... | 79 |
| Appendix C: Road Cut Monitoring Stations: Storm Event Rainfall Characteristics and Runoff Volumes..... | 84 |
| Appendix D: Water Chemistry for Monitoring Station Storm Events..... | 90 |
| Appendix E: Literature Review of Toxicity Thresholds to Acidification for Aquatic Biota..... | 112 |
| Appendix F: Pollutant Export Mass Loadings from Monitoring Station Storm Events..... | 122 |
| Appendix G: EPT XTRAM® HPL Specifications..... | 134 |
| Appendix H: Water Chemistry for Middlebrook Pike Experiment: Simulated Rainfall..... | 135 |
| Appendix I: Water Chemistry for Middlebrook Pike Experiment: Natural Rainfall..... | 143 |

List of Acronyms

| | |
|------------------|--|
| AA | Acid anions |
| AWWA | American Water Works Association |
| Anakee | <i>Geology</i> : Great Smoky Group: Anakeesta Formation |
| ANC | acid neutralizing capacity |
| ANOVA | Analysis of Variance |
| AP | Acid potential |
| AMD | Acid material drainage |
| APM | Acid producing material |
| BC | Base cations |
| Chat | <i>Geology</i> : Chattanooga Shale |
| CMC | Compound maximum concentrations |
| CN | Curve number (hydrology runoff parameter) |
| DEM | Digital elevation map |
| EDTA | Ethylenediaminetetraacetic acid titrimetric method |
| Fentr | <i>Geology</i> : Fentress Formation |
| gpm | Flow rate, gallons per minute |
| HDPE | High density polyethylene |
| HP | Horsepower |
| IC | Ion chromatograph |
| ICP-OES | Inductively-coupled plasma optical emission spectrometer |
| LC ₅₀ | Lethal concentration for 50% of the test organisms |
| LSD | Least significance difference |
| NP/NNP | Neutralization potential; Net neutralization potential |
| NRCS | Natural Resource Conservation Service |
| P | Precipitation / rainfall (depth units, cm) |
| P _e | Runoff volume (depth units, cm) |
| PPA | Potential peroxide acidity |
| psi | Pressure, pounds per square inch |
| RFV | Rainfall (storm) event volume (depth units, cm) |
| RFIA | Rainfall (storm) event average intensity (cm/hr) |
| RFIM | Rainfall (storm) event maximum intensity (cm/hr) |
| S | Potential maximum retention (hydrology runoff parameter) |
| Sand | <i>Geology</i> : Sandstuck Formation |
| Snowb | <i>Geology</i> : Snowbird Group: Roaring Fork |
| Stree | Vegetation class: sparse tree cover |
| TDEC | Tennessee Department of Environment and Conservation |
| TDOT | Tennessee Department of Transportation |
| TVA | Tennessee Valley Authority |
| USEPA | United States Environmental Protection Agency |
| USGS | United States Geological Survey |

Executive Summary

Runoff Water Quality from Highway Cut Slopes in Pyritic Rock Formations: Characterization and Treatment

Road construction through pyritic (sulfidic) rock formations has the potential to cause episodic runoff acidification and impairment to aquatic biota when acid producing materials (APM) are exposed to precipitation. State highways departments (DOTs) across eastern United States in the Appalachia region have recognized the need to handle APM during road construction projects, including Pennsylvania, Virginia, and Tennessee. These state DOTs have developed testing, material handling, and disposal guidance. Specifically, the Tennessee Department of Transportation (TDOT) has encountered APM on several projects in the past decade, which required projects to address special environmental and permitting issues. In order to assist TDOT when APM are encountered on a highway construction site, TDOT (2007) developed Special Provision 107L and standard drawings for material treatment. The document “*Guidelines for Acid Producing Rock Investigation, Testing, Monitoring and Mitigation*” was prepared by Golder Associates to provide technical assistance for APM assessment and management (TDOT 2007). In general, extensive research has been conducted on water quality and remediation of acid mine drainage from sulfidic geologic rock formations. Although much is known about acid material drainage (AMD) chemistry, treatment options, and potential harmful impacts from it, little is known about the water quality generated from post-construction road cuts through APM, and what hydrogeological conditions may produce harmful levels of acidic waters containing sulfates and dissolved metals. The project objectives were to: 1) characterize the water quality from APM highway cut slopes to assess whether long-term environmental issues could occur to receiving streams; 2) assess the potential effectiveness of on-site treatment applications at road cuts through a physical model experiment; and 3) by a review of AMD treatment literature assess the possibility of using passive treatment solutions for the level of acid pollution generated from road cuts in Tennessee.

Characterization of runoff water quality from highway cut slopes through pyrite geology was conducted at ten monitoring stations in Middle and East Tennessee. The ten stations were located among five different surficial geologic formations; they were the Chattanooga Shale, Fentress Formation, Sandsuck Formation, Great Smoky Group (Anakeesta Formation), and the Snowbird Group: Roaring Fork Sandstone. The influence of various environmental factors was investigated on their effect on runoff water quality, which included: rock formation type and age of cut, cut slope and aspect, vegetation, and rainfall magnitude and intensity. In addition, the runoff water quality was assessed to its potential toxic impacts to aquatic biota.

Runoff pH and other chemical parameters varied widely between sites and within sites among different runoff events. Site mean pH ranged from 4.29 to 6.27, and the full range from individual sample events was between 2.97 and 7.70. Site means for acid neutralizing capacity (ANC) ranged from -0.82 to 29.85 meq/L. Mean ANC did not correlate with mean pH values per site, which suggests complex and unique biogeochemical processes occur among the study sites. Where runoff flows over a sedimentary formation adjacent to the pyritic formation base cations (Ca^{2+} , Mg^{2+} , Na^+ , K^+) enter into the runoff greatly controlling ANC. Site means for sum of base cations (Sum BC) ranged from 1.15 to 40.24 meq/L. Decreased pH can result from the displacement of base cations from the rock with Al complexes, where the base cations act as counter ions to electrochemically balance leached sulfate and other acid anions. The lack of available base cations can decrease runoff pH. Other than the dominant oxidation of pyrite

(FeS₂) exposed to water and air, other key geochemical processes affecting pH and ANC, and the overall water quality, includes desorption/adsorption of sulfate ions, hydrolysis and precipitation of dissolved metals, silicon and other mineral weathering, cation/metal exchange (Cu, Cd, Zn, Mn, others), and aluminum dissolution and mobilization. The elevated levels of observed Fe and Al demonstrated metal hydroxides likely formed in the runoff because of the differences between total and dissolved fractions. The mean dissolved concentrations for Fe and Al ranged from 0.13-1.63 mg/L and 0.33-39.00 mg/L, respectively. The overall wide range of chemical parameters and concentrations measured also suggests that other environmental factors may play a role in runoff chemical transformations. External to the road cut rock, surface nitrogen and organics may influence runoff water quality through nitrogen mineralization and nitrification, formation of organic acids, or sulfur/metal adsorption onto organics. These non-pyritic surface processes can affect acidity and proton availability, which in turn influences the runoff pH. The water quality characterization of each road cut site was affected by a unique suite of biogeochemistry processes and controlled by local dominant environmental factors. The factors were found to support assessing risk of harmful runoff acidification.

Geologic rock formations appeared to be the dominant controlling factor for water chemistry of road cut runoff for based on the ten study locations. The mean pH was lowest for the Anakeesta Formation (pH = 4.29), and its mean ANC of 1.83 meq/L was relatively low compared to other sites. Compared with the other formations, the Anakeesta formation exhibited lower concentrations of acid anions (sulfate) and base cations, but higher concentrations of dissolved metals including Al, Fe, Cu, Mn, Si, Cd, and Zn. The Chattanooga Shale and the Roaring Fork Sandstone had mean pH values below 6.0 (5.41 and 5.88, respectively). The runoff chemistry from the Chattanooga Shale was much different than all other formations. Runoff from the Nashville I-840 Chattanooga shale sites contains very high levels of acid anions, base cations, and dissolved metals. Site averages for conductivity exceeded 2,500 μ S/cm. The Nashville sites were fairly new road cuts and they appear to be weathering at a high rate exporting ions in the runoff, in particular higher concentrations of dissolved silicon was observed compared with the other sites. Overall runoff from the Chattanooga Shale contained relatively high levels of dissolved Al, Ba, Cu, Zn, and Ni compared to other rock formations. In general, the runoff chemistry of the Roaring Fork Sandstone was more similar to the other geologic formations. The Fentress and Sandsuck formations had mean pH values above 6.0. The runoff chemistry from these two formations was similar to the others except for the Chattanooga Shale as noted above. With the observed differences of dissolved metals among the rock formations it suggests the basic mineral composition of the local rock appears to greatly influence runoff chemistry from hydrolysis and precipitation of dissolved metals, silicon and other mineral weathering, and cation/metal exchange.

In addition to geologic rock formations, vegetation cover or the lack thereof appeared to influence runoff chemistry from the road cut study sites to some degree. With no vegetative cover with exposed bare rock, runoff ion concentrations were much greater compared to sites with grass and/or tree cover. The highest median pH among the cover types was with trees at 6.4. This outcome suggests that the availability of organic matter from the trees may be playing a role in metal/sulfate adsorption, and sulfur, nitrogen, and cation cycles. Tree cover also protects the rock surface from low intensity rainfall event per leaf interception and evaporation. Runoff chemistry was not correlated with rainfall volume and storm maximum intensity, but average intensity was found to influence the export of sulfate, sum of acid anions (Sum AA), and Sum BC indicating higher intensities have the potential to flush ions from the site. Seasons did not have any significant effect on runoff chemistry among the study sites. Though our study did

not find many relationships with season and weather-related controls, some evidence indicates these variables could influence runoff chemistry from APM. More recently constructed road cuts (young) were found with runoff chemistry with higher concentrations of Sum AA, Sum BC, and dissolved metals. In particular, 'young' sites were higher in dissolved Al and Si indicating mineral weathering rates to be greater compared to the middle and old age sites. The difficulty in interpreting the potential effect of the factors is that data co-varied with the more dominant factors of geologic formation and vegetation. More data from a larger set of study sites would be needed to better statistically evaluate these factors. Overall, the findings do suggest that there are minimal effects from seasons, storm event magnitude, and road cut characteristics of slope and aspect. Also, it suggests the older the road cut the likelihood of acidic runoff is small and that planting vegetative cover at runoffs can be used as an effective management strategy to minimize acidic runoff.

Useful information was obtained from an outdoor experimental study which consisted of constructing four container panels to place pyrite rock, and for three panels different treatments were installed to limit or prevent exposure to the rock. Treatment types were: 1) soil/vegetation cover, 2) shotcrete, and 3) a geosynthetic membrane liner. The panels were subjected to both simulated and natural rainfall. One panel with exposed pyrite rock (no treatment) served as the experimental design control, and provided unique data on runoff chemistry and ion export from the bare rock surface. For the simulated rainfall experiment, pH increased from approximately 3.6 to above 6.0 within four hours. However, rapid oxidation occurred where the next day the initial pH was near 3.6 however reached pH values above 6.0 within two hours. Runoff ANC from the pyrite rock remained low during the simulations due to the lack of base cations to neutralize the acid anions. The Sum AA consists primarily of sulfate, and the rapid decrease in SO_4^{2-} concentrations were measured. Mineral ions from the rock surface are quickly washed off as observed by the rapid decline in conductivity within the first hour of rainfall and basically diminished of ion source within two hours of rainfall. This rapid decline in dissolved metal ions was measured for all metal, but importantly observed with iron and aluminum. These two metal ions are the most relevant to pyrite oxidation and the potential for surface water toxicity, however all the metal ions were found to decline throughout the rainfall simulation period. The dissolved aluminum concentration dropped below 0.2 mg/L with two hours.

During the natural rainfall monitoring from the pyrite rock control, the runoff pH remained low between 3.00 and 3.74 during this six-month period indicating that freshly exposed pyrite geology will take a longer time period to reach a condition where the surface pyrite has completely oxidized. Sulfate export from the rock surface was the major driver for runoff acidification. Elevated levels of dissolved Fe and Al were observed for the runoff from the exposed rock caused by pyrite oxidation and Anakeesta shale weathering. A dilution effect was also observed where pH increased with storm event size (rainfall depth), and conductivity, sulfate, and dissolved iron decreased. With the information also from the simulated rainfall experiments it appears that there is a limited availability of sulfate to cause runoff acidification. This experiment investigated water quality changes from exposed pyritic rock only; in contrast the ten-site characterization study included the potential influence of other geologic formations and site soils adjacent to the exposed pyritic formation, which can contribute base cations to runoff reducing acidification.

Potential water quality impairments of runoff from road cuts were assessed by comparing site chemistry to TDEC regulatory limits on specified chemicals, in addition to published toxicity exposure limits. The dominant concern in runoff that has been in contact with geologic rock formations considered as APM is the chemical parameters resulting from geochemical

acidification processes. These parameters primarily include pH, ANC, and dissolved Al. Other dissolved metals can be a concern depending on hardness where lower hardness concentrations increase toxicological effects. Site averages for runoff pH ranged from acidic to neutral waters where five of the ten sites exceeding a pH of 6.0 and four sites had average pH values below 5.0. Within sites, pH values varied widely, for example the Jamestown 2 site had an average pH of 4.48, but the range was from 3.45 to 7.07. The sites with a pH less than 6.00 are below the state's water quality standards, which pH values are to be between 6 and 9. There is no state water quality standard for dissolved Al, however literature indicates that concentrations should not exceed 0.2 mg/L. Based on a site average, all road cut sites except the Jamestown 3 site exceeded this toxicity threshold. Dissolved Al in runoff was very high at the Nashville I-840 site, indicating a unique geochemical condition under rock/soil acidification. Though the road cut dissolved Al concentrations were above that threshold, runoff events are generally short and runoff is diluted as it comeslingles with the receiving stream. Though some exceedances were observed with pH and dissolved Al, runoff water quality was at levels easily treated using best available technology.

Guidelines on addressing the environmental impacts from acid rock drainage (ARD) at road cut construction sites were developed and published by TDOT with Golder Associates (TDOT 2007). The guidance document "*Guidelines for Acid Producing Rock Investigation, Testing, Monitoring and Mitigation*" was comprehensive and focused on testing, handling, and disposal during highway construction. The TDOT (2007) document also summarized water treatment options for post-construction mitigation of the site runoff. Passive treatment systems are best suited for ARD with low acidity ($< 800 \text{ mg CaCO}_3 / \text{L}$) and low flow rates ($< 50 \text{ L/s}$). Therefore, mass loadings from road cut sites should be less than about 100 to 150 kg $\text{CaCO}_3 / \text{day}$. Relevant to highway runoff from road cuts through pyritic rock are passive systems. In the TDOT (2007) document, several passive treatments were identified as preferred treatment methodologies because of ease of construction, low maintenance, and treatment longevity. They were: limestone beds and oxic/open channels, aerobic wetlands, settling ponds, bio-reactors, and pebble quicklime. These treatments are viable options for on-site passive treatment if the road cut site conditions warrant their construction.

Acknowledgments

This research was funded through the TDOT Research Division by Grant RES#2013-16. The project was managed by Robert Jowers within the Materials and Tests Division with the assistance of Daniel Reeves and Samuel Williams, and Matt Richards within the Environmental Division with the assistance of former staff members Deedee Kaufman and Lina Khoury. Many University of Tennessee students helped with the extensive field and laboratory work needed for this project, and they included: Cory Julien, Brian Williamson, Adrian Gonzalez, Will Simmons, Brandy Manka, Michael Walton, Will Luke, Brett Beeler, Joshua Benavidez, Badal Mahalder, and Max Carter. We appreciate the coordinated work effort with the USGS Nashville Office from Michael Bradley and Tom Byl.

1. INTRODUCTION

1.1 Overview

Exposed pyrite geologic formations with acid producing material (APM) combined with rainfall (water) generates sulfuric acid and can be transported as runoff (Pye and Miller 1990; Fox et al. 1997; Rimstidt and Vauhgan 2003; Hammarstrom et al. 2005). A pyritic rock or soil formation may produce acid when excavated and exposed to air, water, and mediated by Thiobacillus bacteria. Sulfide minerals, most commonly pyrite, produce this acid as a byproduct of the breakdown of the sulfur. This can cause environmental issues where impact can occur from pH change and the material may require special handling and procedures. APM can be found statewide within Tennessee. However, it is predominately found east of the Tennessee River in shales, sandstones, and some siltstones. In Region 1 the Anakeesta formation, the Chattanooga shale, as well as many Pennsylvanian formations contain APM. In Region 2, APM is often associated with Precambrian formations of the Ocoee Supergroup, Devonian Chattanooga Shale, the other Pennsylvanian-age shales. In Region 3 acid producing materials are often found in the Chattanooga Shale.

On highway construction sites, naturally occurring potentially APM may be found in any soil or rock horizon on any project type. Tennessee Department of Transportation (TDOT) has encountered APM on several projects in the past decade, in which these projects present special environmental and permitting issues. In order to assist TDOT when APM are encountered on a highway construction site, TDOT developed Special Provision 107L and standard drawings for material treatment. The document “*Guidelines for Acid Producing Rock Investigation, Testing, Monitoring and Mitigation*” was prepared by Golder Associates to provide technical assistance for APM assessment and management (TDOT 2007). Pre-construction requires sufficient drilling to identify, sample and test historically problematic and potentially problematic formations shall be completed in general accordance with the document (Byerly 1990, 1996). Drilling must be sufficient to show the horizontal and vertical extent of problem layers on the project identifying areas that must have special treatment and provide quantity calculations. During the project design and construction phase, TDOT routinely uses the document for specifying proper handling of APM spoils from highway projects, consisting of removal, disposal, blending, capping, and/or encapsulating APM.

For highway construction projects with potential APM, an acid-based accounting of the samples is required. TDOT maintains a laboratory testing contract with firms capable of providing the appropriate tests. This includes an assessment of the paste pH, Acid Potential (AP), Net Neutralization Potential (NNP) per calcium carbonate deficiency or net acid-base account value, as well as tests of total sulfur and pyritic sulfur (Sobek et al. 1978). All samples are assessed for whether or not they are representative of the material out in the field. Multiple parameters may be needed in order to assess whether or not a soil or rock can produce acid in the field. Additionally, site assessments of the same material placed in the field may also need to be completed or addressed. This approach to environmental mitigation of APM is expensive, thus driving the need to better understand the potential extent of the environmental problem.

Runoff from these cut slopes may enter nearby streams, which can become acidified and potentially harm aquatic life (Huckabee et al. 1975; Mathews and Morgan 1982; Kuchen et al. 1994). Based on numerous acid mine drainage and acid rain studies, many references have

described the potentially harmful effects of stream acidification on biota. Most toxicological studies focus on the acute lethal end-point. However, sub-lethal stress on trout can occur from episodic acidification events, and the degree of biota stress is dependent on the extent of pH change during stormflow, and the duration and frequency of events (Gagen et al. 1993; Baldigo et al. 2009; Neff et al. 2009). In general, toxic effects of pH on fish and macroinvertebrates can be classified as follows: 1) slight impairment = pH 5.5 to 6.4; 2) moderate impairment = pH 5.0 to 5.5; 3) severe impairment = pH 4.0 to 5.0; and 4) lethal = pH < 4.0. Stream with acid neutralizing capacity (ANC) less than 0 $\mu\text{eq/L}$ are considered acidic (Wigington et al. 1996). Tennessee Department of Environment and Conservation (TDEC) recommends an ANC greater than 50 $\mu\text{eq/L}$ during baseflow unless evidence suggests streams naturally occur below this concentration. Although pH and ANC provide water quality targets for stream acidification, these constituents may be surrogate for the toxic effects of dissolved aluminum. Dissolved aluminum in the form of inorganic monomeric aluminum (Al_{IM}) is regarded as the most toxic dissolved metal for fish and macroinvertebrates in acidified stream waters (Neville and Campbell 1988). Fish gill ion transport is disrupted by replacing needed calcium on gill surfaces with increased Al_{IM} concentrations (Booth et al. 1988; Ingersoll et al. 1990). Driscoll et al. (2001) suggests that Al_{IM} concentrations above 2.0 $\mu\text{mol/L}$ as an appropriate threshold for toxicity. TDEC water quality criteria per Chapter 004-40-03 require a pH between 6.0 and 9.0, and no pH change over 1.0 unit within a 24-hour period. Although much is known about the harmful impacts from acid mine drainage, little is known about the water quality generated from post-construction road cuts through APM, and what hydrogeological conditions may produce harmful levels of acidic waters containing sulfates and dissolved metals.

As noted above, road cuts occur through different geologic rock formations (e.g., Anakeesta metamorphic rock, Chattanooga shale) and varying orientations of inclined rock layers. The export of acidic waters from road cuts may be dependent on the design with respect to whether cut slopes are vertical or near-vertical, and length of exposure. The type of rock, soils, and vegetation also controls sulfate export from drainage surfaces, large or small (Cai et al. 2010, 2011a,b; Neff et al. 2012). A basic study on the runoff chemistry and hydrological export of sulfate and dissolved metals can provide useful information on the extent of the problem and if at levels potentially causing environmental harm. In addition, this information can provide the necessary data to design various treatment options.

Various on-site and permanent treatment options are available for APM runoff, including limestone channels and constructed wetlands (Byerly 1990). Performance of treatment options is not well known from highway cut slopes because of a lack of information on cut slope runoff and water quality. Knowing the amount of sulfate and protons [H^+ , pH] generated from exposed highway cut slopes with different local geologies and geotechnical engineering designs is critical for designing successful treatment systems. Over-designed systems could increase construction costs, and under-designed systems could harm the environment. More detailed information of how much acid runoff is generated is especially needed in environmentally sensitive areas.

1.2 Project Objectives

The project objectives were to: 1) characterize the water quality of AMD from highway cut slopes to assess whether long-term environmental issues occur to receiving streams; 2) assess the potential effectiveness of on-site treatment applications at road cuts through a physical model experiment; and 3) assess the possibility of using passive treatment solutions to treat expected

levels of acid pollution from road cut sites. Monitoring sites of different hydrogeological conditions were selected, sampled for runoff water, and characterized for chemical properties, which could be compared with federal and state water quality standards. Physical model experiment provides information on the runoff water quality for three treatments: soil/vegetation, shotcrete, and geoliner, amendments compared to an untreated control. With information on the runoff chemistry, it provides the basis to assess what possible passive treatment options are available to reduce acidity and transport of other harmful pollutants, i.e., dissolved monomeric aluminum. This research is proposed to examine more cost efficient and environmentally safe ways to deal with APM from cut slopes.

It is noted that the U.S. Geological Survey (USGS) also conducted concurrently an investigation of AMD water quality from road cuts through APM. Their investigation examined water from rock seams and examining micro-scale biogeochemistry. The USGS generated a bibliography for acid-rock drainage and selected acid-mine drainage issues related to acid-rock drainage from transportation activities (Bradley and Worland 2015). The University of Tennessee at Knoxville (UTK) collaborated with the USGS staff on data exchange or other supportive activities.

1.3 Scope of Work

Following the project's three objectives, the scope of work included:

- Section 2.0: Characterization of Runoff Water Quality from Road Cuts through APM
- Section 3.0: Experimental Testing of Treatments for AMD Highway Cut Slopes
- Section 4.0: Passive Water Quality Treatment Options at Road Cuts through APM

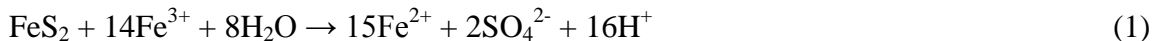
Details of the study methods for each scope of work are described in the individual sections, as listed above. Results and conclusions are presented per section. The Executive Summary serves as the overall summary and project conclusions.

2. Characterization of Runoff Water Quality from Road Cuts through APM

2.1 Introduction

Road construction through pyritic (sulfidic) rock formations has the potential to cause episodic runoff acidification and impairment to aquatic biota when acid producing materials (APM) are exposed to precipitation (Huckabee et al. 1975; Mathews and Morgan 1982; Morgan et al 1982; Kuchen et al. 1994; DeNicola and Stapleton 2002). In addition to water quality impacts, acidic runoff with high sulfate levels can degrade highway construction materials (concrete and metals), and effect soil and rock stability of road cut fill materials through excessive ion dissolution and/or crystal development causing material upheaving (Miller et al 1976; Vear and Curtis 1981; Berube et al. 1986; Pye and Miller 1990; Sahat and Sum 1990; Byerly 1996; Ji et al. 2006). State highways departments in the eastern United States have recognized the need to handle APM during road construction projects, including Pennsylvania, Virginia, Tennessee, and others in which they have developed testing, material handling, and disposal guidance (Ammons et al. 1990; Byerly 1996; Orndorff and Daniels 2002; Hammarstrom et al. 2005; Siceree 2006; Scheetz and Ellsworth 2007; TDOT 2007). Recent APM guidance for highway construction has greatly reduced severe environmental problems that have resulted in the past such as fish kills and mineral-stained waters; however little is known about post-construction runoff water quality. This study focused on characterizing runoff water quality from existing road cuts through sulfidic rock formations in East and Middle Tennessee.

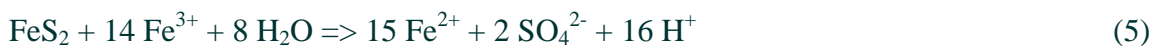
Extensive research has been conducted on water quality and remediation of acid mine drainage from sulfidic rock formations (Bigham and Norstrom 2000; Baker and Banfield 2003; Bird 2003; Rimstidti and Vauhgan 2003; Grande et al. 2004; Sracek et al. 2004; Johnson et al. 2005). This acid mine drainage research provides the fundamental background on the biogeochemical processes that occurs with groundwater seeps and surface runoff, including the microbial mediation of sulfide oxidation on pyrite rock surfaces (Nordstorm 2000; Fowler et al. 2001; Rohwerder et al. 2003; Kock and Schippers 2008; Hallberg 2009). Fundamentally, pyrite (FeS_2) reacts with water and oxygen where sulfide oxidation and ferrous iron (Fe^{2+}) is converted to ferric iron (Fe^{3+}) as follows:



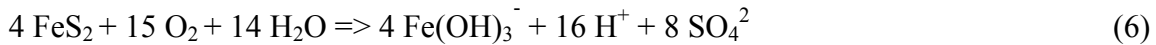
The rate of reaction in Equation 3 is pH dependent, where at low pH values around 2 to 3 the rate is very slow and no bacteria can assist in the production of protons (EnviroSci Inquiry 2011). At pH values around 5 the reaction is much faster. Ferrous iron may be converted to the commonly yellow-orange flock termed yellowboy in stagnant pools of water:



The above reaction is the hydrolysis of iron, which is pH dependent. Solid formation and precipitation occurs when the pH is above 3.5. This reaction in itself provides acidity. Another reaction which creates a self-propagating reaction by the oxidation of additional pyrite to ferric iron (Equation 5). This reaction is very rapid and continues until either ferric iron or pyrite is depleted.



The overall summary reaction, which included the creation of yellowboy is as follows:



The loose porous rust $\text{Fe}(\text{OH})_3$ can slowly transform into a crystallized mineral form written as $\text{Fe}_2\text{O}_3 \cdot \text{H}_2\text{O}$, the familiar red-brown staining. This red-brown staining can be observed on exposed road cut through pyritic rock formations.

The basic research on environmental geochemistry established from studying the acid mine drainage provides critical information interpreting the potential effects of water quality from post-construction road cuts through pyritic formations. A few studies examined runoff water quality from exposed pyritic rock surfaces other than road cuts, such as landfills, natural slopes, and excavated APM (Miller et al. 1976; Morgan et al. 1982; Fox et al. 1997; Igarishi and Oyama 1999). These studies found water to have high levels of free acidity ($\text{pH} < 5$), and elevated levels of dissolved metals (Al, Cu, Fe, Mn, and Zn). Sufficient availability of oxygen formed yellowboy and reddish-brown iron precipitates. One of the first papers published on runoff water quality from road cut through pyritic rock was by Cendreo et al. (1977) in Spain. Their study measured pH and %S (percent sulfur by weight) at eight road cuts observing a wide range of pH values from 2.5 to 8.0, and %S from trace amounts to 0.72%. As expected with the weathering of pyrite, runoff pH above 5 to 6 were directly correlated with %S content above 0.2%. Cendreo et al. (1997) also noted poor vegetation cover occurred at highly acidic sites. Adams et al. (1999) studied one road cut site near Buchanan, Georgia through a pyritic schist, and the receiving stream Kiser Creek. Springs and seeps from the rock cut site kept water flowing in the drainage ditch year-round, and pH measures ranged from 2.4 to 6.9, and averaged 3.8. This road-side water contained sulfate concentrations between 62 and 195 mg/L, and dissolved iron between 0.13 and 0.18 mg/L. The pH values of waters within all stream monitoring locations downstream of the road-cut runoff effluent were above 6.0.

Orndorff and Daniels (2004) conducted a survey of acid material drainage (AMD) from road cuts through the state of Virginia, examining runoff and rock chemistry for 25 sites. The focus on this study was classifying the potential severity of acidic runoff from different geologic formations with pyrite using the potential peroxide acidity (test), total S, and “fizz test” (Sobek et al 1978) on rock samples. Water samples were collected in a dug shallow well at the road cut base or an available ditch for 10 of the 25 sites. Orndorff and Daniels (2004) found water chemistry to vary widely among these ten sites. They found pH to range from 2.5 to 7.1 (average pH = 4.54) and total sulfur ranged from 7 to 543 mg/L. Dissolved metals also varied to a large extent among sites and even from the same site. Dissolved Fe ranged between < 0.1 to 249.3 mg/L with most concentrations below 1.0 mg/L, dissolved Al between < 0.1 to 254.5 mg/L with most concentrations below 20 mg/L, dissolved Cu between < 0.1 to 0.29 mg/L, and dissolved Zn between < 0.1 to 10.4 mg/L. These wide ranges in water chemistries indicate multiple environmental factors apparently control the runoff water quality from road cut sites.

Runoff from road cuts through pyritic rock can be potentially harm to aquatic life because waters become acidified and may contain toxic dissolved metals (Huckabee et al. 1975; Bacon and Mass 1979; Mathews and Morgan 1982; Morgan et al. 1982; Kuchen et al. 1994). Based on numerous acid mine drainage and acid rain studies, many references have described the potentially harmful effects of stream acidification on biota (Driscoll et al. 1980; Fromm 1980; Exley et al. 1991; DeLonay 1993; Hermann et al. 1993; Simonin et al. 1993; Harvey and Jackson 1995; Newman and Dolloff 1995). Most toxicological studies associated with acidified waters

focus on the acute lethal end-point of pH or dissolved Al. However, sub-lethal stress on fish can occur from episodic acidification events, and the degree of biota stress is dependent on the extent of pH change during stormflow, and the duration and frequency of events (Caroline et al. 1992; Gagen et al. 1993; MacAvoy and Bulger 2004; Baldigo et al. 2009; Neff et al. 2009). In general, toxic effects of pH on fish and macroinvertebrates can be classified as follows: 1) slight impairment = pH 5.5 to 6.4; 2) moderate impairment = pH 5.0 to 5.5; 3) severe impairment = pH 4.0 to 5.0; and 4) lethal = pH < 4.0. Although pH provides water quality targets for stream acidification, pH may be surrogate for the toxic effects of dissolved Al. Dissolved Al in the form of inorganic monomeric aluminum (Al_{IM}) is regarded as the most toxic dissolved metal for fish and macroinvertebrates in acidified stream waters (Neville and Campbell 1988). Inorganic aluminum increases with decreases in pH and predominate in waters below pH 5.0 (Driscoll and Postek 1995; Sullivan and Cosby 1998). Driscoll et al. (2001) suggests that Al_{IM} concentrations above 0.2 mg/L as an appropriate threshold for toxicity. Like H^+ (pH) toxicity, the primary mechanism of Al toxicity is disturbance of gill ion transport (Booth et al. 1988) when pH is 4.2–4.8 (most severe at pH 4.5), and asphyxia when pH is 5.5–6.4 (most severe at pH 6.1) (Neville and Campbell 1988). In general toxic metal ions including dissolved Al compete with metal cations (particularly Ca^{2+}) on fish gills (Di Toro et al. 2001) and facilitate greater loss of critical blood ions (mostly Na^+). In waters of low hardness, less dissolved Ca^{2+} is available and fish are more vulnerable to Na^+ ion loss induced by elevated pH and dissolved Al (Cleveland et al. 1991; Ingersoll et al. 1990). Although much is known about the harmful impacts from acid mine drainage, little is known about the water quality generated from post-construction road cuts through APM, and what hydrogeological conditions may produce harmful levels of acidic waters containing sulfates and dissolved metals.

The study objectives were to: 1) characterize the runoff water quality from existing road cut through pyritic rock formations and APM to assess acidity conditions, export of dissolved metals, and other chemical parameters; 2) investigate environmental factors that may be influencing the water chemistry of road cut runoff, including geological rock formation, vegetative cover, and road cut type, age, slope, and aspect; and 3) assess whether runoff water quality may potentially cause impairment, a stressed environment for aquatic biota. This study is unique compared to the limited previous published articles on road-cut water quality, in that the chemical parameters analyzed only included pH and sulfur, and Orndorff and Daniels (2003) measuring a few dissolved metals. No study compared anion-cation balances and a comprehensive suite of water quality parameters. In addition, environmental factors that possible can affect runoff chemistry other than rock type had not be investigated to date.

2.2 Methods

2.2.1 Study Area

Runoff from highway cut slopes through pyrite geology was collected at ten monitoring stations in Middle and East Tennessee (Figure 2.1). The ten stations were located among five different surficial geologic formations; they were the Chattanooga Shale, Fentress Formation, Sandsuck Formation, Great Smoky Group (Anakeesta Formation), and the Snowbird Group: Roaring Fork Sandstone (Table 2.1). Site photos are shown in Appendix A, and state-wide surficial geology maps for pyritic rock formations are in Appendix B. Additional site information on the road cuts was compiled in Table 2.2 consisting of formation type and age of cut, cut slope and aspect, and vegetation.



Figure 2.1. Tennessee state map showing locations of the ten runoff monitoring stations (shown as white squares). Image Credits: *Google and Google Earth image overlay.*

2.2.2 Rainfall and Runoff Hydrology

Rainfall volumes per storm event and study site were estimated from Tennessee Valley Authority (TVA) published data. TVA weather station data were used because installation of full weather stations at each site was cost prohibitive. Tables 2.3 and 2.4 summarize the TVA weather stations locations in latitude and longitude, and the weather stations used to compile rainfall volumes per road cut monitoring station. For each storm event in which water samples were collected, an estimate of event rainfall volume (in units of depth) was computed using a weighted approach based on distance absolute deviations. An absolute deviation represents the linear distance in lat./long. degrees between a weather station and the monitoring site.

Runoff volumes from the road cut monitoring sites were collected from rock walls using plastic roof gutters (Figure 2.2). Roof gutters were attached with stainless steel bolts and any space between the rock wall and the gutter were filled with a synthetic polymer foam and/or caulk adhesive. Gutters were sloped at least with a 2% grade into a downspout, which collected water was directed to passive collection systems consisting of two or three flow divider buckets and a terminal collection bucket, all of which were 5-gal (18.9 L) in size (Pinson et al. 2003). Based on Pinson et al. (2003), the flow dividers consisted of a stainless-steel circular crown containing 22.5° V-notch weirs, with the crown screwed onto the bucket (Figure 2.2). The first and second dividers consisted of 12 V-notches in order to handle the high initial flow rate, whereas if used the 3rd divider had 24 notches. Once the bucket completely filled, water and sediment overflow was evenly divided among the V-notches, and flow from a single notch was

Table 2.1. Study site highway locations in Tennessee, rock formations, and general site notes.

| Site Name | County | Latitude | Longitude | Highway | Formation | General Site Notes |
|---------------|------------|---------------|---------------|---------|------------------------------|---|
| Grainger 1 | Grainger | 36°21'1.52"N | 83°20'16.76"W | SR-32 | Chattanooga Shale | Near Bean Station; site – contains degraded erosion control mesh behind a wire mesh. |
| Jamestown 1 | Fentress | 36°29'49.48"N | 84°58'02.75"W | SR-28 | Fentress Formation | Gradual slope; recent constructed road cut; vegetated with grass. |
| Jamestown 2 | Fentress | 36°29'42.95"N | 84°58'05.10"W | SR-28 | Fentress Formation | Same conditions as the Jamestown 1 Site |
| Jamestown 3 | Fentress | 36°26'28.27"N | 84°57'46.52"W | SR-52 | Fentress Formation | Vertical cut, older road cut; mature vegetation in places. |
| Nashville 1 | Williamson | 35°48'52.44"N | 86°58'30.49"W | I-840 | Chattanooga Shale | High cut, east of SR 246. Dc on road cut shelf, multiple seeps with various sources - Mfp, Dc, Ou formations. |
| Nashville 2 | Williamson | 35°49'17.21"N | 86°57'11.81"W | I-840 | Chattanooga Shale | East of SR 246. Dc formation at top of road cut. Multiple seeps. |
| Ocoee 1 | Polk | 35° 7'13.81"N | 84°34'5.49"W | SR-30 | Sandsuck Formation | Near the Clear Creek Trailhead, Vertical cut |
| Ocoee 2 | Polk | 35° 7'13.90"N | 84°34'4.71"W | SR-30 | Sandsuck Formation | Same conditions as Ocoee 1 Site |
| Ocoee 3 | Polk | 35° 2'35.90"N | 84°26'59.61"W | US-64 | Great Smoky Group: Anakeesta | East of road to Boyd's Gap overlook. Vertical cut |
| Sevierville 1 | Sevier | 35°49'1.80"N | 83°27'39.95"W | CR 416 | Snowbird Group: Roaring Fork | Bean Station near Dixie Hwy. Vertical Cut. |

Table 2.2. Road cut site information on type, age of cut, cut slope, facing aspect, and vegetation.

| Site Name | Formation | Type | Age | Cut Slope | Facing Aspect | Vegetation |
|---------------|------------------------------|---------------|------------|-----------|---------------|---|
| Grainger 1 | Chattanooga Shale | Single cut | < 20 yrs | 60° | East | None; forest above cut |
| Jamestown 1 | Fentress Formation | Earthen cover | < 20 yrs | 45° | West | Grass cover |
| Jamestown 2 | Fentress Formation | Earthen cover | < 20 yrs | 45° | West | Grass cover |
| Jamestown 3 | Fentress Formation | Terraced cut | > 60 yrs | 90° | North | Deciduous forest cover with herbal growth |
| Nashville 1 | Chattanooga Shale | Terraced cut | < 20 yrs | 90° | North | None |
| Nashville 2 | Chattanooga Shale | Terraced cut | < 20 yrs | 90° | North | None |
| Ocoee 1 | Sandsuck Formation | Single cut | > 60 yrs | 60° | South | Deciduous forest cover with herbal growth |
| Ocoee 2 | Sandsuck Formation | Single cut | > 60 yrs | 60° | South | Deciduous forest cover with herbal growth |
| Ocoee 3 | Great Smoky Group: Anakeesta | Single cut | 35-60 yrs | 90° | East | Sparse deciduous trees (Stree) |
| Sevierville 1 | Snowbird Group: Roaring Fork | Single cut | 35- 60 yrs | 90° | West | Deciduous forest cover |

Table 2.3. Weather station locations used to estimate storm event volumes at road cut monitoring stations. Weather stations maintained by the Tennessee Valley Authority.

| Weather Station Name | Location ID | Station Latitude | Station Longitude |
|---------------------------------------|-------------|------------------|-------------------|
| Wolf near Byrdstown | BYGT1 | 36.56028 | -85.07306 |
| Clear Fork near Robbins | CFKT1 | 36.38833 | -84.63028 |
| Columbia | CLBT1 | 35.66417 | -87.03250 |
| Copperhill | CPRT1 | 35.00389 | -84.38695 |
| Higdons Store | EPWG1 | 34.90333 | -84.43611 |
| Etowah | ETOT1 | 35.33028 | -84.52111 |
| Harpeth at Franklin, TN | FRAT1 | 35.92056 | -86.86555 |
| Gatlinburg | GTTT1 | 35.69167 | -83.53445 |
| Morristown | MSTT1 | 36.27250 | -83.22611 |
| Beaver Creek near Monticello | MTCK2 | 36.79750 | -84.89611 |
| Parksville | PKST1 | 35.09500 | -84.64917 |
| Appalachia Powerhouse near Turtletown | TURT1 | 35.18250 | -84.43833 |

Table 2.4. Road cut monitoring stations and the nearest TVA weather stations used to estimate storm event volumes. TVA Weather Station latitude and longitude locations defined in Table 2.3. Lat., long, and absolute (Δy , Δx) deviations summarized per road cut monitoring station.

| Road Cut Site Name | Road Cut Site Lat. | Road Cut Site Long. | TVA Weather Station | Lat. deviation (Δy) | Long. Deviation (Δx) | Absolute deviation |
|--------------------|--------------------|---------------------|---------------------|-------------------------------|--------------------------------|--------------------|
| Grainger 1 | 36.35042 | -83.33799 | MSTT1 | -0.078 | 0.112 | 0.190 |
| Jamestown 1 | 36.49708 | -84.96743 | BYGT1 | 0.063 | -0.106 | 0.169 |
| | | | MTCK2 | 0.300 | 0.071 | 0.372 |
| Jamestown 2 | 36.49526 | -84.96808 | BYGT1 | 0.065 | -0.105 | 0.170 |
| | | | MTCK2 | 0.302 | 0.072 | 0.374 |
| Jamestown 3 | 36.44119 | -84.96292 | BYGT1 | 0.119 | -0.110 | 0.229 |
| | | | CFKT1 | -0.053 | 0.333 | 0.385 |
| | | | MTCK2 | 0.356 | 0.067 | 0.423 |
| Nashville 1 | 35.81457 | -86.97514 | CLBT1 | -0.150 | -0.057 | 0.208 |
| | | | FRAT1 | 0.106 | 0.110 | 0.216 |
| Nashville 2 | 35.82145 | -86.95328 | FRAT1 | 0.099 | 0.088 | 0.187 |
| | | | CLBT1 | -0.157 | -0.079 | 0.236 |
| Ocoee 1 | 35.12050 | -84.56819 | PKST1 | -0.026 | -0.081 | 0.106 |
| | | | TURT1 | 0.062 | 0.130 | 0.192 |
| | | | ETOT1 | 0.210 | 0.047 | 0.257 |
| | | | CPRT1 | -0.117 | 0.181 | 0.298 |
| Ocoee 2 | 35.12053 | -84.56797 | PKST1 | -0.026 | -0.081 | 0.107 |
| | | | TURT1 | 0.062 | 0.130 | 0.192 |
| | | | ETOT1 | 0.210 | 0.047 | 0.257 |
| Ocoee 3 | 35.04331 | -84.44989 | CPRT1 | -0.117 | 0.181 | 0.298 |
| | | | CPRT1 | -0.039 | 0.063 | 0.102 |
| | | | TURT1 | 0.139 | 0.012 | 0.151 |
| Sevierville 1 | 35.81717 | -83.46110 | EPWG1 | -0.140 | 0.014 | 0.154 |
| | | | GTTT1 | -0.125 | -0.073 | 0.199 |



Figure 2.2. Photos of Pinson et al. (2003) flow divider buckets for passive runoff collection at the monitoring sites: a) Nashville 2 (N2) and b) Ocoee 1 (O1).

directed to the next bucket. A triangular leveling device constructed of angle iron and stainless steel adjustment bolts at each corner was used to ensure that the flow divider was level so outlet flow from the buckets was evenly divided. The first flow divider limits the maximum peak runoff rate that can be handled with this arrangement, which for a 12-weir divider is about 30 L/s. Hoomehr et al. (2013) successfully used this collection system for measuring runoff and sediment from coal reclamation sites in East Tennessee. In addition to Figure 2.2 above, also see Appendix A site photos.

The original intent of the Pinson et al (2003) runoff collection devices was to estimate volumes, however the total runoff volumes often exceeded the capacity of the station set-up, typically with three buckets (one 12 V-notch weir, and one 24 V-notch weir). Although runoff per storm event often exceeded the device capacity, they did provide a more thoroughly-mixed water sample for chemical analysis. For all stations and storm events sampled, runoff volumes were estimated using the standard Natural Resource Conservation Service method (NRCS 1986). The basic equations for this runoff estimates are as follow:

$$P_e = (P - 0.2*S)^2 / (P + 0.8*S); S = (25400/CN) - 254$$

where, P_e is the runoff depth (cm), P is precipitation depth (cm), S is the potential maximum retention, and CN is the curve number which are obtained from standard NRCS (1986) tables based on the degree of vegetation cover. The event total runoff volume (m^3) was computed by multiplying runoff depth (cm/100) by the site drainage area (m^2). Drainage areas for each station were estimated by a USGS digital elevation map (DEM) and field checked. The CN , computed S , and drainage area for each site are summarized in Table 2.5. Computed runoff volumes for each monitoring station and precipitation depths per storm event are summarized in Appendix C.

2.2.3 Water Sample Collections

Per collection site, runoff from 4 to 5 storm events was collected among different seasons, with the aim to collect 2 summer, 1 fall, 1 winter, and 2 spring runoff samples. Runoff water samples were collected with a minimum of 72 hours of dry weather prior to the storm event. Samples were collected within a maximum of two weeks after the storm event, which could not be avoided with the travel distances among all the sites.

Table 2.5. NRCS rainfall-runoff equation parameters curve number (CN) and potential maximum retention (S) or each monitoring station for metric precipitation depths.

| Road Cut Site Name | CN | S | Drainage Area (m ²) |
|--------------------|----|-------|---------------------------------|
| Grainger 1 | 97 | 7.86 | 47.38 |
| Jamestown 1 | 87 | 37.95 | 252.70 |
| Jamestown 2 | 87 | 37.95 | 363.25 |
| Jamestown 3 | 93 | 19.12 | 44.59 |
| Nashville 1 | 98 | 5.18 | 36.23 |
| Nashville 2 | 98 | 5.18 | 27.87 |
| Ocoee 1 | 93 | 19.12 | 44.59 |
| Ocoee 2 | 93 | 19.12 | 37.16 |
| Ocoee 3 | 93 | 19.12 | 52.96 |
| Sevierville 1 | 93 | 19.12 | 39.02 |

2.2.4 Water Quality Chemical Analysis

Water analyses were conducted at the University of Tennessee, Department of Civil and Environmental Engineering Water Quality laboratory by graduate students. Analyses followed American Water Works Association (AWWA) Standard Methods for use of an ion chromatograph (Dionex™ IC), inductively-coupled plasma optical emission spectrometer (ThermoFisher™, ICP-OES), wet chemistry titration procedures for pH and conductivity. U.S. Environmental Protection Agency (USEPA) and AWWA Standard Methods are the same, but test numbers differ; a cross-referencing list of tests is provided in Table 2.6. Water quality analyses include the following chemical parameters: pH, conductivity, sulfate, nitrate, chloride, calcium, magnesium, and dissolved metals (iron, aluminum, manganese, silica, copper, and zinc). Subsets of the four samples per site were analyzed for total metals. Hardness was computed from calcium, magnesium, and iron (IC/ICP) concentrations (mg/L) rather than using an ethylenediaminetetraacetic acid (EDTA) titrimetric method. Acid neutralizing capacity (ANC) was as an ion balance, where:

$$ANC = \sum \text{Base Cations} + \sum \text{Dissolved Metals} - \sum \text{Acid Anions}$$

Similar to the modified Molot et al. (1989) version used by Hyer et al. (1995), the change in the concentration sum of acid anions (sulfate, nitrate, chloride, and phosphate), and the sum of base cations (Na⁺, K⁺, Ca²⁺, and Mg²⁺) and dissolved metals (Al³⁺, Ba²⁺, Cd²⁺, Cu²⁺, Fe³⁺, Mn²⁺, Ni²⁺, Si⁴⁺, and Zn²⁺) was used to determine the relative contribution of each quantity to the total ANC change. Valances for the dissolved metals were assumed as shown. This approach is also consistent with methods used by Wellington and Driscoll (2004) to quantify the contribution of ion concentration change to overall ANC change for water chemistry. A comprehensive summary of the laboratory chemical analysis and parameter estimates are in Appendix D.

2.2.5 Site Water Quality Characterization

Water samples were collected from each of the 5-gallon buckets and analyzed separately. Because complete mixing is assumed as runoff flows from the gutter into the Pinson collection systems and sequentially from one bucket to the other, concentrations of each chemical parameter represents an average of each bucket sub-sample per site and event. Average concentrations were summarized by geologic formation. Mass loadings for each runoff event were computed from the average concentration multiplied by the event flow volume, and site

Table 2.6. Summary of water quality analyses, analytical chemical procedures, equipment, and method references.

| Analysis | Procedure | Equipment | Method References |
|--|---|---------------------------------|--|
| pH | Potentiometric | PC-Titration Plus™ | EPA Method 150.1 (laboratory) |
| Conductivity | Potentiometric | PC-Titration Plus™ | EPA Method 120.1 |
| Total Suspended Solids | Membrane Filtration | Vacuum Filtration | Standard Methods 2540D |
| Anions: (NO ₃ ⁻ , Cl ⁻ , PO ₄ ³⁻ , SO ₄ ²⁻) Monovalent Cations: (NH ₄ ⁺) | Ion Chromatography | Dionex™ Ion Chromatograph | Standard Methods 4110 |
| Total and Dissolved Metals (Na, K, Mg, Ca, Ba, Cd, Mn, Al, Fe, Cu, Ni, Zn, and Si) | Inductively-Coupled Plasma, Optical Emission Spectrometer | Thermo-Fisher™ Iris Intrepid II | Standard Methods 3120B EPA Method 6010B EPA Method 3005A |

export normalized per day. Units for site export of chemical mass loadings were grams per day (g/day), and only computed for ions and dissolved metals.

Results were compared to State Water Quality Standards, and published thresholds for biotic toxicity. The chemical parameters with standards or public thresholds include pH, ANC, and dissolved metals Al and Zn (Table 2.7). In general, a stream pH below 5.5 and dissolved metals (Al, Zn) above 0.2 mg/L can initiate aquatic biota stress and impairment. A detailed summary of toxicological effects from stream water acidification can be found in Appendix E.

TDEC water quality criteria (Rule 0400-40-03-.03) for the designated use Fish and Aquatic Life provide a means to compare the potential environmental impacts from highway road-cuts. The pH value shall not be outside the following ranges: 6.0 – 9.0 in wadeable streams, and values fluctuate more than 1.0 unit over a period of 24 hours. There shall be no turbidity, total suspended solids, or color in such amounts or of such character that will materially affect fish and aquatic life, which is a narrative criterion rather than numeric. The waters shall not contain iron at concentrations that cause toxicity or in such amounts that interfere with habitat due to precipitation or bacteria growth though no numeric limits are provided. Compound maximum concentrations (CMC) for dissolved metals include: cadmium (2.0 µg/L), copper (13.0 µg/L), lead (65 µg/L), nickel (470 µg/L), and zinc (120 µg/L). However, the CMC for these metals are hardness dependent, and toxicity is reduced with higher hardness levels. Dissolved aluminum does not have a TDEC regulatory limit, but literature indicates that Al concentrations greater than 0.2 mg/L may cause harm to aquatic biota (Table 2.7).

2.2.6 Statistical Analysis

Descriptive statistics were used to summarize site concentrations and mass loadings for water samples per storm event collected in terms of averages, standard deviations, and range (minimums and maximums). Statistics were applied to the site storm event chemistry and environmental factors: season, geology, vegetative cover, road-cut age, aspect, and slope (Table 2.2). Seasons were classified as: spring = March 15 through May 15, summer = May 15 through September 15, fall = September 15 through December 15, and winter = December 15 through March 15.

Analysis of Variance (ANOVA) means separation (Least Significance Difference, LSD) was used to test for significant differences in runoff chemistry among seasons. ANOVA LSD means separation was also used to test for significant differences in environmental site factors,

Table 2.7. Summary of toxic effects of pH, and aluminum and zinc concentrations on fish and benthic macroinvertebrates from published literature.

| Chemical | Fish | Benthic Macroinvertebrates |
|---|--|---|
| pH 5.5-6.4 5.0-5.5 4.0-5.0 < 4.0 | Reduced growth Reduced abundance; adverse effect to mortality; harmful to the eggs and fry Lethal to salmonids | Slightly impacted Moderately impacted <i>Baetis muticus</i> , <i>Heptagenia lateralis</i> and <i>R. semicolorata</i> absent; Native mayfly <i>B. alpinus</i> declined. Severely impacted Lower taxonomic richness; Scarce empididae, <i>Isoperla rivulorum</i> , <i>Rhithrogena</i> spp. and <i>Baetis</i> spp. (Ephemeroptera) Significantly fewer individuals and taxa; Reduced abundance resulted primarily from reduced abundance of mayflies (Ephemeroptera) |
| Aluminum $Al_{tot} > 0.2$ mg/L $Al_{tot} > 0.4$ mg/L $Al_{in} > 0.2$ mg/L | Loss of Na and Cl; Measureable reductions in survival and growth; Significant mortality of brook trout | Reduced density of Ephemeroptera and Ceratopogonidae Acute toxicity LC_{50} for <i>Hyalella azteca</i> (Crustacea), <i>Pisidium</i> spp. (Bivalvia), <i>Enallagma</i> sp. (Odonata) Mortality of <i>Gyraulus</i> sp. (Gastropoda), <i>Hyalella azteca</i> (Crustacea), <i>Chironomids</i> (Chironomidae) |
| Zinc > 0.047 mg/L > 0.11 mg/L > 0.219 mg/L | Fish avoidance Start to affect survival | Ceriodaphnia dubia abundance reduced by 50% |

including geology, vegetative cover, road-cut age, aspect and slope. Linear regression was applied to test for relationships between chemical parameters values/concentrations (dependent variable) and rainfall characteristics (independent variables). The rainfall characteristics included storm event volume (RFV, cm), average intensity (RFIA, cm/hr), and maximum intensity (RFIM, cm/hr). Statistical computations were conducted with SPSS v.23.

2.3 Results

2.3.1 Road Cut Site Chemistry

Site runoff chemistry was highly variable among the study sites as observed by the parameter site means, standard deviations, and ranges (Table 2.8). Chemistry differences among sites were due to many environmental factors including: APM geology, vegetative cover, physical conditions as to cut slope and aspect, road cut age, and others. In addition, variability in chemistry is possibly due to rainfall event intensity and volumes, and the seasonal differences in rainfall characteristics and air temperatures were investigated and results summarized in the next

Table 2.8. Summary of site chemistry by parameter. Site identifiers are: G1 is Grainger; J1, J2, & J3 are Jamestown; N1 & N2 are Nashville; O1, O2, & O3 are Ocoee; and S1 is Sevierville. Latitude and longitude coordinates are listed in Table 2.4. Site means, standard deviations (StDev), and range. Appendix D contains the details of all site and storm event chemistry.

| Chemical Parameter | SITE | | | | | | | | | |
|--|-----------------------------------|-----------------------------------|-----------------------------------|-----------------------------------|-----------------------------------|-----------------------------------|-----------------------------------|-----------------------------------|-----------------------------------|-----------------------------------|
| | G1 | J1 | J2 | J3 | N1 | N2 | O1 | O2 | O3 | S1 |
| TSS (mg/L) | 2114 (827) 100 – 2645 | 172 (146) 39 – 535 | 72 (18) 39 – 96 | 46 (8) 35 – 63 | 2057 (1012) 39 – 3136 | 350 (94) 135 – 442 | 351 (245) 38 – 802 | 131 (124) 37 – 460 | 51 (10) 39 – 72 | 83 (40) 38 – 185 |
| Conductivity (μ S/cm) | 173 (29) 117 – 209 | 112 (14) 93 – 133 | 298 (132) 22 – 428 | 31 (6) 20 – 38 | 5217 (388) 4686 – 6031 | 2533 (889) 1451 – 3460 | 174 (69) 92 – 348 | 225 (55) 140 – 309 | 479 (301) 237 – 1147 | 462 (307) 225 – 1104 |
| Hardness (mg/L) | 18.2 (2.6) 15.1 – 23.3 | 15.0 (2.9) 12.6 – 21.8 | 35.5 (8.2) 26.9 – 45.5 | 9.8 (0.6) 9.1 – 11.3 | 654.3 (79.2) 604.7 – 826.6 | 670.7 (63.5) 540.0 – 750.7 | 59.1 (5.8) 48.0 – 67.9 | 40.0 (4.1) 34.8 – 46.2 | 30.7 (6.8) 22.0 – 44.1 | 85.1 (15.2) 72.2 – 121.4 |
| pH [units] | 4.91 (0.34) 4.27 – 5.50 | 6.76 (0.26) 6.57 – 7.37 | 4.84 (1.48) 3.45 – 7.07 | 6.27 (0.19) 5.97 – 6.50 | 6.12 (0.40) 5.32 – 6.53 | 4.94 (0.74) 3.91 – 6.40 | 6.16 (1.13) 4.60 – 7.57 | 6.81 (0.98) 4.53 – 7.70 | 4.29 (1.01) 2.97 – 6.13 | 5.88 (0.77) 4.20 – 6.63 |
| ANC (meq/L) | -0.27 (0.32) -0.73 – 0.18 | -0.82 (0.57) -1.35 – 0.23 | -7.79 (5.03) -13.87 – -0.26 | 0.14 (0.06) 0.03 – 0.24 | -4.32 (6.09) -11.37 – 2.88 | 29.85 (3.74) 24.95 – 36.65 | 3.48 (0.70) 1.84 – 4.15 | 1.24 (0.30) 0.74 – 1.89 | 1.83 (0.97) -0.22 – 3.36 | 3.02 (0.45) 2.14 – 3.68 |
| Cl ⁻ (meq/L) | 0.17 (0.005) 0.16 – 0.17 | 0.17 (0.01) 0.16 – 0.19 | 0.46 (0.19) 0.21 – 0.66 | 0.13 (0.02) 0.09 – 0.18 | 0.41 (0.07) 0.33 – 0.53 | 0.42 (0.06) 0.33 – 0.52 | 0.14 (0.04) 0.08 – 0.21 | 0.17 (0.01) 0.16 – 0.19 | 0.19 (0.04) 0.16 – 0.26 | 0.20 (0.05) 0.16 – 0.30 |
| SO ₄ ²⁻ (meq/L) | 1.88 (0.11) 1.71 – 2.03 | 1.33 (0.44) 0.35 – 1.75 | 9.23 (5.50) 1.87 – 16.42 | 0.37 (0.07) 0.28 – 0.50 | 43.43 (5.53) 36.33 – 49.61 | 16.22 (2.90) 9.51 – 18.16 | 0.80 (0.31) 0.42 – 1.51 | 1.04 (0.10) 0.89 – 1.22 | 2.05 (0.52) 1.36 – 2.90 | 2.77 (0.84) 2.15 – 4.62 |
| NO ₃ ⁻ (meq/L) | 0.24 (0.01) 0.22 – 0.26 | 0.23 (0.01) 0.21 – 0.25 | 0.55 (0.24) 0.18 – 0.95 | 0.16 (0.04) 0.10 – 0.24 | 0.46 (0.10) 0.29 – 0.59 | 0.54 (0.18) 0.24 – 0.76 | 0.32 (0.11) 0.19 – 0.54 | 0.20 (0.01) 0.18 – 0.23 | 0.25 (0.06) 0.20 – 0.39 | 0.21 (0.04) 0.16 – 0.30 |
| PO ₄ ³⁻ (meq/L) | 0.40 (0.02) 0.38 – 0.44 | 0.39 (0.02) 0.34 – 0.40 | 1.09 (0.13) 0.91 – 1.23 | 0.23 (0.05) 0.17 – 0.37 | 0.53 (0.28) 0.28 – 0.93 | 0.92 (0.25) 0.60 – 1.34 | 0.27 (0.07) 0.18 – 0.39 | 0.32 (0.02) 0.28 – 0.35 | 0.38 (0.07) 0.31 – 0.53 | 0.35 (0.04) 0.30 – 0.44 |
| Sum AA (meq/L) | 2.68 (0.13) 2.52 – 2.85 | 2.12 (0.45) 1.15 – 2.55 | 11.32 (5.76) 3.44 – 18.49 | 0.89 (0.10) 0.77 – 1.13 | 44.84 (5.85) 37.24 – 50.76 | 18.10 (2.89) 11.41 – 20.14 | 1.54 (0.41) 0.98 – 2.55 | 1.73 (0.12) 1.53 – 1.93 | 2.88 (0.66) 2.15 – 3.96 | 3.52 (0.95) 2.77 – 5.58 |
| Na ⁺ (meq/L) | 0.21 (0.11) 0.04 – 0.30 | 0.15 (0.06) 0.00 – 0.22 | 0.23 (0.41) 0.00 – 1.02 | 0.07 (0.02) 0.02 – 0.09 | 1.01 (0.35) 0.04 – 1.26 | 2.59 (3.82) 0.00 – 11.93 | 0.47 (0.16) 0.08 – 0.65 | 0.24 (0.08) 0.08 – 0.40 | 1.28 (0.71) 0.00 – 2.66 | 0.60 (0.32) 0.00 – 1.05 |
| K ⁺ (meq/L) | 0.05 (0.01) 0.04 – 0.05 | 0.017 (0.01) 0.15 – 0.18 | 0.07 (0.02) 0.04 – 0.09 | 0.04 (0.00) 0.03 – 0.04 | 0.02 (0.00) 0.02 – 0.02 | 0.06 (0.04) 0.04 – 0.16 | 0.01 (0.00) 0.00 – 0.01 | 0.03 (0.00) 0.02 – 0.04 | 0.06 (0.02) 0.02 – 0.09 | 0.08 (0.01) 0.07 – 0.11 |
| Mg ²⁺ (meq/L) | 0.76 (0.11) 0.64 – 0.97 | 0.24 (0.05) 0.19 – 0.35 | 1.13 (0.14) 0.92 – 1.26 | 0.16 (0.01) 0.15 – 0.18 | 13.28 (1.25) 12.11 – 16.36 | 10.52 (2.91) 3.52 – 12.63 | 1.31 (0.13) 1.05 – 1.53 | 0.77 (0.13) 0.62 – 1.03 | 0.26 (0.10) 0.12 – 0.47 | 2.25 (0.39) 1.92 – 3.23 |
| Ca ²⁺ (meq/L) | 0.44 (0.07) 0.37 – 0.57 | 0.60 (0.12) 0.49 – 0.87 | 1.05 (0.33) 0.66 – 1.38 | 0.39 (0.03) 0.36 – 0.45 | 24.62 (3.27) 22.44 – 31.39 | 27.06 (2.07) 24.79 – 31.14 | 2.16 (0.23) 1.75 – 2.51 | 1.53 (0.15) 1.35 – 1.78 | 1.34 (0.29) 0.99 – 1.94 | 2.88 (0.52) 2.42 – 4.08 |
| Sum BC (meq/L) | 1.45 (0.26) 1.08 – 1.90 | 1.15 (0.21) 0.86 – 1.61 | 2.48 (0.76) 1.78 – 3.74 | 0.66 (0.04) 0.61 – 0.77 | 38.94 (4.16) 36.08 – 48.53 | 40.24 (1.78) 38.70 – 43.87 | 3.95 (0.47) 2.91 – 4.53 | 2.56 (0.29) 2.16 – 3.08 | 2.94 (1.00) 1.13 – 4.68 | 5.80 (1.08) 4.96 – 8.14 |
| Dis Al (mg/L) | 0.72 (0.30) 0.25 – 0.97 | 0.33 (0.15) 0.00 – 0.56 | 0.61 (0.38) 0.28 – 1.38 | 0.11 (0.03) 0.02 – 0.14 | 2.41 (1.15) 0.04 – 3.40 | 39.00 (16.17) 0.02 – 50.03 | 0.26 (0.11) 0.01 – 0.36 | 0.24 (0.18) 0.00 – 0.61 | 5.66 (1.92) 2.88 – 9.02 | 1.77 (0.96) 0.00 – 2.97 |
| Dis Cu (mg/L) | 0.06 (0.01) 0.04 – 0.07 | 0.06 (0.01) 0.04 – 0.07 | 0.16 (0.04) 0.07 – 0.20 | 0.02 (0.00) 0.01 – 0.02 | 0.50 (0.14) 0.22 – 0.63 | 0.94 (0.26) 0.26 – 1.09 | 0.04 (0.01) 0.03 – 0.06 | 0.07 (0.2) 0.04 – 0.12 | 0.66 (0.27) 0.25 – 1.14 | 0.31 (0.13) 0.15 – 0.58 |

| Chemical Parameter | SITE | | | | | | | | | |
|--------------------|------------------------------|----------------------------|----------------------------|----------------------------|--------------------------------|--------------------------------|-----------------------------|----------------------------|-----------------------------|----------------------------|
| | G1 | J1 | J2 | J3 | N1 | N2 | O1 | O2 | O3 | S1 |
| Dis Fe (mg/L) | 0.16 (0.08) 0.05 – 0.25 | 0.13 (0.07) 0.05 – 0.22 | 0.71 (1.08) 0.05 – 2.65 | 0.05 (0.01) 0.03 – 0.06 | 0.47 (0.19) 0.08 – 0.62 | 1.63 (1.20) 0.10 – 3.45 | 0.04 (0.01) 0.01 – 0.06 | 0.08 (0.04) 0.02 – 0.13 | 0.71 (0.29) 0.15 – 1.12 | 0.33 (0.18) 0.05 – 0.63 |
| Dis Mn (mg/L) | 0.66 (0.16) 0.40 – 0.80 | 0.26 (0.14) 0.00 – 0.34 | 2.10 (0.29) 1.51 – 2.35 | 0.07 (0.03) 0.00 – 0.09 | 2.32 (0.65) 0.96 – 2.65 | 5.01 (1.72) 1.22 – 6.09 | 0.02 (0.01) 0.00 – 0.02 | 0.21 (0.14) 0.00 – 0.49 | 2.92 (1.33) 0.69 – 4.79 | 1.56 (0.80) 0.35 – 2.96 |
| Dis Si (mg/L) | 5.83 (0.56) 5.48 – 7.14 | 0.65 (0.18) 0.54 – 1.07 | 6.05 (1.45) 3.82 – 7.68 | 2.50 (0.19) 2.30 – 2.88 | 7.87 (1.45) 6.87 – 11.70 | 19.66 (5.37) 6.49 – 22.72 | 7.34 (1.06) 4.74 – 8.55 | 2.54 (0.38) 2.18 – 3.50 | 6.36 (2.10) 2.99 – 10.72 | 2.94 (0.62) 2.35 – 4.19 |
| Dis Zn (mg/L) | 0.38 (0.13) 0.19 – 0.50 | 0.23 (0.10) 0.02 – 0.29 | 0.11 (0.02) 0.07 – 0.14 | 0.07 (0.02) 0.03 – 0.09 | 1.95 (0.73) 0.26 – 2.38 | 7.61 (3.06) 0.16 – 9.03 | 0.05 (0.01) 0.02 – 0.05 | 0.18 (0.11) 0.02 – 0.44 | 2.44 (1.21) 0.34 – 4.05 | 1.26 (0.65) 0.23 – 2.11 |
| Dis Ba (mg/L) | 0.12 (0.02) 0.09 – 0.14 | 0.06 (0.01) 0.05 – 0.07 | 0.23 (0.04) 0.18 – 0.29 | 0.05 (0.00) 0.04 – 0.06 | 0.50 (0.08) 0.34 – 0.57 | 0.46 (0.09) 0.32 – 0.57 | 0.08 (0.01) 0.07 – 0.09 | 0.08 (0.01) 0.07 – 0.11 | 0.47 (0.13) 0.38 – 0.80 | 0.35 (0.11) 0.21 – 0.57 |
| Dis Cd (mg/L) | 0.02 (0.01) 0.01 – 0.03 | 0.01 (0.01) 0.00 – 0.02 | 0.07 (0.01) 0.05 – 0.09 | 0.01 (0.00) 0.00 – 0.01 | 0.12 (0.05) 0.00 – 0.18 | 0.13 (0.05) 0.07 – 0.23 | 0.01 (0.00) 0.01 – 0.02 | 0.02 (0.01) 0.00 – 0.05 | 0.16 (0.10) 0.00 – 0.32 | 0.13 (0.07) 0.02 – 0.28 |
| Dis Ni (mg/L) | 0.09 (0.01) 0.07 – 0.11 | 0.02 (0.01) 0.01 – 0.03 | 0.12 (0.05) 0.06 – 0.21 | 0.00 (0.00) 0.00 – 0.00 | 0.30 (0.12) 0.17 – 0.59 | 1.87 (0.71) 0.15 – 2.39 | 0.00 (0.00) 0.00 – 0.00 | 0.01 (0.01) 0.00 – 0.02 | 0.12 (0.08) 0.01 – 0.29 | 0.05 (0.2) 0.01 – 0.08 |
| Total Al (mg/L) | 8.91 (1.22) 7.70 – 10.48 | 1.68 (0.68) 0.95 – 2.31 | 1.41 (0.96) 0.32 – 2.39 | 0.64 (0.33) 0.31 – 1.08 | 57.54 (21.56) 35.20 – 86.9 | 44.44 (22.92) 23.41 – 70.04 | 1.15 (0.98) 0.32 – 2.30 | 1.08 (0.54) 0.75 – 2.03 | 8.58 (9.12) 1.65 – 22.76 | 1.13 (0.63) 0.53 – 2.06 |
| Total Cu (mg/L) | 0.04 (0.01) 0.03 – 0.06 | 0.06 (0.01) 0.01 – 0.07 | 0.17 (0.04) 0.07 – 0.21 | 0.01 (0.01) 0.00 – 0.03 | 0.19 (0.21) 0.00 – 0.48 | 0.72 (0.87) 0.05 – 1.96 | 0.04 (0.01) 0.03 – 0.06 | 0.00 (0.01) 0.00 – 0.02 | 0.00 (0.01) 0.00 – 0.02 | 0.01 (0.01) 0.00 – 0.02 |
| Total Fe (mg/L) | 11.49 (1.88) 9.91 – 14.06 | 1.63 (0.91) 0.62 – 2.68 | 3.18 (2.96) 0.35 – 5.76 | 1.04 (0.80) 0.30 – 2.00 | 253.9 (135.2) 130.4 – 445.7 | 320.4 (244.1) 97.6 – 646.9 | 1.81 (1.66) 0.10 – 3.58 | 1.48 (2.15) 0.29 – 5.32 | 7.33 (7.31) 2.62 – 20.05 | 4.19 (2.22) 1.44 – 5.91 |
| Total Mn (mg/L) | 0.71 (0.13) 0.58 – 0.87 | 0.13 (0.04) 0.07 – 0.17 | 1.20 (1.25) 0.02 – 2.29 | 0.22 (0.11) 0.08 – 0.36 | 2.15 (0.44) 1.53 – 2.56 | 2.50 (1.04) 1.28 – 3.51 | 0.15 (0.13) 0.02 – 0.29 | 0.21 (0.22) 0.04 – 0.59 | 1.52 (1.35) 0.49 – 3.50 | 0.33 (0.23) 0.07 – 0.64 |
| Total Si (mg/L) | 9.36 (0.68) 8.54 – 10.17 | 3.46 (1.47) 1.82 – 4.75 | 4.17 (3.16) 0.56 – 6.93 | 2.18 (0.54) 1.54 – 2.77 | 29.78 (6.42) 23.62 – 38.77 | 19.85 (7.91) 12.22 – 28.16 | 8.71 (3.23) 5.35 – 12.66 | 4.63 (0.89) 3.27 – 5.71 | 9.44 (8.20) 2.00 – 21.29 | 3.66 (1.14) 1.98 – 5.17 |
| Total Zn (mg/L) | 0.19 (0.04) 0.14 – 0.22 | 0.21 (0.14) 0.03 – 0.34 | 0.07 (0.03) 0.01 – 0.11 | 0.06 (0.05) 0.02 – 0.12 | 2.45 (0.78) 1.65 – 3.44 | 4.13 (2.08) 1.43 – 6.08 | 0.20 (0.19) 0.03 – 0.45 | 0.07 (0.06) 0.03 – 0.17 | 0.65 (0.50) 0.24 – 1.43 | 0.22 (0.11) 0.10 – 0.32 |
| Total Cd (mg/L) | 0.01 (0.005) 0.00 – 0.01 | 0.13 (0.21) 0.00 – 0.44 | 0.02 (0.03) 0.00 – 0.06 | 0.01 (0.00) 0.00 – 0.01 | 0.29 (0.24) 0.04 – 0.56 | 0.07 (0.05) 0.00 – 0.11 | 0.09 (0.13) 0.00 – 0.28 | 0.02 (0.01) 0.00 – 0.05 | 0.01 (0.01) 0.00 – 0.03 | 0.01 (0.01) 0.00 – 0.03 |
| Total Ni (mg/L) | 0.08 (0.03) 0.06 – 0.12 | 0.15 (0.14) 0.01 – 0.30 | 0.04 (0.03) 0.00 – 0.08 | 0.00 (0.00) 0.00 – 0.01 | 1.36 (0.38) 0.91 – 1.80 | 1.23 (0.68) 0.50 – 2.12 | 0.05 (0.10) 0.00 – 0.24 | 0.02 (0.02) 0.00 – 0.05 | 0.05 (0.02) 0.03 – 0.08 | 0.03 (0.02) 0.00 – 0.07 |

Sample numbers (N) per site for all parameters except total metal concentrations were as follow: G1 = 8, J1 = 9, J2 = 7, J3 = 12, N1 = 11, N2 = 8, O1 = 15, O2 = 13, O3 = 14, and S1 = 12. Four of each of these samples were randomly selected and analyzed for total metal concentrations.

TSS = total suspended solids; Sum AA = sum of acid anions; Sum BC = sum of base cations; Dis = dissolved.

sub-section (2.3.2). In this sub-section, water chemistry of the road-cut runoff is reviewed collectively for the ten study sites, and compared to water quality criteria.

Suspended Solids: The mean concentrations for total suspended solids (TSS) ranged from 48 to 2,114 mg/L among the sites, which was highly variable (Table 2.8). Two sites Nashville 1 and Grainger 1 had the highest mean concentrations above 2,000 mg/L. These two sites were non-vegetated and the source of these suspended solids was unknown. It is possible the solids are iron precipitates, fine sediment such as sand/silt/clay, or organic material that collected in the buckets. Six sites were below 200 mg/L, which would be more typical for runoff waters. The mean concentrations for Nashville 2 and Ocoee 1 sites were approximately 350 mg/L.

Conductivity: Conductivity is an indicator to ion concentrations in waters. The Nashville sites (N1 and N2) had very high conductivity values with means above 2,500 $\mu\text{S}/\text{cm}$, which conductivity correlates with the reported high hardness values (Table 2.8). Consistent with pyritic geologic formations with underburden limestone, the main contributing ions are sulfate, calcium, magnesium, iron, and aluminum. Jamestown 3 has low conductivity in the range from 20 to 38 $\mu\text{S}/\text{cm}$, which road cut was old and highly-vegetated with tree cover. Mean conductivities for the seven other study sites ranged from 112 to 479 $\mu\text{S}/\text{cm}$ and these values would be expected for surface waters.

Hardness: Site characteristics for hardness concentrations correlated with the pattern observed with conductivity, where the Nashville sites (N1 and N2) were much higher than all other sites with means of 645.3 and 670.7 mg/L, respectively (Table 2.8). The other sites generally had mean hardness values that were low below 35 mg/L, which would be considered “soft” water. These sites included the Jamestown sites (J1, J2, and J3), Ocoee 3, and Grainger 1. Ocoee sites (O2 and O3) had mean hardness concentrations below 60 mg/L.

Acidity Indicators (pH and ANC): Overall site means for pH ranged from 4.29 to 6.76, and ANC ranged from -7.79 to 29.85 meq/L (Table 2.8). Five sites had mean pH values below 6.0, which were: Grainger 1, Jamestown 2, Nashville 2, Ocoee 3, and Sevierville 1. Four sites had ANC concentrations below zero which indicates a greater concentration of negatively-charged ions compared to positively-charged ions in solution. These sites were: Grainger 1, Jamestown 1 and 2, and Nashville 1. Though the state regulations do not have a water quality standard for ANC, above 50 $\mu\text{eq}/\text{L}$ is generally considered healthy for aquatic biota. The Jamestown (1 and 2) sites were relatively in close proximity to each other, likewise so were the Nashville sites (1 and 2) but acidity indicators were very different. This observation illustrates the spatially variable of water quality from road cuts. The USGS measured pH at seeps and generally found lower pH values than from our runoff water samples (M. Bradley, *personal communication*).

Acid Anions: Sum of acid anions (Sum AA) include negatively-charged ions: chloride, nitrate, chloride, and phosphate, in which the sum of these ions and mean concentrations for the study sites ranged from 0.89 to 44.84 meq/L (Table 2.8). The Nashville 1 site had the mean highest concentration, and the Jamestown 3 had the lowest. Nashville 2 and Jamestown 2 were above 10 meq/L. The other sites were all below 3.5 meq/L. In all cases, sulfate (SO_4^{2-}) were significantly greater than the other anions. Sulfate is the by-product of oxidation of pyrite (FeS), therefore this result was expected with a study designed to collect runoff from exposed pyrite geology to air and water. Sulfate concentrations ranged from 0.4 to 43.4 meq/L (17.6 to 2,084.9 mg/L). Most site mean concentrations were below 100 mg/L, which are typical for runoff from pyrite geology. The Nashville 1 and 2 sites with concentrations above 1,000 mg/L are considered very high. Sulfate is the primary driver for runoff acidification unless there are base cation sources that neutralize this acid anion.

Base Cations: Sum of base cations (Sum BC) include positivity-charges ions: sodium, potassium, calcium, and magnesium, in which the sum of the ions and mean concentrations for the study sites ranged from 0.66 to 40.24 meq/L (Table 2.8). Site mean concentrations for Sum BC were greatest at both the Nashville (N1 and N2) sites at 38.94 and 40.24 meq/L, respectively. The Nashville sites included the Chattanooga shale with a limestone overburden exposed to precipitation, which likely is the source of the high levels of base cations. All other sites had mean concentrations of Sum BC were below 6 meq/L, which lower concentrations potentially limit the neutralization capacity at each site. Mean concentrations of calcium and magnesium were consistently greater than sodium and potassium, which is expected from pyritic shales and limestone rock exposed to precipitation. In most cases mean concentrations of calcium was greater than magnesium but not always, for example Grainger 1 and Jamestown 2 magnesium was slightly greater. The overall mean calcium and magnesium concentrations for the Nashville sites were 512.0 mg/L and 147.2 mg/L, respectively. The overall mean calcium and magnesium concentrations for all other sites excluding the Nashville sites were 28.1 mg/L and 10.6 mg/L, respectively.

Aluminum: Site mean concentrations for total Al ranged from 0.64 to 57.54 mg/L, and dissolved Al ranged from 0.11 to 39.00 mg/L (Table 2.8). Nashville 1 site was observed with the very high dissolved concentration. Mean concentrations for dissolved Al for six sites were below 1.0 mg/L, and among those sites pH values were above 6. Dissolved Al is generally more available from soil and rock source with pH below 5.0. Total Al concentrations were very high at both Nashville sites, which was unique among all the study sites. Total Al concentrations were about 9 mg/L for Grainger 1 and Ocoee 3 with all remaining sites with concentrations were below 2 mg/L. With total concentrations greater than dissolved, it appears much of the potentially available Al was bound to soil particles but as noted above is dependent on pH. Aquatic biota is exposed to dissolved concentration and is the concentration that will be reviewed for toxicity in Sub-section 2.3.3.

Iron: Site mean concentrations for total Fe ranged from 1.04 to 320.4 mg/L, and dissolved Fe ranged from 0.05 to 1.63 mg/L (Table 2.8). The Nashville (N1 and N2) sites were observed with the very high total and dissolved Fe concentrations: total concentrations were 253.9 and 320.4 mg/L and dissolved concentrations were 0.47 and 1.63 mg/L, respectively. It is evident from these differences that most the free iron precipitated from ferrous (Fe^{2+}) to ferric (Fe^{3+}) oxidation number, with the precipitate (road-cut red stain, rust) as Fe_2O_3 . Figure 2.2a illustrates the red staining on the road cut surface at the Nashville (N1) site. Total Fe concentrations for all other sites were below 12 mg/L. Dissolved Fe concentrations for all other sites were below 0.7 mg/L. Though the total concentrations at the other sites were lower than the Nashville sites, the large lower difference with dissolved Fe was similar indicating the iron was oxidized and precipitates.

Silicon: Site mean concentrations for total Si ranged from 3.46 to 29.78 mg/L, and dissolved Si ranged from 0.65 to 19.66 mg/L (Table 2.8). The Nashville (N1 and N2) sites were observed with the highest total and dissolved Si concentrations: total concentrations were 29.78 and 19.85 mg/L and dissolved concentrations were 7.87 and 19.66 mg/L, respectively. Dissolved Si concentrations for all other sites were mostly below 7 mg/L. Mineralization of silicon was evident from the differences from total and dissolved concentrations, where dissolved concentrations are typical for surface waters with shale and limestone formations.

Zinc: Site mean concentrations for total Zn ranged from 0.07 to 4.13 mg/L, and dissolved Zn ranged from 0.05 to 7.61 mg/L (Table 2.8). The Nashville (N1 and N2) sites were observed with the highest total and dissolved Zn concentrations: total concentrations were 2.45 and 4.13 mg/L

and dissolved concentrations were 1.95 and 7.61 mg/L, respectively. The higher mean dissolved Zn for Site N2 compared with total was unexplained and additional site testing is recommended.

Total and Dissolved Metals: Site mean concentrations for total and dissolved metals other than the metals reported above included Cu, Mn, Ba, Cd, and Ni. The mean concentration ranges for dissolved metals, for all sites included: 0.02 – 0.94 mg/L, 0.02 – 5.01 mg/L, 0.05 – 0.50 mg/L, 0.01 – 0.13 mg/L, and < 0.01 – 0.03 mg/L, respectively. Consistent with the other metals, the dissolved metal concentrations were highest at the Nashville sites. Those dissolved metals that may cause impairment are summarized below (Table 2.7, Appendix E).

2.3.2 Environmental Factors Associated with Road Cut Site Chemistry

Runoff chemistry was examined for relationships to various environmental factors, which included rainfall characteristics (storm event volume and intensities), season, and site characteristics including geology, vegetative cover, road cut slope and aspect, and age. Storm event characteristics are summarized in Table 2.9. Statistical analyses were conducted on these data and summarized below.

Rainfall Characteristics: Runoff samples were collected from a range of rainfall event volumes and intensities per site, in which it was assessed to whether event characteristics affected runoff chemistry values. Linear regression analyses were completed with the chemical parameters as the dependent variable and three rainfall event characteristics (volumes, average intensity, and maximum intensity) as independent variable. No significant relationships were observed between chemical parameter values/concentrations and storm event volumes (RFV); all statistical p-levels were greater than 0.9. Chemical ions sulfate (SO_4^{2-}), Sum AA, Sum BC were significantly related to storm event average intensities (RFIA) though weakly correlated as shown in Figure 2.3 ($p = 0.04$, $R^2 = 0.19$; $p = 0.04$, $R^2 = 0.20$, and $p = 0.05$, $R^2 = 0.19$, respectively). Results from these significance relationships were from the larger data set providing greater statistical power ($N_{\text{total}} = 109$), however the poor correlations indicate that many factors other than rainfall intensity influences ion export from the road cut sites. With regards to Sum BC and observed in Figure 2.3, the higher concentrations were from the Nashville sites and these data points strongly influence the statistical results. Dissolved metals were not significantly related to RFIA (p values between 0.18 and 0.32), which indicates metals are less freely available at the rock surfaces exposed to storm events. Chemical parameter values/ concentrations were not significantly related to storm event maximum intensity (RFIM), in which significance levels were greater than 0.85 for the acid anions and base cations, and greater than 0.40 for the dissolved metals.

Seasons: Runoff chemistry was generally not influenced by seasons. Except for a very few chemical parameters, no significant relationships were observed. TSS was found significantly different between winter and summer with means of 75.4 mg/L and 609.6 mg/L, respectively ($p = 0.01$). Differences are likely due to more available organic matter during warm months. Two dissolved metals were found to significantly differ between seasons. Copper was different between winter and spring ($p = 0.03$), and cadmium was different between winter and all other seasons ($p = 0.003$ - 0.018). With only dissolved Cu and Cd found to differ from winter to other seasons and the other dissolved metals did not differ, these relationships are unexplained.

Geology: Runoff chemistry was influenced by road-cut exposed formations, which included the five different types: Chattanooga Shale (Chat), Fentress Formation (Fentr), Sandsuck Formation (Sand), Great Smoky Group: Anakeesta Formation (Anakee), and Snowbird Group: Roaring Fork Sandstone (Snowb). A summary of the runoff chemistry is in Table 2.10. The

Table 2.9. Summary of storm event characteristics for estimated volumes, and average and maximum intensities, and runoff volumes. Site means and (standard deviations) and range. Appendix C contains the details of all site and storm event data.

| Site | No. of Events | Duration (days) | Rainfall Volume (cm) | Rainfall Average Intensity (cm/hr) | Rainfall Maximum Intensity (cm/hr) | Runoff Volumes (m ³) |
|------------------|---------------|----------------------|-----------------------------|------------------------------------|------------------------------------|----------------------------------|
| Grainger (G1) | 8 | 3.1 (1.46) 1 – 6 | 3.74 (2.08) 0.61 – 7.52 | 0.19 (0.09) 0.08 – 0.33 | 0.76 (0.29) 0.23 – 1.12 | 1.41 (0.93) 0.08 – 3.15 |
| Jamestown (J1) | 9 | 1.44 (0.73) 1 – 3 | 2.38 (1.17) 0.76 – 5.18 | 0.14 (0.10) 0.05 – 0.33 | 0.69 (0.63) 0.13 – 2.11 | 1.47 (1.75) 0.02 – 6.01 |
| Jamestown (J2) | 7 | 1.43 (0.97) 1 – 3 | 2.02 (0.59) 0.76 – 2.54 | 0.12 (0.07) 0.05 – 0.23 | 0.05 (0.34) 0.13 – 1.17 | 1.29 (0.67) 0.02 – 2.07 |
| Jamestown (J3) | 12 | 1.50 (0.80) 1 – 3 | 2.20 (1.33) 0.71 – 5.84 | 0.13 (0.10) 0.05 – 0.36 | 0.66 (0.63) 0.15 – 2.29 | 0.44 (0.47) 0.02 – 1.80 |
| Nashville (N1) | 11 | 1.65 (0.50) 1 – 2 | 3.05 (2.73) 0.61 – 7.67 | 0.21 (0.14) 0.05 – 0.48 | 0.78 (0.59) 0.15 – 2.11 | 0.92 (0.82) 0.09 – 2.57 |
| Nashville (N2) | 8 | 1.75 (0.46) 1 – 2 | 2.81 (2.01) 0.61 – 5.72 | 0.18 (0.11) 0.05 – 0.38 | 0.64 (0.40) 0.15 – 1.30 | 0.64 (0.54) 0.07 – 1.43 |
| Ocoee (O1) | 15 | 2.60 (1.55) 1 – 6 | 3.08 (3.23) 0.48 – 13.41 | 0.11 (0.06) 0.03 – 0.23 | 0.98 (0.69) 0.10 – 2.26 | 0.81 (1.29) 0.02 – 5.07 |
| Ocoee (O2) | 13 | 2.62 (1.66) 1 – 6 | 3.28 (3.43) 0.48 – 13.41 | 0.11 (0.06) 0.03 – 0.23 | 1.00 (0.74) 0.10 – 2.26 | 0.74 (1.14) 0.02 – 4.22 |
| Ocoee (O3) | 14 | 2.36 (1.60) 1 – 6 | 3.41 (4.22) 0.33 – 16.08 | 0.12 (0.09) 0.03 – 0.33 | 0.55 (0.45) 0.13 – 1.52 | 1.18 (2.00) 0.02 – 7.41 |
| Sevierville (S1) | 12 | 2.67 (1.50) 1 – 6 | 2.85 (1.61) 1.30 – 7.32 | 0.20 (0.23) 0.08 – 0.91 | 0.82 (0.73) 0.25 – 2.92 | 0.57 (0.54) 0.12 – 2.12 |

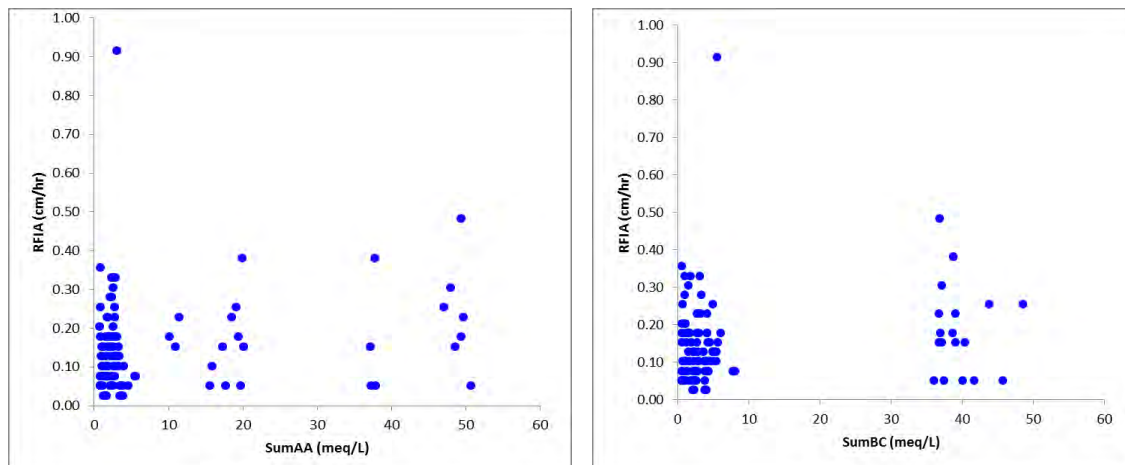


Figure 2.3. Scatterplots of sum of acid anions (Sum AA) and base cations (Sum BC) with storm event average intensity (RFIA). RFIA vs Sum AA ($p = 0.04$, $R^2 = 0.20$), RFIA vs Sum BC ($p = 0.04$, $R^2 = 0.19$).

chemistry among the five formations varied greatly. Mean pH values were: 5.41, 6.07, 6.47, 4.29, and 5.88, respectively. Mean ANC concentrations: 7.00, -2.14, 2.44, 1.83, and 3.02, respectively. The lack of correlation between pH and ANC indicates that the different road cut formations export a unique composition of acid anions, base cations, and dissolved metals (Figure 2.4). For example, the Chattanooga Shale had the greatest mean ANC with a mean pH of 5.41 and the Fentress Formation with the lowest mean ANC and a mean pH of 6.07. The Anakeesta Formation had the lowest mean pH of 4.29 with a mean ANC of 1.83 meq/L.

The runoff chemistry of the Chattanooga Shale was much different from the other rock types studied, apparently dominated by the chemistry at the Nashville (N1 and N2) sites. The Grainger 1 site was another site through the Chattanooga Shale in which road-cut runoff water was collected and analyzed. Runoff conductivity and hardness were much higher compared to the other rock types indicating high concentrations of ions and dissolved metals (Figure 2.5). The mean ANC was 7.00 meq/L, though relatively acidic compared to the other sites it indicated the presence of base cations in greater concentrations than the other sites (Table 2.10). Calcium and magnesium concentrations were significantly greater from the Chattanooga Shale road-cut runoff compared with the other rock types, with means of 18.18 meq/L and 8.75 mg/L. All other rock types were below 3.0 meq/L for Ca²⁺ and Mg²⁺ concentrations. The mean pH of 5.41 indicated acidity from acid anions, which was primarily from sulfate with a concentration of 23.06 meq/L. All other acid anions for this geologic formation were below 1 meq/L. In addition to the higher concentrations of acid anions and base cations, dissolved metals were of greater concentrations in the road-cut runoff from the Chattanooga Shale compared to the other rock types (Figure 2.6).

Table 2.10. Summary of runoff chemistry for the different formations along road-cut exposed to rainfall. Formations are Great Smoky Group: Anakeesta Formation (Anakee), Chattanooga Shale (Chat), Fentress Formation (Fentr), Sandsuck Formation (Sand), and Snowbird Group: Roaring Fork Sandstone (Snowb). Means and (standard deviations) provided per parameter and geologic formation. Letters ^{A, B, C, D} indicate significant differences per chemical ($p \leq 0.05$).

| Chemical Parameter | Geologic Formation | | | | |
|---------------------------------------|---------------------------------|---------------------------------|---------------------------------|---------------------------------|---------------------------------|
| | Anakee | Chat | Fentr | Sand | Snowb |
| TSS (mg/L) | 51.1 (10.5) ^B | 1,567.9 (1,108.7) ^A | 93.2 (97.7) ^B | 249.0 (224.7) ^B | 82.7 (40.0) ^B |
| Conductivity (µS/cm) | 479.5 (300.8) ^B | 2,927.2 (2,207.1) ^A | 124.2 (125.6) ^B | 197.8 (66.9) ^B | 462.1 (306.9) ^B |
| Hardness (mg/L) | 30.6 (6.8) ^B | 470.7 (305.1) ^A | 17.9 (11.4) ^B | 50.2 (10.6) ^B | 85.1 (15.2) ^B |
| pH [units] | 4.29 (1.00) ^C | 5.41 (0.77) ^A | 6.07 (1.04) ^B | 6.47 (1.10) ^B | 5.88 (0.77) ^A |
| ANC (meq/L) | 1.83 (0.97) ^C | 7.00 (15.79) ^A | -2.14 (4.11) ^B | 2.44 (1.26) ^C | 3.02 (0.46) ^C |
| Cl ⁻ (meq/L) | 0.19 (0.04) ^B | 0.34 (0.28) ^A | 0.22 (0.16) ^B | 0.16 (0.03) ^B | 0.20 (0.05) ^B |
| SO ₄ ²⁻ (meq/L) | 2.05 (0.51) ^B | 23.06 (18.50) ^A | 2.89 (4.56) ^B | 0.91 (0.26) ^B | 2.77 (0.84) ^B |
| NO ₃ ⁻ (meq/L) | 0.25 (0.06) ^B | 0.42 (0.17) ^A | 0.28 (0.20) ^B | 0.27 (0.10) ^B | 0.21 (0.04) ^B |
| PO ₄ ³⁻ (meq/L) | 0.38 (0.07) ^C | 0.61 (0.31) ^A | 0.49 (0.36) ^B | 0.29 (0.06) ^C | 0.35 (0.04) ^C |
| Sum AA (meq/L) | 2.88 (0.66) ^B | 24.42 (18.69) ^A | 3.89 (5.18) ^B | 1.63 (0.32) ^B | 3.52 (0.95) ^B |
| Na ⁺ (meq/L) | 1.28 (0.71) ^A | 1.24 (2.21) ^A | 0.13 (0.21) ^B | 0.36 (0.18) ^{BC} | 0.60 (0.32) ^C |
| K ⁺ (meq/L) | 0.06 (0.02) ^D | 0.04 (0.03) ^A | 0.09 (0.06) ^B | 0.02 (0.01) ^C | 0.08 (0.01) ^B |
| Mg ²⁺ (meq/L) | 0.26 (0.10) ^B | 8.75 (5.67) ^A | 0.43 (0.42) ^B | 1.06 (0.30) ^B | 2.25 (0.39) ^B |
| Ca ²⁺ (meq/L) | 1.34 (0.29) ^B | 18.18 (11.99) ^A | 0.62 (0.32) ^B | 1.87 (0.37) ^B | 2.88 (0.52) ^B |
| Sum BC (meq/L) | 2.94 (1.00) ^B | 28.21 (17.92) ^A | 1.27 (0.84) ^B | 3.30 (0.80) ^B | 5.80 (1.08) ^B |
| Dis Al (mg/L) | 5.66 (1.92) ^B | 12.75 (19.31) ^A | 0.31 (0.28) ^B | 0.25 (0.14) ^B | 1.77 (0.96) ^B |
| Dis Cu (mg/L) | 0.66 (0.27) ^C | 0.50 (0.38) ^A | 0.07 (0.06) ^B | 0.06 (0.02) ^B | 0.31 (0.13) ^D |
| Dis Fe (mg/L) | 0.71 (0.29) ^A | 0.72 (0.88) ^A | 0.24 (0.58) ^B | 0.06 (0.03) ^C | 0.33 (0.18) ^B |
| Dis Mn (mg/L) | 2.92 (1.33) ^A | 2.62 (1.99) ^A | 0.64 (0.88) ^B | 0.11 (0.14) ^C | 1.56 (0.80) ^{AB} |
| Dis Si (mg/L) | 6.36 (2.10) ^C | 10.76 (6.64) ^A | 2.79 (2.19) ^B | 5.11 (2.56) ^C | 2.94 (0.62) ^B |
| Dis Zn (mg/L) | 2.43 (1.21) ^A | 3.16 (3.44) ^A | 0.13 (0.09) ^B | 0.11 (0.10) ^B | 1.25 (0.65) ^C |
| Dis Ba (mg/L) | 0.47 (0.13) ^C | 0.37 (0.18) ^A | 0.10 (0.08) ^B | 0.08 (0.01) ^B | 0.34 (0.11) ^A |
| Dis Cd (mg/L) | 0.15 (0.10) ^A | 0.09 (0.06) ^A | 0.02 (0.03) ^B | 0.01 (0.01) ^B | 0.12 (0.07) ^A |
| Dis Ni (mg/L) | 0.12 (0.08) ^D | 0.71 (0.86) ^A | 0.04 (0.06) ^B | 0.01 (0.01) ^C | 0.05 (0.02) ^B |

Sum AA = sum of acid anions; Sum BC = sum of base cations; Dis = dissolved

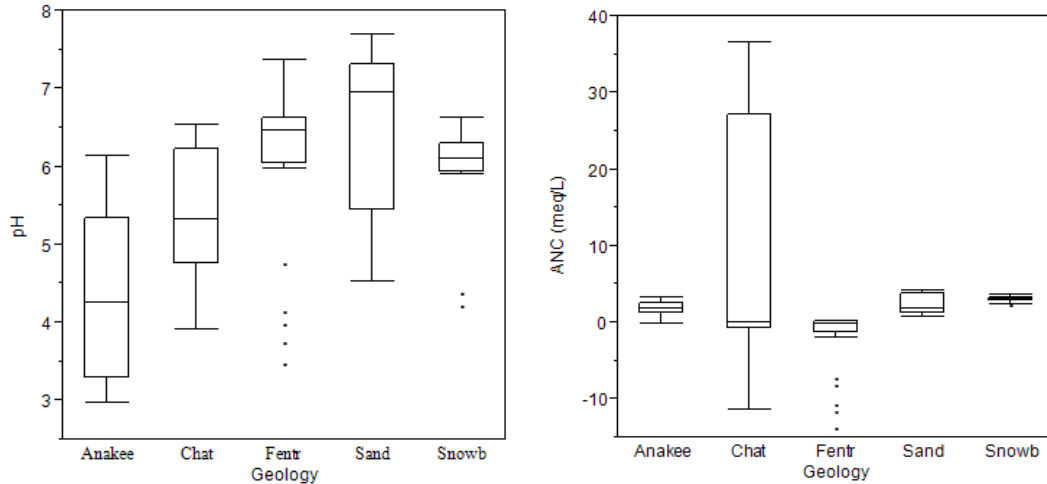


Figure 2.4. Boxplots of runoff chemistry for pH and ANC (meq/L) for the five different road-cut formations. Formations are Great Smoky Group: Anakeesta Formation (Anakee), Chattanooga Shale (Chat), Fentress: Formation (Fentr), Sandsuck Formation (Sand), and Snowbird Group: Roaring Fork Sandstone (Snowb).

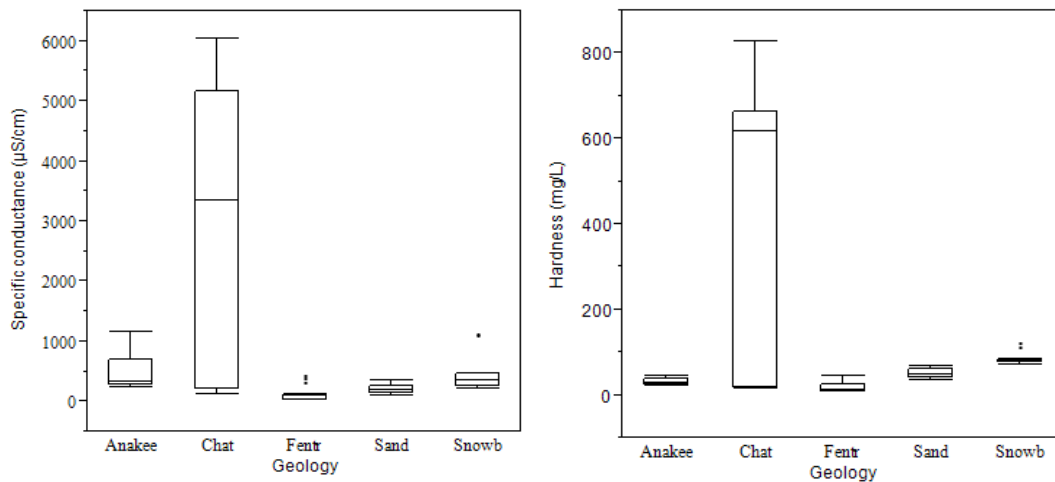


Figure 2.5. Boxplots of runoff chemistry for conductivity ($\mu\text{S}/\text{cm}$) and total hardness (mg/L) for the five different road-cut formations. Formations are: Great Smoky Group: Anakeesta Formation (Anakee) Chattanooga Shale (Chat), Fentress Formation (Fentr), Sandsuck Formation (Sand), and Snowbird Group: Roaring Fork Sandstone. (Snowb).

The runoff chemistry of the Anakeesta Formation was found with the lowest mean pH among the geologic formations studied (Table 2.10, Figure 2.4). ANC was positive with a concentration of 1.83 meq/L governed by the Sum BC concentrations slightly greater than the Sum AA. Other than the Chattanooga Shale, runoff from this rock formation was high in dissolved metals (Figure 2.6). Of the dissolved metals, Al, Fe, Cu, Mn, Si, and Zn were found elevated. It appears that the low pH from the Anakeesta Formation was governed by mineral weathering, hydrolysis and mobilization. Though the mean pH for the Roaring Fork Sandstone (Snowb) had a higher value of 5.88 than the Anakeesta Formation, the governing cause for runoff acidity appears to be the same with metals dissolution and hydrolysis.

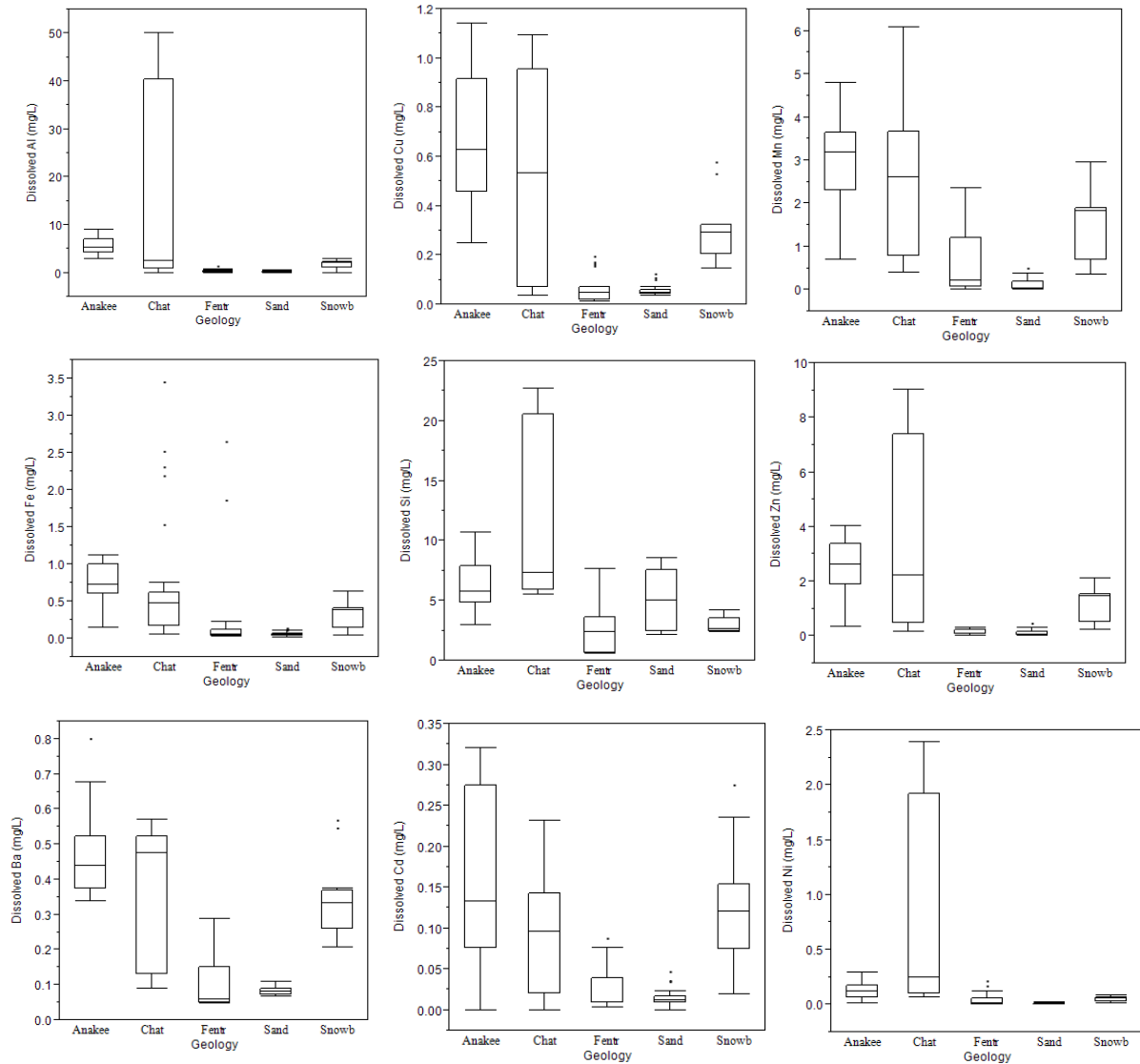


Figure 2.6. Boxplots of runoff chemistry for dissolved metals (mg/L) including Al, Cu, Mn, Fe, Si, Zn, Ba, Cd, and Ni for the five different road-cut formations. Formations are: Great Smoky Group: Anakeesta Formation (Anakee) Chattanooga Shale (Chat), Fentress Formation (Fentr), Sandsuck Formation (Sand), and Snowbird Group: Roaring Fork Sandstone (Snowb).

The runoff chemistry of the Fentress Formation had a pH of 6.01 but its runoff had the lowest mean ANC concentration of -2.14 meq/L compared with the other rock types studied (Table 2.10). The Sum AA concentration was much greater than the Sum BC explaining one cause of the low ANC. Overall, the ion and dissolved metal concentrations were low compared to the other rock types. Ion concentrations, ANC, and pH values were highly variable as observed by the event standard deviation in Table 2.10. With the low conductivity and ANC this site is vulnerable to acidification even through the relatively neutral pH.

The runoff chemistry of the Sandsuck Formation had the highest mean pH (6.47) among the study sites, and a relatively high ANC concentration of 2.44 meq/L. The more neutral acidity at this site was largely governed by the Sum BC and primarily available calcium and magnesium in

the runoff. Dissolved iron was especially low along with other dissolved metals. The runoff chemistry of the Roaring Fork Sandstone was similar to the Sandsuck Formation. The mean pH of the Roaring Fork Sandstone was 5.88, and the ANC was 3.02 meq/L.

Vegetative Cover: Runoff chemistry appeared to be influenced by vegetative cover based on the four classes assessed: none, grass, dense forest (forest) and sparse forest (stree). Table 2.2 summarized these classes per study site. Three sites Nashville (N1 and N2) and Grainger (G1) were exposed road cuts with no vegetation (none), and differed significantly in runoff chemistry for sites with vegetation: grass, dense forest (forest) and sparse forest (stree). The chemical parameters that differed significantly included: TSS, conductivity, hardness, sulfate, Sum AA, Sum BC, and dissolved metals ($p < 0.01$). The dissolved metals that very dominantly different from no vegetation and vegetation (all cover types) were Al, Cu, and Fe. Other dissolved metals had a few cases where they were not significantly different, such as Mn was not differ between none and sparse forest (Stree) ($p = 0.48$), and grass and forest ($p = 0.08$). Though it appears that vegetative cover influenced runoff acidification, it is difficult to assess the statistical collinearity between geology and vegetative cover. As observed in Figure 2.5, conductivity and hardness were significantly greater for the Chattanooga shale, which included Nashville (N1 and N2) and Grainger (G1) sites, which were also exposed with no vegetation (Figure 2.7).

The effect of vegetative cover can be interpreted by comparing pH and ANC by the two classification groups: geology and vegetative cover (Figures 2.4 and 2.8). In Figure 2.8, pH for the dense forest cover (Forest) was greatest among the classes and significantly differed for sparse forest cover (Stree) and none ($p < 0.01$). However, it was not significantly different than with grass cover ($p = 0.19$). Cover types grass and none were also similar ($p = 0.10$). The low pH for the sparse forest cover also corresponds with the Anakeesta Formation as observed in Figure 2.4 so this vegetative cover class could not be distinguished as unique with respect to runoff pH. The Chattanooga shale had the highest ANC concentrations which also correlated with no road-cut vegetative cover, which also is associated with the high ion concentrations of hardness, Sum AA, and Sum BC. ANC was significantly different between none and both dense and sparse forest covers ($p < 0.01$), and grass and both dense and sparse forest covers ($p < 0.05$). ANC concentrations were not significantly different between dense and sparse forest covers

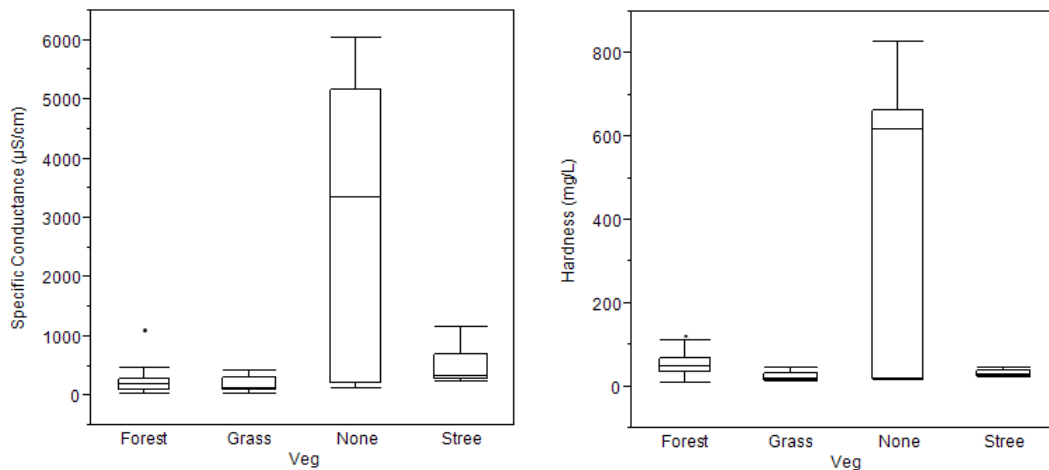


Figure 2.7. Boxplots of runoff chemistry for conductivity ($\mu\text{S}/\text{cm}$) and total hardness (mg/L) for four different road-cut vegetative covers: none (None), grass (Grass), dense forest (Forest) and sparse forest (Stree).

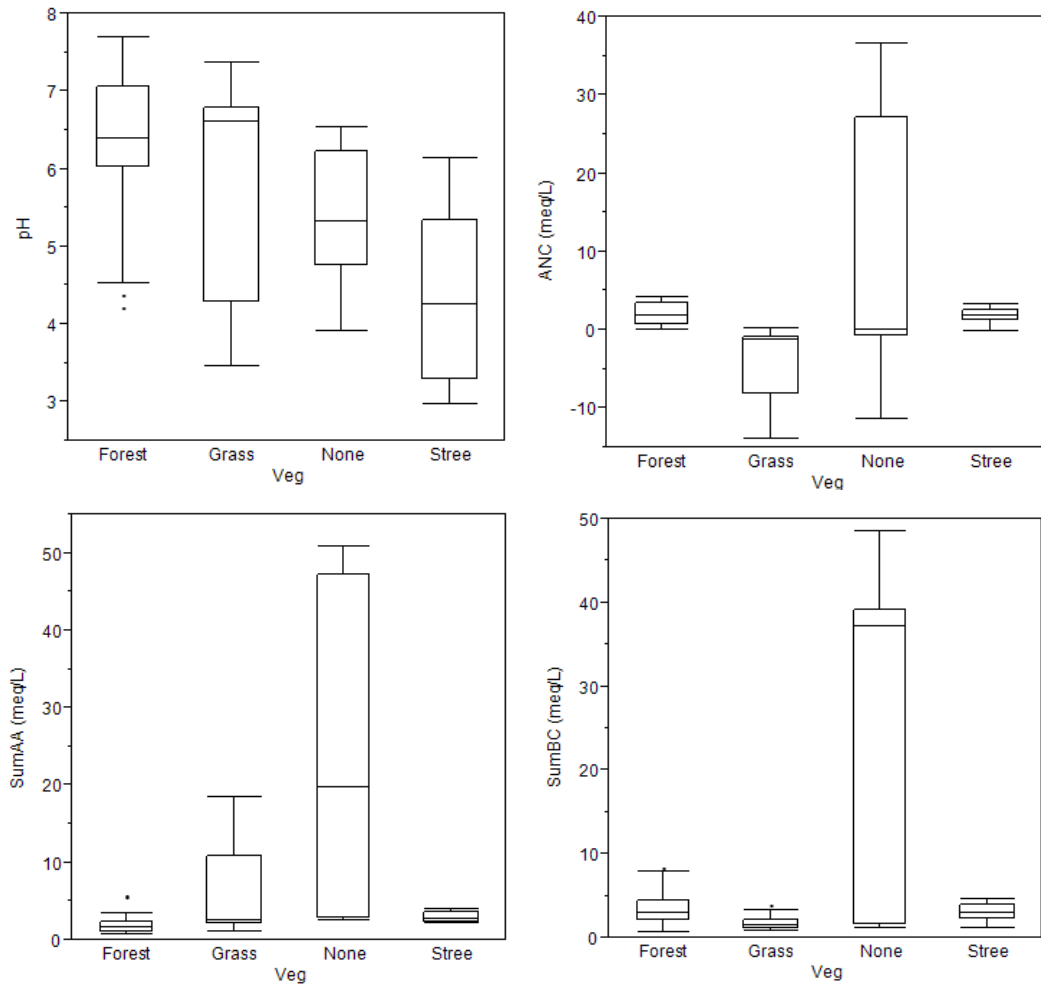


Figure 2.8. Boxplots of runoff chemistry for pH, ANC (meq/L), Sum AA (meq/L), and Sum BC (meq/L) for four different road-cut vegetative covers: none (None), grass (Grass), dense forest (Forest) and sparse forest (Stree).

ranging between 0 and 5 meq/L ($p = 0.93$). Overall, it appears that geology is more dominant than vegetative cover in controlling runoff chemistry but vegetative has an effect. A study specifically designed would need to be implemented to quantify that effect.

Road Cut Age: Runoff chemistry appeared to be influenced by road cut age based on the three classes assessed: old (> 60 yrs), mid (35-60 yrs), and young (< 20 yrs). As observed statistically with the vegetative cover classes, it appears runoff chemistry were strongly governed by three sites Nashville (N1 and N2) and Grainger (G1) which were classified as young, and with no vegetative cover. The sites were also excavated through the Chattanooga Shale. Most the chemical parameters from road cut runoff were significantly different between young and mid-aged sites and young and old sites ($p < 0.05$). The chemical parameters included TSS, conductivity, hardness, pH, chloride (Cl^-), nitrate (NO_3^-), orthophosphate (PO_4^{3-}), Sum AA, Ca^{2+} , Mg^{2+} , Sum BC, and dissolved Si, Ni, Ba, and Cd. Between old and mid-aged sites, pH was also significantly different ($p < 0.01$). ANC concentrations were not differently different among the three classes (Figure 2.9). Dissolved Al was only significantly different between young and old sites ($p < 0.01$). Dissolved Fe was significantly different between young and old sites, and mid

and old site ($p < 0.01$), but not between young and mid-aged sites ($p = 0.68$). Dissolved Mn and Zn were similar to that observed with dissolved Fe.

Road Cut Aspect and Slope: The influence of road cut aspect and slope appeared to have the least potential effect on runoff chemistry because outcomes of the statistical analysis resulted in non-significant relationships except for the north aspect. Similarly, as observed with other environmental factors summarized above, runoff chemistry for road cut aspect appeared to be strongly governed by the two Nashville (N1 and N2) sites which were classified as north (Figure 2.10). These two sites were also classified as no vegetation (none), and young road cut age. In Figure 2.10, hardness concentrations were high for the north aspect class compared with east, west and south reflecting the Nashville site concentrations (Table 2.8). Site collinearity among environmental factors, particularly geology and vegetation affects the statistical analysis. Although several chemical parameters tended to be significantly different between the north aspect and the west, east, and south aspects ($p < 0.05$). The chemical parameters included: conductivity, pH, hardness, ANC, SO_4^{2-} , NO_3^- , Sum AA, Mg^{2+} , Ca^{2+} , K^+ , Sum BC, dissolved Al and Si. Dissolved Fe was not significantly different among road cut aspects. North facing slopes did tend to receive more shading and colder winter temperatures, but the statistics on season did not show any significant patterns. In general, road cut slopes were not significantly different among the chemical parameters in runoff ($p > 0.2$).

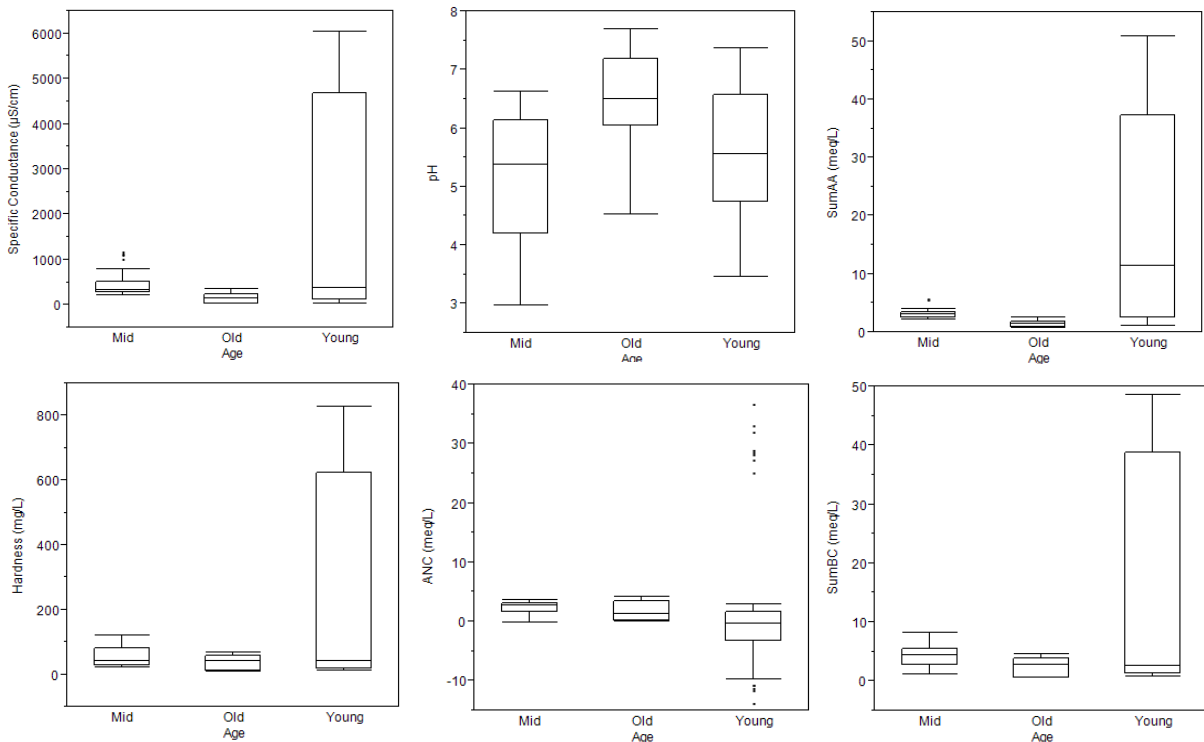


Figure 2.9. Boxplots of runoff chemistry for conductivity ($\mu\text{S}/\text{cm}$), total hardness (mg/L), pH, ANC (meq/L), Sum AA (meq/L), and Sum BC (meq/L) for three different road-cut ages: young, mid, and old.

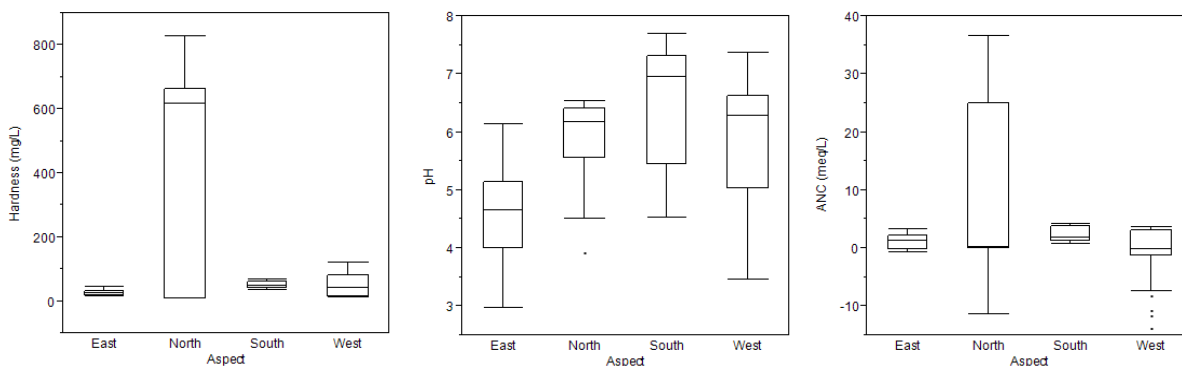


Figure 2.10. Boxplots of runoff chemistry for total hardness (mg/L), pH, and ANC (meq/L) for four different road-cut aspects: east, west, north, and south.

2.3.3 Mass Export of Chemistry Parameters in Runoff from Road Cuts

Mass export of acid anions, base cations, and dissolved metals were computed to quantify the expected loadings and ranges that leave road-cut sites and enter nearby receiving streams (Table 2.11). Per storm event, the chemical parameter concentration (mg/L) was multiplied by the runoff volume (m^3) and divided by the number of runoff days (Appendices D and F). The calculations required a conversion of $g/1000\text{ mg}$ and $1000\text{ L}/m^3$ to obtain units of g/day . These data are needed to assess roadside treatment options because reduction of stormwater runoff acidity is dependent on a mass balance between acid anions, base cations, and dissolved metals, and chemical oxidation, hydrolysis, and different biogeochemical processes.

Sulfate and calcium were the ions measured with the greatest mass exports ranging from < 0.00 to $5,855\text{ g/day}$ and < 0.00 to $1,185\text{ g/day}$, respectively (Table 2.11). The overall median among all sites was much less than the maximums reported for SO_4^{2-} and Ca^{2+} , which were 24.85 g/day and 8.42 g/day , respectively (Appendix F). Sites with the greatest SO_4^{2-} and Ca^{2+} mass loadings were Nashville (N1 and N2), Jamestown (J2), and Sevierville (S1) sites, which road-cut were through multiple rock formations including the Chattanooga Shale, Fentress Formation, and Roaring Fork Sandstone (Snowb). Among these same sites, Mg^{2+} concentrations were higher compared with the other sites. Mass export of dissolved Al was greatest at the Nashville (N2) site with a maximum event of 29.9 g/day , though the overall median among all sites was 0.21 g/day . Mass export of dissolved Fe was generally low with an overall median among all sites was 0.05 g/day , and a range between < 0.00 and 3.30 g/day . Other than dissolved Si, all other metals exhibited low mass export with overall event medians less than 0.15 g/day .

2.3.4 Potential Water Quality Impairment from Road Cut Site Chemistry

Potential water quality impairment to receiving streams from road-cut runoff were assessed by comparing site chemistry to TDEC regulatory limits on specified chemicals, in addition to published toxicity exposure limits (Section 2.2.5; Table 2.7). The dominant concern in runoff water quality is from its contact with rock formations containing APM and the acidification that results from biogeochemical processes. These parameters primarily include pH, ANC, and dissolved Al. Other dissolved metals can be a concern depending on hardness where higher hardness concentrations reduce toxicological effects. TSS was measured in this study and samples were generally below 200 mg/L , which would be typical for surface runoff during rain

Table 2.11. Site mass loading export (g/day) of acid anions, base cations, and dissolved metals per storm event, and summarized by means, standard deviations (StDev), and range. Site identifiers are: G1 is Grainger; J1, J2, & J3 are Jamestown; N1 & N2 are Nashville; O1, O2, & O3 are Ocoee; and S1 is Sevierville. Latitude and longitude coordinates are listed in Table 2.4.

| Chemical Parameter | SITE | | | | | | | | | |
|-------------------------------|--------------------------------|---------------------------------|-----------------------------------|------------------------------|-----------------------------------|-----------------------------------|-------------------------------|-------------------------------|--------------------------------|--------------------------------|
| | G1 | J1 | J2 | J3 | N1 | N2 | O1 | O2 | O3 | S1 |
| Cl ⁻ | 3.05 (2.16) 0.16 – 7.52 | 7.67 (10.56) 0.01 – 34.80 | 22.85 (17.32) 0.01 – 1399.6 | 1.90 (2.54) 0.03 – 9.39 | 9.45 (11.57) 0.84 – 41.99 | 4.90 (3.62) 0.80 – 10.50 | 1.56 (1.96) 0.01 – 7.11 | 1.66 (1.89) 0.01 – 6.53 | 2.71 (3.56) 0.01 – 11.88 | 2.96 (4.20) 0.23 – 12.71 |
| SO ₄ ²⁻ | 45.45 (29.56) 2.31 – 102.28 | 98.42 (155.76) 0.01 – 504.95 | 564.19 (446.16) 0.01 – 1399.61 | 6.38 (7.83) 0.11 – 29.62 | 1345.3 (1619.1) 150.1 – 5855.1 | 268.45 (219.36) 37.21 – 601.20 | 12.16 (15.98) 0.09 – 57.86 | 13.20 (14.63) 0.09 – 49.95 | 37.06 (49.33) 0.01 – 178.76 | 52.59 (73.37) 4.07 – 247.08 |
| NO ₃ ⁻ | 7.56 (5.14) 0.37 – 17.77 | 19.23 (28.23) 0.01 – 92.42 | 44.65 (30.53) 0.01 – 92.42 | 3.72 (4.51) 0.10 – 16.45 | 18.56 (23.56) 1.91 – 85.24 | 11.72 (11.04) 1.67 – 31.82 | 6.61 (9.39) 0.03 – 37.45 | 3.40 (3.95) 0.02 – 13.84 | 6.03 (7.93) 0.01 – 27.03 | 5.89 (8.68) 0.39 – 27.03 |
| PO ₄ ³⁻ | 6.42 (4.22) 0.34 – 14.48 | 15.85 (22.39) 0.01 – 73.58 | 40.75 (26.34) 0.01 – 73.58 | 2.53 (2.76) 0.05 – 10.17 | 12.73 (19.78) 0.56 – 69.95 | 9.16 (5.62) 1.14 – 16.17 | 2.88 (3.78) 0.01 – 13.61 | 2.67 (3.06) 0.02 – 10.92 | 4.86 (6.30) 0.01 – 20.19 | 5.14 (7.53) 0.37 – 23.42 |
| Na ⁺ | 2.78 (2.31) 0.02 – 7.21 | 4.74 (7.21) 0.01 – 23.30 | 8.79 (12.14) 0.01 – 24.24 | 0.65 (0.91) 0.01 – 3.35 | 15.66 (17.83) 0.06 – 64.51 | 23.76 (34.49) 0.01 – 104.29 | 3.25 (4.49) 0.03 – 15.94 | 1.75 (2.64) 0.01 – 9.78 | 13.55 (21.34) 0.01 – 65.38 | 10.05 (19.19) 0.01 – 65.38 |
| K ⁺ | 0.94 (0.66) 0.04 – 2.27 | 8.30 (11.75) 0.01 – 38.64 | 7.60 (12.70) 0.01 – 38.64 | 0.56 (0.71) 0.01 – 2.66 | 0.42 (0.46) 0.04 – 1.69 | 0.88 (0.79) 0.08 – 2.44 | 0.09 (0.12) 0.01 – 0.42 | 0.31 (0.38) 0.01 – 1.39 | 0.90 (1.46) 0.01 – 5.43 | 1.35 (1.97) 0.10 – 6.05 |
| Mg ²⁺ | 4.89 (3.47) 0.20 – 11.46 | 3.53 (4.55) 0.01 – 15.12 | 15.01 (10.01) 0.01 – 29.87 | 0.75 (0.96) 0.02 – 3.57 | 101.14 (111.62) 11.29 – 394.09 | 44.03 (38.31) 6.19 – 109.84 | 4.64 (5.74) 0.04 – 20.84 | 2.69 (3.23) 0.02 – 10.99 | 1.16 (1.68) 0.01 – 6.14 | 8.85 (13.76) 0.92 – 53.76 |
| Ca ²⁺ | 4.74 (3.54) 0.19 – 11.86 | 14.18 (17.68) 0.01 – 59.05 | 26.55 (19.91) 0.01 – 59.05 | 2.88 (3.50) 0.06 – 13.01 | 305.94 (336.99) 39.05 – 1185.7 | 185.84 (142.89) 30.20 – 417.12 | 12.62 (15.39) 0.09 – 57.02 | 8.56 (10.46) 0.05 – 37.64 | 10.81 (13.99) 0.01 – 41.61 | 20.82 (29.66) 1.91 – 113.49 |
| Dis Al | 0.40 (0.31) 0.01 – 1.00 | 0.49 (0.77) 0.01 – 2.49 | 0.89 (0.80) 0.01 – 2.49 | 0.04 (0.06) 0.01 – 0.22 | 1.68 (2.39) 0.01 – 8.51 | 14.75 (10.67) 0.01 – 28.89 | 0.08 (0.11) 0.01 – 0.39 | 0.09 (0.17) 0.01 – 0.65 | 1.86 (2.27) 0.01 – 6.48 | 1.14 (2.05) 0.01 – 6.48 |
| Dis Cu | 0.03 (0.03) 0.01 – 0.09 | 0.08 (0.12) 0.01 – 0.40 | 0.20 (0.15) 0.01 – 0.40 | 0.01 (0.01) 0.01 – 0.04 | 0.31 (0.37) 0.02 – 1.33 | 0.34 (0.25) 0.01 – 0.69 | 0.01 (0.02) 0.01 – 0.06 | 0.02 (0.04) 0.01 – 0.13 | 0.26 (0.39) 0.01 – 1.27 | 0.19 (0.36) 0.01 – 1.27 |
| Dis Fe | 0.09 (0.07) 0.01 – 0.20 | 0.22 (0.41) 0.01 – 1.29 | 0.95 (1.31) 0.01 – 4.86 | 0.02 (0.03) 0.01 – 0.10 | 0.31 (0.42) 0.01 – 1.52 | 0.60 (0.77) 0.01 – 2.34 | 0.01 (0.02) 0.01 – 0.06 | 0.03 (0.04) 0.01 – 0.14 | 0.23 (0.33) 0.01 – 1.17 | 0.20 (0.37) 0.01 – 1.17 |
| Dis Mn | 0.36 (0.29) 0.01 – 0.97 | 0.39 (0.62) 0.01 – 1.98 | 2.22 (1.58) 0.01 – 4.86 | 0.03 (0.04) 0.01 – 0.16 | 1.48 (1.84) 0.07 – 6.67 | 1.80 (1.32) 0.06 – 4.11 | 0.01 (0.01) 0.01 – 0.02 | 0.08 (0.14) 0.01 – 0.51 | 1.05 (1.65) 0.01 – 5.68 | 0.97 (1.76) 0.01 – 5.68 |
| Dis Si | 3.02 (2.10) 0.14 – 7.15 | 0.78 (1.01) 0.01 – 3.37 | 6.05 (4.43) 0.01 – 13.33 | 0.98 (1.33) 0.02 – 4.98 | 4.78 (5.32) 0.67 – 18.88 | 7.40 (5.78) 0.31 – 16.26 | 2.15 (2.71) 0.02 – 9.96 | 0.74 (0.89) 0.01 – 2.95 | 2.33 (2.88) 0.01 – 9.65 | 1.62 (2.80) 0.09 – 9.65 |
| Dis Zn | 0.21 (0.17) 0.01 – 0.57 | 0.32 (0.51) 0.01 – 1.62 | 0.30 (0.54) 0.01 – 1.62 | 0.03 (0.04) 0.01 – 0.16 | 1.24 (1.64) 0.05 – 5.87 | 2.88 (2.14) 0.01 – 5.96 | 0.01 (0.02) 0.01 – 0.07 | 0.06 (0.12) 0.01 – 0.46 | 0.86 (1.39) 0.01 – 4.90 | 0.81 (1.50) 0.01 – 4.90 |
| Dis Ba | 0.06 (0.05) 0.01 – 0.16 | 0.08 (0.12) 0.01 – 0.38 | 0.27 (0.18) 0.01 – 0.51 | 0.02 (0.02) < 0.01 – 0.02 | 0.32 (0.39) 0.02 – 1.43 | 0.16 (0.12) 0.02 – 0.35 | 0.02 (0.03) 0.01 – 0.11 | 0.03 (0.03) 0.01 – 0.12 | 0.18 (0.23) 0.01 – 0.80 | 0.17 (0.27) 0.01 – 0.80 |
| Dis Cd | 0.01 (0.01) 0.01 – 0.03 | 0.01 (0.01) 0.01 – 0.02 | 0.07 (0.06) 0.01 – 0.18 | < 0.01 (-) < 0.01 – 0.02 | 0.06 (0.07) 0.01 – 0.25 | 0.04 (0.04) 0.01 – 0.11 | 0.01 (0.01) 0.01 – 0.02 | 0.01 (0.01) 0.01 – 0.05 | 0.05 (0.09) 0.01 – 0.34 | 0.07 (0.12) 0.01 – 0.34 |
| Dis Ni | 0.05 (0.04) 0.01 – 0.12 | 0.02 (0.03) 0.01 – 0.10 | 0.13 (0.10) 0.01 – 0.30 | < 0.01 (-) < 0.01 | 0.19 (0.23) 0.02 – 0.81 | 0.74 (0.61) 0.01 – 1.71 | < 0.01 (-) < 0.01 – 0.01 | 0.01 (0.01) 0.01 – 0.02 | 0.04 (0.06) 0.01 – 0.23 | 0.03 (0.06) 0.01 – 0.23 |

events (Table 2.8). TSS was observed to be high above 2,000 mg/L at the Nashville 1 and Grainger 1 sites, which likely consisted on eroded silts/sands from the exposed road cuts.

There is no state water quality standard for conductivity, however the USEPA proposed a limit of 300 $\mu\text{S}/\text{cm}$, in which a study by Pond et al. (2008) found impairment to a mayfly family (EPA 2010). Conductivity is a surrogate measure of ions in a water sample, and in waters from the Appalachian coal mining areas the ions of greatest concentrations are typically sulfate and calcium. Hardness is generally correlated with conductivity. Five of the ten sites exceeded 300 $\mu\text{S}/\text{cm}$, with the Nashville (N1 and N2) sites far exceeding this conductivity with site averages over 2,500 $\mu\text{S}/\text{cm}$ (Table 2.8). Runoff from the Nashville sites contained high levels of acid anions, base cations, and dissolved metals.

Site averages for runoff pH ranged from acidic to neutral waters where five of the ten sites exceeding a pH of 6.0 and four sites had average pH values below 5.0 (Table 2.8). Within sites, pH values varied widely, for example the Jamestown 2 site had an average pH of 4.48, but the range was from 3.45 to 7.07. The sites with a pH less than 6.00 are below the state's water quality standards, which pH values are to be between 6 and 9. Treatment options are discussed in Section 4.0.

There is no state water quality standard for ANC, however TDEC recommends for TMDL management a target of $> 50 \mu\text{eq}/\text{L}^{-1}$. In general, streams with ANC less than $0 \mu\text{eq}/\text{L}^{-1}$ would be considered acidic. Four of the ten sites had averages below $0 \mu\text{eq}/\text{L}^{-1}$ and due to the high event variability average ANC did not correlate exactly with their corresponding average pH values. The Nashville 2 site had a high average ANC concentration of 29.85 meq/L resulting from the high concentrations of base cations and dissolved aluminum.

There is no state water quality standard for dissolved aluminum, however literature indicates that concentrations should not exceed 0.2 mg/L (Baldigo and Murdoch 1997; Schwartz et al. 2014). Based on a site average, all road cut sites except the Jamestown 3 site exceeded this toxicity threshold (Table 2.8). Dissolved aluminum in runoff was very high at the Nashville 2 site, indicating a unique geochemical condition under rock/soil acidification. Dissolved aluminum in the form of inorganic monomeric aluminum (Al_{IM}) is regarded as the most toxic dissolved metal for fish and macroinvertebrates in acidified stream waters (Driscoll et al. 1980; Driscoll 1985; Hermann et al. 1993). Fish gill ion transport is disrupted by replacing needed calcium on gill surfaces with increased concentrations of monomeric aluminum (Appendix E). Increased dissolved Al can also cause excessive whole body loss of sodium in trout, resulting in loss of ion regulation. A duration or dose threshold should be considered in particular since stream acidification is episodic in nature (Gagen et al. 1993). Baldigo and Murdoch (1997) found significant mortality of brook trout when dissolved aluminum exceeded 0.20 mg/L for more than two days. Though the road cut dissolved aluminum concentrations were above that threshold, runoff events are generally short and runoff is diluted as it comeslingles with the receiving stream. Most the research has been conducted on trout, therefore how this dissolved Al threshold transfers to other fish species is not well known.

Dissolved iron does not have a state water quality standard, but in high concentrations in groundwater or seeps in which the water is anaerobic then becomes exposed to oxygen will form a $\text{Fe}(\text{OH})_3$ precipitate. This iron precipitate forms an orange-yellow flock (yellowboy) at the bottom of pools. In excessive it may be harmful to benthic macroinvertebrates (Barbour et al. 1999).

Of the dissolved metals analyzed for as part of this study that have state water quality standards, they include cadmium, copper, nickel, and zinc. State CMC standards are as follows cadmium (0.002 mg/L), copper (0.013 mg/L), nickel (0.47 mg/L), and zinc (0.12 mg/L). These standards do not account for hardness. Without accounting for hardness, site averages for cadmium and copper were exceeded for the dataset. In the case of cadmium, it is possible the results are due to the ICP-OES instrument tolerance limit. Five of the ten sites had averages for nickel that exceeded the threshold, and seven sites had averages for zinc that exceeded the threshold. These metals will form precipitates at higher pH levels, thus there are treatment options which are discussed in Section 4.0.

2.4 Discussion and Conclusion

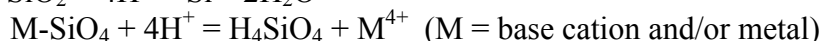
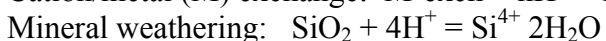
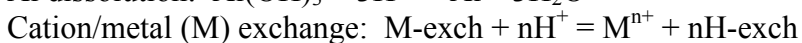
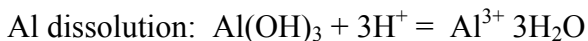
Runoff water quality from road cuts through pyritic rock formations were characterized among ten study sites in Middle and East Tennessee. Sites were monitored for over one year obtaining water samples through all four seasons. Sites were selected to vary environmental factors including geologic rock formation, vegetation cover, and road cut type (age, slope and aspect). In addition to the standard chemical parameters used as indicators, pH, ANC, conductivity, hardness, and TSS, chemical parameters included a comprehensive set of anions, cations, and total/dissolved metals to better understand the biogeochemical processes occurring at the study sites. Very few studies have been conducted on runoff chemistry from road cuts through pyritic rock formations, and none have examined the possible influence of different environmental factors on runoff chemistry. Of the three published articles identified, chemical analyses were limited to pH and total sulfur (Cendro et al. 1977) and pH, total sulfur (sulfate), and a few dissolved metals including Fe and Al (Adams et al. 1999; Orndorff and Daniels 2004).

As with the previous studies (Cendro et al. 1977; Adams et al. 1997; Orndorff and Daniels 2004), pH and other chemical parameters varied widely between sites and within sites among different runoff events. For example, Orndorff and Daniels (2004) reported a pH range from 2.5 to 7.1, dissolved Fe from < 0.1 mg/L to 249.3 mg/L, and dissolved Al from < 0.1 mg/L to 254.5 mg/L for different rock formations in Virginia. Adams et al. (1997) reported pH values between 2.4 and 6.9 for a road cut site in Georgia. In this study, site mean pH also varied widely from 4.29 to 6.27, and the full sample event range between 2.97 and 7.70 (Table 2.8). Other chemical parameters varied widely in this study, and the fact that pH and ANC did not correlate suggests complex and unique biogeochemical processes are occurring among the study sites. Availability of base cations (Ca^{2+} , Mg^{2+} , Na^+ , K^+) from the pyritic formation and/or an adjacent over- or under-burden sedimentary formation can control runoff ANC. Decreased pH can result from the displacement of base cations from the rock with Al complexes, where the base cations act as counter ions to electrochemically balance leached sulfate and other acid anions (Cronan and Schofield 1990; Fernandez et al. 1993; Cai et al. 2010). The lack of available base cations can decrease runoff pH. Other than the dominant oxidation of pyrite (FeS_2) exposed to water and air (Pye and Miller 1990; Rimstidt and Vauhgan 2003; Rohwerder et al. 2003), other key geochemical processes affecting pH and ANC and the overall water quality includes desorption/adsorption of sulfate ions, hydrolysis and precipitation of dissolved metals, silicon and other mineral weathering, cation/metal exchange (Cu, Cd, Zn, Mn, others), and aluminum dissolution and mobilization (Dahlgren et al. 1993; Mitchell et al. 2001; Essington 2004; Palmer et al. 2004; Houle et al. 2006; Sokolva and Alekeeva 2008). These processes can occur at groundwater seeps where water is exposed to oxygen and with the rainfall that becomes in contact with rock surfaces.

The wide range in values of chemical parameters observed in the runoff sampled from this road cut study suggests that other environmental factors may play a role in determining the water quality at any particular location and period of time. If runoff comes in contact with surface soils with organic matter, sulfur and metal adsorption onto organics is possible potentially increasing pH and shifting dissolved metal concentrations (Spratt et al. 1998; Lawrence 2002; Inandar et al. 2004). Also under this environmental condition, microbial activity in organic-rich soils can produce acidity through nitrogen mineralization and nitrification and formation of organic acids (Gobran et al. 1998; Deyton et al. 2009; Gonzalez 2018). Mullholland (1993) notes that depending on the near sub-surface pathways of water initiated from precipitation, water chemistry varies due to differences in controlling biogeochemical processes.

Though many environmental can influence runoff chemical quality at road cut sites, rock formation type appeared to be the controlling factor determining runoff water chemistry (Table 2.10; Figures 2.4-2.6). The mean pH was lowest for the Anakeesta Formation (pH = 4.29), and its mean ANC of 1.83 meq/L was relatively low compared to other. Compared with the other formations, the Anakeesta formation exhibited generally low concentrations of acid anions (sulfate) and base cations, but higher concentrations of dissolved metals including Al, Fe, Cu, Mn, Si, Cd, and Zn. The Chattanooga Shale and the Roaring Fork Sandstone had mean pH values below 6.0 (5.41 and 5.88, respectively). The runoff chemistry from the Chattanooga Shale was much different than all other formations, which sites included Nashville (N1, N2), and Grainger (G1). Runoff from the Nashville sites contains very high levels of acid anions, base cations, and dissolved metals. Site averages for conductivity at the Nashville sites exceeded 2,500 $\mu\text{S}/\text{cm}$, but it was relatively low for the Grainger at 173 $\mu\text{S}/\text{cm}$. The Nashville sites are fairly new road cuts and those sites appear to be weathering at a high rate exporting ions in the runoff compared to the other sites, particularly the higher concentrations of dissolved silicon. Overall runoff from the Chattanooga Shale contained high levels of dissolved Al, Ba, Cu, Zn, and Ni. The runoff chemistry of the Roaring Fork Sandstone was more similar to the other geologic formations. The Fentress and Sandsuck formations had mean pH values above 6.0. Runoff chemistry from these two formations was similar to the other rock formations except for the Chattanooga Shale as noted above.

With the observed differences of dissolved metals among the rock formations it suggests the basic mineral composition of the local rock appears to greatly influence runoff chemistry; however no data were available on the exact mineral composition of the study site's rock formations. Other published studies observed a few correlations sulfur content, but none completed a comprehensive analysis of mineral and metal composition. For example, Centrero et al. (1977) correlated %S with pH, where %S estimates below 0.2% were generally above a pH of 5 to 6. Orndorff and Daniels (2003) conducted PPA and %S on rock but there was no significant correlation with runoff pH and dissolved metals. In some cases, PPA values above 25 mg/L (CaCO_3 equivalence) results in higher concentrations of dissolved metals. A more comprehensive examination of mineral content could provide additional information on the geochemical processes influencing runoff acidification such as mineral weathering and aluminum dissolution, hydrolysis and precipitation of dissolved metals and silicon, and cation/metal exchange. These chemical reactions are defined as the following:



Though chemical reactions on the rock surface associated with weathering, Adams et al. (1999) observed that the comingling of waters from seeps and runoff from different rock surfaces affected water chemistry. It is evident that the waters leaving road cut sites potentially undergo several biogeochemical transformations influencing the pH, sulfate, and dissolved metals. As with others, they observed the iron precipitate yellowboy in ponded pools of water at the road cuts. Many environmental factors appear to influence road cut runoff, and further study is needed to more specifically quantify chemical relationships between the parent minerals and runoff water quality, and the comingling of flow paths through different rock types. .

In addition to rock formations affected runoff water quality at road cut sites, it appears that vegetation cover or the lack thereof can possibly influence runoff chemistry to some degree (Figures 2.7 and 2.8). With no vegetative cover with exposed bare rock, ion concentrations in runoff were much greater compared to sites with grass and/or tree cover. Figure 2.7 shows the differences among cover types for conductivity and hardness, and Figure 2.8 shows the differences for ANC, Sum AA, and Sum BC. Median pH for no cover was approximately 5.3. The highest median pH among the cover types was with trees at 6.4, which included the Ocoee (O1 and O2) sites, Jamestown (J3) site, and Sevierville (S1) site. This outcome suggests that organic matter availability from the trees may be playing a role in metal/sulfate adsorption, and sulfur, nitrogen, and cation cycles (Cape et al. 1992; Johnson and Lindberg 1992; Draaijers et al. 1997; Mitchell et al. 2001, Barker et al. 2002; Holzmüller et al. 2007; Cai et al. 2010). Tree cover also protects the rock surface from low intensity rainfall event per leaf interception and evaporation.

The other environmental factors investigated included rainfall event volume and intensity, seasons, and road cut types (age, slope, and aspect). Runoff chemistry was not correlated with rainfall volume and storm maximum intensity, but average intensity was found to influence the export of sulfate, Sum AA, and Sum BC indicating higher rainfall intensities potentially flush ions from road cuts. Seasons did not have any significant effect on runoff chemistry among the study sites. Though our study did not find many relationships with season and weather-related controls, some evidence indicates these variables could influence runoff chemistry from APM (Olyphant et al. 1991). More recently constructed road cuts (young) were found with runoff chemistry with higher concentrations of anions, cations, and dissolved metals. In particular, 'young' sites were higher in dissolved Al and Si indicating mineral weathering rates to be greater compared to the middle and old age sites. The difficulty in interpreting the potential effect of the factors is that data co-varies with the more dominant factors of geologic formation and vegetation. More data from a larger set of study sets would be needed to statistically evaluate these factors. Overall, the findings do suggest that there are minimal effects from seasons, storm event magnitude, and road cut characteristics of slope and aspect. Also, it suggests the older the road cut the likelihood of acidic runoff is small and that planting vegetative cover at runoffs can be used as an effective management strategy to minimize acidic runoff.

Potential environmental impacts of runoff water quality from road cuts through pyritic geologic formations were investigated by comparing the analyses of collected waters to the state's Water Quality Standards. The water collected was directly from the road cut as shown in Figure 2.2 and Appendix A, which does not necessarily infer it's the water chemistry entering the receiving stream. As indicated by Adams et al. (1999), acidity of road cut runoff was greatly reduced prior to entering stream, and there were no downstream effects measured. Their study did not analyze for base cations or many of the dissolved metals so understanding why road cut water acidity was neutralized was left unknown but likely was due to the addition of calcium and magnesium ions. In this study, pH was measured below 6.0 often (Table 2.8). Even sites with a

mean pH above 6.0, single events included pH measurements below 6.0. The Tennessee Water Quality Standards require effluent discharges to be between 6.0 and 9.0, unless outside this range represents locally natural conditions. Runoff conductivity was very high at the Nashville road cut sites with a mean above 2,500 $\mu\text{S}/\text{cm}$, which levels could negatively impact certain benthic macroinvertebrates (Pond et al. 2008). In sedimentary rock formations conductivities in the range of 400 to 700 $\mu\text{S}/\text{cm}$ are commonly observed. Though there is no Standard for dissolved Al, several studies indicate that levels above 0.2 mg/L can be harmful to fish and some benthic macroinvertebrates (Driscoll et al. 1980; Hermann et al. 1993; Galen et al. 1993; Baldigo and Murdoch 1997). The presence of dissolved organic carbon (not measured) can greatly reduce the toxicological effect of dissolved Al (Driscoll 1985). The study findings observed road cut site exceedances above state Standards for two dissolved metals. State CMC standards for dissolved metals are: cadmium (0.002 mg/L), copper (0.013 mg/L), nickel (0.47 mg/L), and zinc (0.12 mg/L). These standards do not account for water hardness. Without accounting for hardness, site averages for cadmium and copper were exceeded for the dataset. The sites with dissolved metals exceedances were also the sites with higher hardness concentrations which reduce or eliminate the toxic effects of cadmium and copper.

Runoff water quality for this study was similar to that found in the Orndorff and Daniels (2004) study for 25 Virginia sites, inferring water quality from road cuts through pyritic rock formations have expected ranges and event episodic variabilities. The need for site treatment needs to consider whether there is a source of base cations from an overburden or underburden sedimentary rock formation and its exposure to rainfall (Miller et al. 1976; Sobek et al. 2000). For example, a well-vegetated road cut with an adjacent limestone formation would be a low risk site for acidic runoff. In contrast, a vertical road cut bare of vegetation completely through a pyrite shale formation could generate runoff pH below 6.0, low ANC, low hardness, and higher concentrations of dissolved metals, which has a high risk for biological impairment in receiving streams. Findings from the study, coupled with published methods for AP, NP, NNP and PPA could be applied in developing a risk map for runoff acidity and toxic metals. As observed from this study (Table 2.8), runoff chemistry was at levels/concentrations that are treatable with passive methods (Morgan et al. 1982; Gazea et al. 1996; Webb et al. 1998; Gibert et al. 2005; Johnson et al. 2005; Sheoran and Sheoran 2006; Nyquist and Greger 2009; Ray et al. 2009; Cruz Viggli et al. 2010). The potential for runoff treatment options at pyritic road cuts will be described in more detail in Chapter 4.0.

3. Experimental Testing of Treatments for APM Highway Cut Slopes

3.1 Introduction

Acid producing material (APM) removed from highway construction projects through pyritic formations must be handled according to special disposal requirements. These requirements are described in the Tennessee Department of Transportation document Special Provision 107L and standard drawings for material treatment (TDOT 2007). Though APM handling has been well developed, little was known about acidification of runoff from the completed road cut surface. A few studies examined runoff water quality from exposed pyritic rock surfaces at road cuts and natural land surfaces (Miller et al. 1976; Morgan et al. 1982; Fox et al. 1997; Igarishi and Oyama 1999). These studies found water to have high levels of free acidity ($\text{pH} < 5$), and elevated levels of dissolved metals (Al, Cu, Fe, Mn, and Zn). Yellowboy and reddish-brown iron precipitates forms with sufficient availability of oxygen. A study by Orndorff and Daniels (2004) is the only study closely related to characterization of water quality from road cuts in the Appalachian region. They found ranges in pH from 2.5 to 7.1, and total sulfur from 7 to 543 mg/L. Dissolved metals in the runoff were highly variable between sites and storm events. Similar results were found in this study (Chapter 2) with ranges in pH from 4.29 to 6.76. Sulfate concentrations ranged from 17.6 to 2,084.9 mg/L. Results indicated that variability was primarily dependent on the geologic formation in which the road cut was constructed and whether vegetative cover was present. Road cut age appeared to be another factor to runoff acidification, with older road cuts being less acidic. It appeared from this study that vegetative cover could be an effective design and/or management strategy to reduce the potential for acidic runoff leaving highway road cuts. In addition to vegetative cover, there is a research need to explore other options for road cut design to prevent acidic drainage from road cuts.

In order to provide TDOT additional design options, an experimental study design was developed to explore the possible use of different cover materials over exposed pyritic rock. As noted above, vegetation was one material, but needs to include soil and vegetation together. Soil at the rock interface may serve as a sink for sulfate ions if soil water pH remains low and sulfate concentrations are elevated from microbial processes at the surface. Soils also provide a source of base cations to neutralize soil acidity. Soil and vegetation reduce exposure of the pyritic rock surface to rainfall, and vegetation consumes nitrate therefore reducing soil acidity. In general, the biogeochemical processes associated with this cover are complex and rely on natural environmental factors such as seasons (cold/hot), and moisture (wet/dry). Highways departments have used shotcrete as a stabilizing material for rock road cuts to prevent weathering and spalling (Qiao and Zhou 2017; Hayward Baker 2018; Hitech Rockfall 2018; Prometheus 2018). Shotcrete application is the spraying of concrete, Portland cement, aggregate and water, by a pneumatic pressurized gun or nozzle applied to a thickness of three to six inches, which can be unreinforced or reinforced with welded wire mesh or steel fibers. An innovative highway project used shotcrete and vegetation cover to form a 'green wall' and effectively stabilize a vertical rock slope (Medla et al. 2017). Not only can shotcrete stabilize rock surfaces, its use can neutralize runoff acidity because it's a silica calcium based product. Geosynthetic membrane liner products are many, and have been used for many applications (Fettig 2018; Geosynthetica 2018). Solmax (2018) integrated a geoliner and grass vegetation system for construction sites (GSE Lite Earth™). Most applications at highway construction sites are on soil surfaces with moderate slopes. Few examples in literature were found where geoliners were used on vertical rock surfaces. If geoliners are to be used on vertical rock surfaces engineered attachments will be needed to insure long-term durability of that material.

The objective of this study was to test possible designs for road cut cover materials to prevent precipitation contact with exposed pyrite geology and limit export of acid pollutants in runoff. The materials selected for study included soil/vegetation, shotcrete, and a geosynthetic membrane liner. The project Task 2 consisted of a field laboratory-scale experiment, which was modified from the originally proposed full-scale highway road-cut application and testing. The reason for the task scope change was the manufacturer declined to guarantee their product and its application on pyritic rock road cuts. This study did not assess cover material performance for durability only its ability to reduce acid runoff.

3.2 Methods

3.2.1 Experimental Set-up

The experimental design included constructing four container panels to place pyrite rock, and three of the four panels treatments were installed to limit or prevent exposure to the rock. The treatment types were: 1) soil/vegetation cover, 2) shotcrete, and 3) a geosynthetic membrane liner. One panel with exposed pyrite rock (no treatment) served as the experimental design control. The basic construction design of the container panel is shown in Figure 3.1. Four container panels were constructed on UTK’s Facilities Planning Division property (5723 Middlebrook Pike, Knoxville, Tennessee 37996-0040). The study site was located at an outdoor location on the property so that the pyrite rock was exposed to sunlight and rainfall that would naturally occur at a road cut. A rainfall gauge was installed at the study site. The study design included a two-phase rainfall exposure methodology. The first phase consisted of exposing the four panels with artificial rainfall from a precipitation simulation tower (Figure 3.2). The second phase consisted of exposing the four panels to natural rainfall for about a six-month period.

Anakeesta pyritic rock collected from a rockslide which occurred on May 28, 2015 in the Great Smoky Mountains National Park (with National Park Service permission) was used for this study. Rock pieces varied in size from approximately 0.5 ft to 2-ft A-axis length. Prior to experimental implementation, the collected rock was stored in a cool, dry indoor location on the UTK campus. The rock was geochemically tested by TDOT, and the collected rock was considered pyritic per Golder Associate (2007) document guidelines, with properties determined as follows:

| Paste pH | Neutralization Potential (Tons/1000 Tons) | Total Sulfur (%) | Potential Acidity (Tons/Acre) | +CaCO ₃ (Tons/1000 Tons) | Total Pyritic Sulfur (%) |
|----------|---|------------------|-------------------------------|-------------------------------------|--------------------------|
| 4.3 | 35.50 | 2.670 | 83.44 | -47.94 | 0.94 |

The four container panels constructed of wood lumber, plastic gutter runoff collectors, and rainfall simulators were constructed in April 2016 (Figure 3.2). The container panels measure approximately 3-ft by 4-ft in area, with 0.5 ft sidewalls and a sample port at the bottom (Figure 3.1). The panels were inclined at approximately 15° slope allowing for effective drainage and a longer water contact time than with steeper slopes. The Anakeesta rock was placed in each panel container with no mortar or binding material applied between the rocks so not to change the runoff chemistry during the rainfall events. The rock pieces were placed to minimize gap size. Surface areas of the exposed rock to rainfall in each panel were estimated using scaled photographs. The soil/vegetation treatment consisted of turf grass sod purchased at Home Depot, and placed over the rock. The shotcrete treatment used was mixed according to standard AASHTO T.22 materials as described by Part 6, Section 622 of the TDOT Standard Specifications for Road and Bridge Construction document (Januray1, 2015). This material

consisted of 4,000 psi concrete with no aggregates, and was applied on the rock surface filling spaces between the rocks, and forming a smooth surface $\frac{3}{4}$ inch in thickness. The geosynthetic membrane liner consisted of black 36 mil (30 ounce) high pressure liner (HPL) of woven polyester fabric and PVC/Elvaloy/ KEE film manufactured by EPT XTRAM®, which completely covered the rock surface in the panel container. The EPT XTRAM® HPL specifications are in Appendix G.

The rainfall simulator consisted of an 18-ft tall steel tower with PVC water supply pipes and special discharge nozzles to simulate natural rainfall droplets (Figure 3.2). The water supply consisted of city tap water filling a small open surface plastic tank, set for one week for dechlorination, and a 1.5 HP Gould submersible pump was used to delivery water to the nozzles. The nozzles were 1/8-inch diameter size from Spraying System Company (Model # GG-SS 4.3W). These nozzles had a 0.078-inch nominal diameter and a wide angle spray of 120°. The simulation system operated at 10 psi pressure delivering 0.43 gpm, which equates to 2.55-inch depth for a day (approximately 0.21 inch/hr) on the 12 ft² container panel area. This flow rate was selected to reflect a storm event intensity of a 1-year 24-hour return frequency for the Knoxville area. Plastic sheeting was placed on the simulation tower as shown in Figure 3.2 to eliminate spray dispersion from wind.

The simulated rainfall experiments at a rate of approximately 0.21 inch/hr were conducted for a period of 4.317 hours for an initial test, followed by one-day rest, and then run for a period of 2.383 hours. This simulation was conducted for the pyrite rock and soil/vegetation panels. The sample frequency for runoff water was: 0.00, 0.833, 1.667, 2.383, 3.117, 3.833, 4.133, and 4.317 hours for the initial test. The frequency for the second run on the second day was: 0.00, 0.833, 1.667, and 2.383 hours. The simulation for the shotcrete and geoliner treatments, and sample frequency for monitoring was a single day for 0.00, 0.833, 1.667, 2.383, 3.117, 3.833, and 4.133. Water from the source tank pool was collected at the beginning and end of each simulation experiment. Runoff water samples from each simulation and test and container panel were analyzed for chemical parameters at UTK Water Quality Laboratory. The source tank water was also analyzed for chemistry.

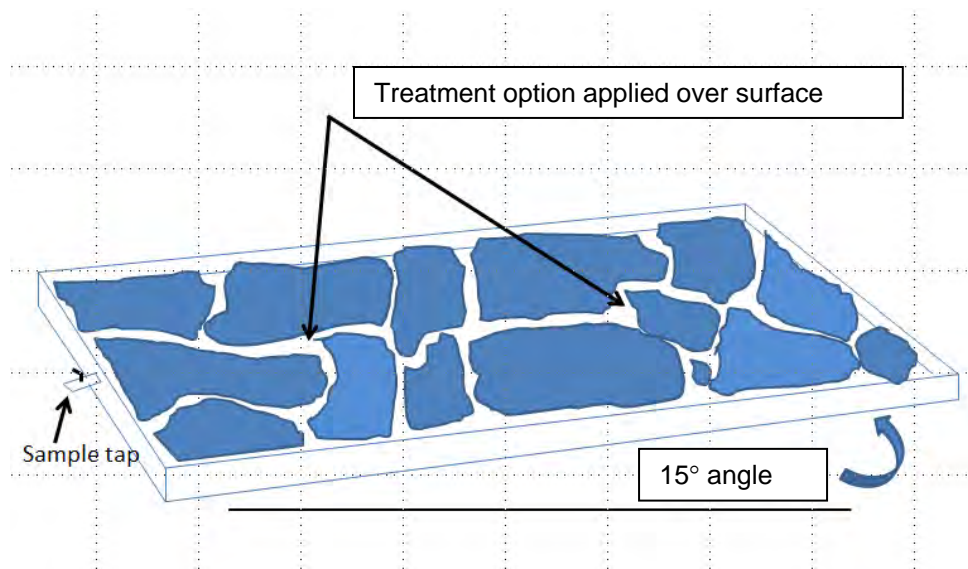


Figure 3.1. Schematic of the experimental test panel with pyritic rock.



Figure 3.2. Photos of the experiment set-up on Middlebrook Pike, Knoxville, Tennessee showing Anakeesa Formation rock treatment trays (*upper right photo* – exposed rock and *lower right photo* – soil/grass), and the rainfall simulator (*lower left photo*).

After the stimulated rainfall experiments, the four test panels were exposed to natural rainfall from June 2016 through December 2016, where runoff water samples were collected and samples analyzed at the UTK Water Quality Laboratory. Runoff volumes from natural rainfall for each of the treatment panels were measured by water depth measurements in the bucket. For very large storms exceeding the bucket capacity, a linear regression relationship was developed from the measurements within the bucket's capacity (Table 3.1). This experiment was conducted for a six-month period to investigate treatment performance under natural climatic conditions. Water chemistry data of the runoff for both the stimulated rainfall experiment and natural rainfall can be found in Appendices H and I, respectively.

Table 3.1. Linear regression relationships for the four 12-ft² area panels, one control and three treatments. Runoff volume (∇) is in ft³, and rainfall depth (RD) is in inches.

| Treatment | Regression Equation | Correlation | Significance |
|-----------------|--------------------------------------|---------------|--------------|
| Pyrite Rock | $\nabla r = -0.01720 + 0.96575 (RD)$ | $R^2 = 0.988$ | $P < 0.001$ |
| Soil/Vegetation | $\nabla v = -0.01431 + 0.57838 (RD)$ | $R^2 = 0.934$ | $P < 0.001$ |
| Shotcrete | $\nabla s = -0.00048 + 0.96575 (RD)$ | $R^2 = 0.961$ | $P < 0.001$ |
| Geoliner | $\nabla g = -0.00002 + 0.71862 (RD)$ | $R^2 = 0.969$ | $P < 0.001$ |

Note: the y-intercept in the linear regression equation represents hydrologic initial abstraction.

3.2.2 Water Quality Chemical Analysis

Water analyses followed American Water Works Association (AWWA) Standard Methods for use of an ion chromatograph (Dionex™, IC), inductively-coupled plasma optical emission spectrometer (ThermoFisher™, ICP-OES), wet chemistry titration procedures for pH and conductivity. U.S. Environmental Protection Agency (USEPA) and AWWA Standard Methods are the same, but test numbers differ; a cross-referencing list of tests is provided in Table 2.6. Water quality analyses include the following chemical parameters: pH, specific conductance, sulfate, nitrate, chloride, calcium, magnesium, and dissolved metals (iron, aluminum, manganese, silica, copper, and zinc).

3.2.3 Data Analysis

Simulation rainfall and runoff chemistry data were summarized per experimental treatment and the control as chemical parameter concentration plots over time to qualitatively observe changes during the simulation event. The main goal was to observe how long pyrite oxidation would continue and influence runoff pH, and the export of sulfate and iron from the exposed rock surfaces. Per chemical parameters, chemical units/concentrations were plotted against experimental time to examine the change over the simulation period. A qualitative assessment of the patterns observed was interpreted with regards to the effectiveness of the treatments by comparing the control (pyrite rock panel) with the three different treatments.

Natural rainfall and runoff chemistry data were summarized per experimental treatment and the control for each storm event measured. Statistical analysis was conducted on event mean concentrations per treatment using a single-factor ANOVA to test whether collectively there was a significant difference among the four treatments. In order to test significant differences between individual treatment combinations, a two-sample t-test (assuming unequal variances) was used. Chemistry from the rainfall water was also statistically compared to the treatment runoff chemistries. In addition, runoff volumes were compared with runoff chemistry to assess the influence of storm event magnitude on the chemistry. A regression model was used to assess whether runoff chemistry was influenced the volume magnitude of the storm event.

3.3 Results

3.3.1 Simulated Rainfall Experiments

As would be expected the runoff chemistry of the control panel with the exposed pyrite rock changed over the simulation period. The pH increased from approximately 3.6 to above 6.0 within four hours (Figure 3.3). However, rapid oxidation occurred where the next day the initial

pH was near 3.6 however reached pH values above 6.0 within two hours. Runoff ANC from the pyrite rock remained low during the simulations due to the lack of base cations to neutralize the acid anions (Figures 3.4-3.6). The sum of acid anions (Sum AA) consists primarily of sulfate, and the rapid decrease in SO_4^{2-} concentrations can be observed in Figure 3.7. Mineral ions from the rock surface are quickly washed off as observed by the rapid decline in conductivity within the first hour of rainfall and basically diminished of ion source within two hours of rainfall (Figure 3.8). This rapid decline in dissolved metal ions can be observed for iron and aluminum in Figures 3.9 and 3.10. These two metal ions are the most relevant to pyrite oxidation and the potential for surface water toxicity, however all the metal ions were found to decline throughout the rainfall simulation period (Table 3.2). The dissolved aluminum concentration dropped below 0.2 mg/L with two hours. These plots from the rainfall simulator experiments demonstrate that runoff chemistry from exposed pyrite rock changes rapidly with acidification conditions occurring for about an hour for an intensity of approximately 0.21 inch/hour (equivalent to 1-year 24-hour return frequency, Knoxville area).

The soil/vegetation treatment using turf grass sod purchased commercially at Home Depot results in runoff chemistry changes for ANC, sum of base cations (Sum BC), Sum AA, conductivity, and sulfate; however, runoff pH remained approximately the same during the experimental runs at about 6.8 (Figures 3.4-3.10). ANC declined on day 1 (D1), but increased on day 2 (D2). The chemistry with the turf grass is complicated with sum of base cations remaining constant on D1 but declining rapidly on D2, whereas the sum of acid anions was observed to decline on both days. The complex chemistry for this treatment was likely due to the nitrogen-phosphorus-potassium fertilization and calcium liming of turf sod. Overall, the runoff pH was moderated from the soil and grass vegetation compared with the pyrite rock control. However, in order to better understand the chemistry and the potential use of grass sod as a treatment further research on the biogeochemical processes would be needed. It appears some sulfate and potassium were present with export occurring at different rates, and mineralization/nitrification occurring on the second day with moist soil conditions for the microbial community (Table 3.2).

The runoff chemistry for shotcrete and geoliner treatments remained relatively constant through the first day of rainfall simulations (Figures 3.4-3.10). Because of that observation, a second day of simulations was not conducted. The runoff pH was approximately 9.0 and 7.8 for the shotcrete and geoliner treatments, respectively. The runoff ANC was approximately 2.0 meq/L and 1.3 meq/L for the shotcrete and geoliner treatments, respectively. ANC was generally low because it essentially consisted of rain water with low ion content. The sum of base cations and conductivity were elevated initially for the shotcrete treatment, but quickly declined after one hour, which was due to high levels of calcium being washed off (Table 3.2). This result would be expected because shotcrete is a calcium-based product. Both these treatments would be effective controls for potential acidification of runoff from exposed pyrite rock along newly constructed road cuts.

A summary table of the runoff chemistry for the pyrite rock panel (control), the three treatment panels (soil/vegetation, shotcrete, and geoliner), and the water supply (source) tank is expressed as minimums and maximums (Table 3.2).

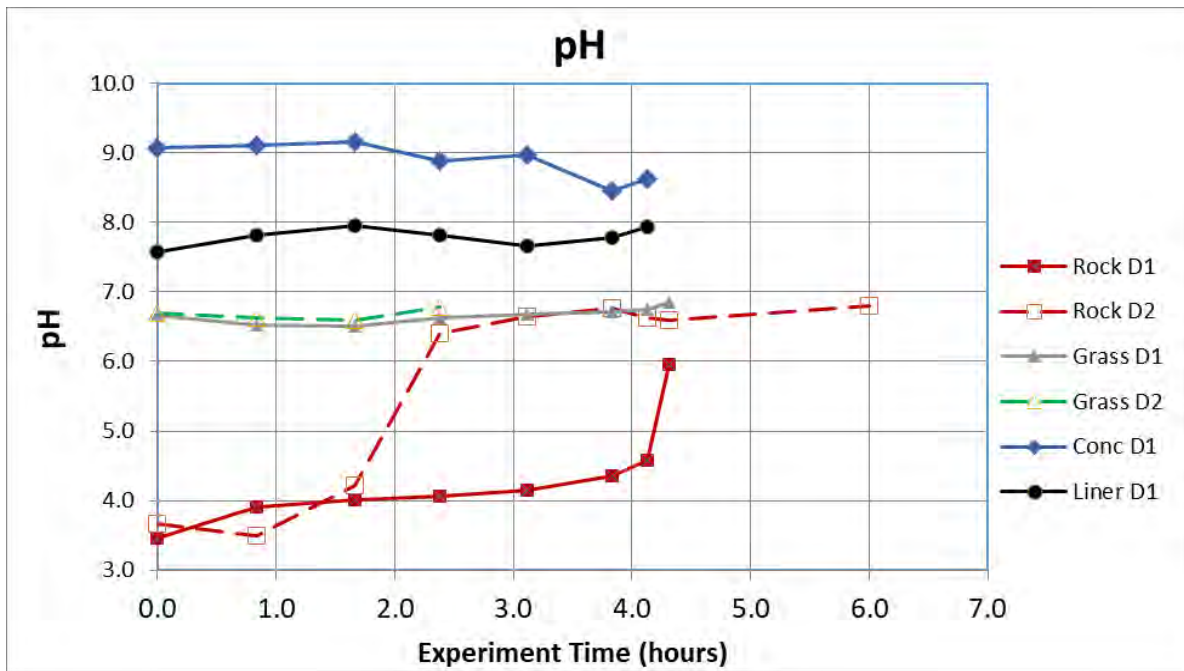


Figure 3.3. Runoff pH from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2).

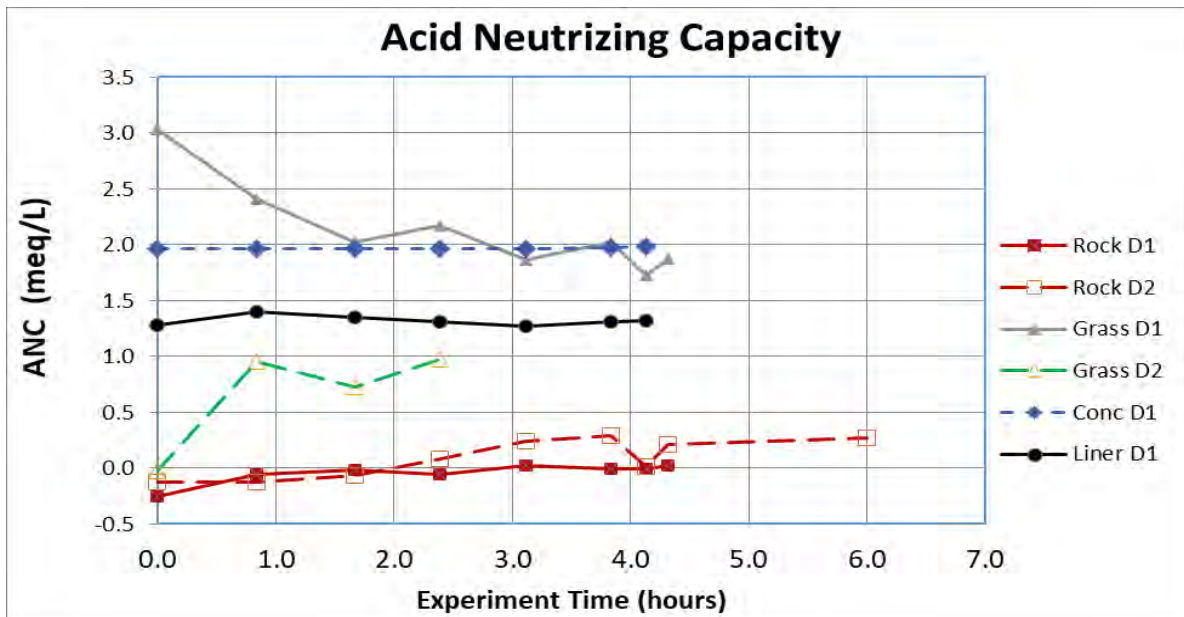


Figure 3.4. Runoff ANC (meq/L) from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2).

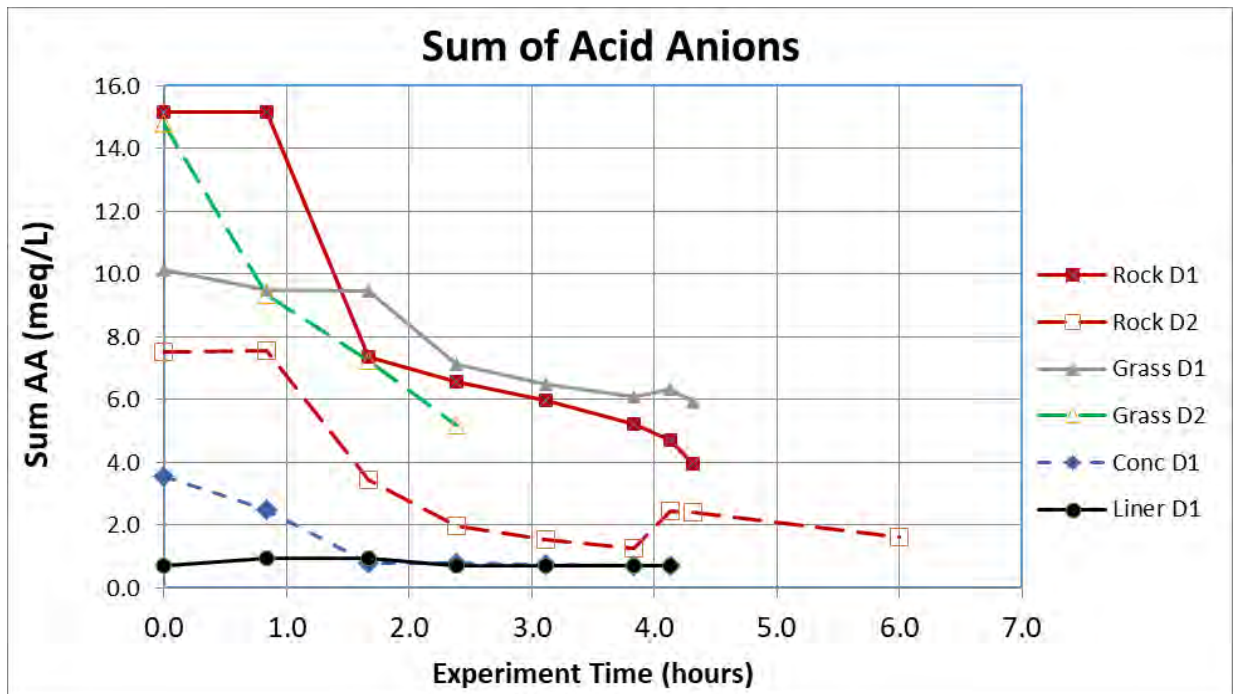


Figure 3.5. Runoff sum of acid anions (meq/L) from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2).

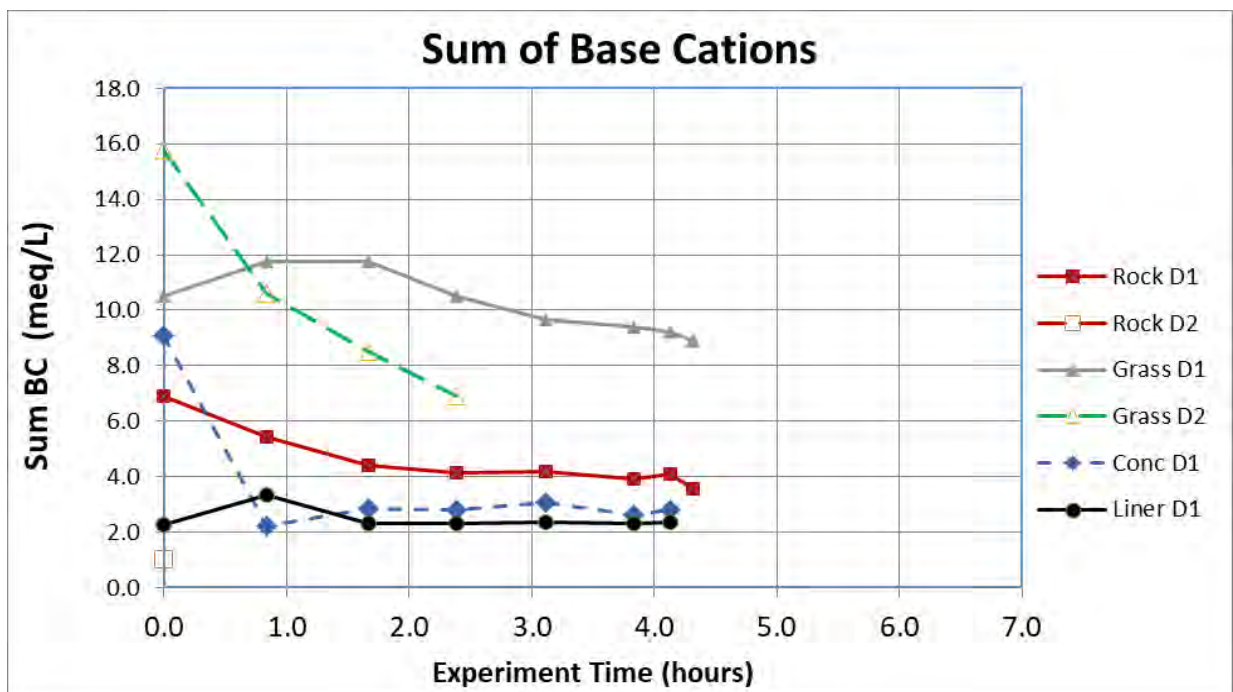


Figure 3.6. Runoff sum of base cations (meq/L) from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2).

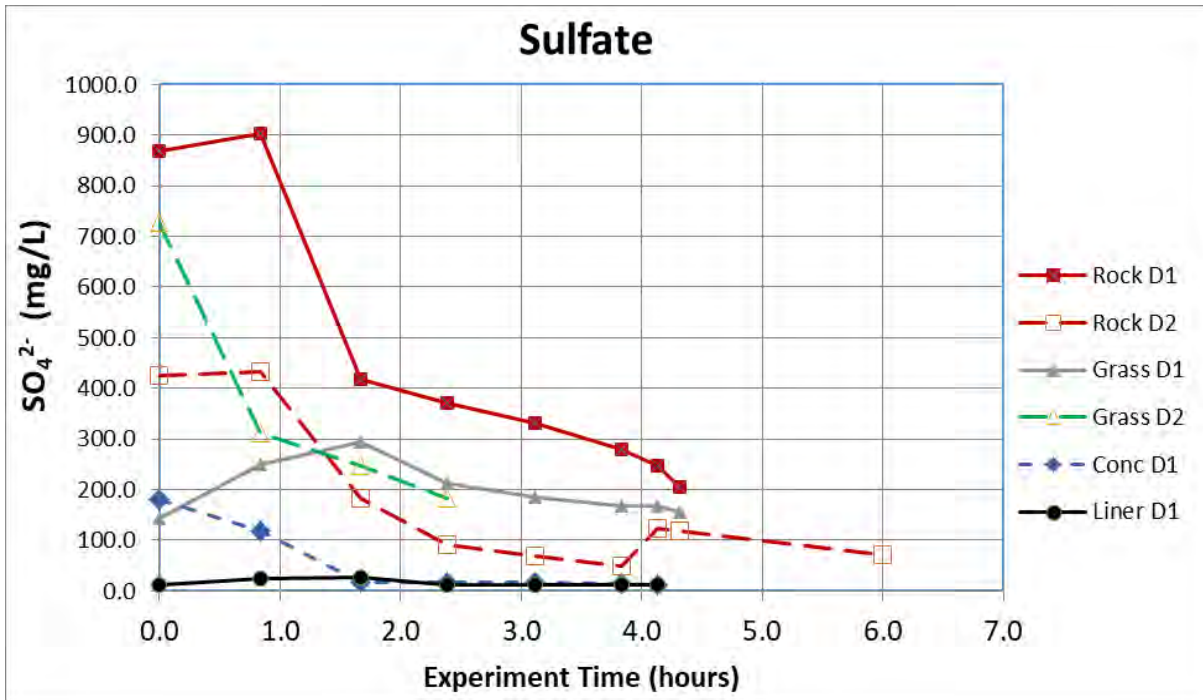


Figure 3.7. Runoff sulfate (mg/L) from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2).

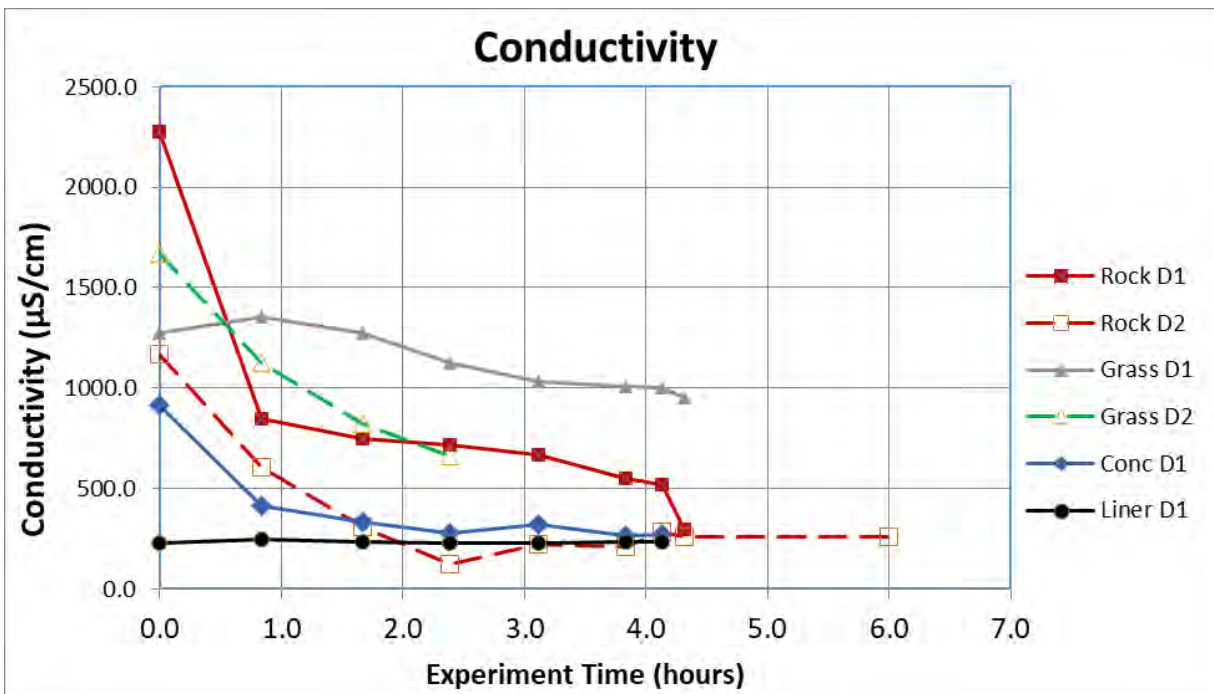


Figure 3.8. Runoff conductivity ($\mu\text{S}/\text{cm}$) from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2).

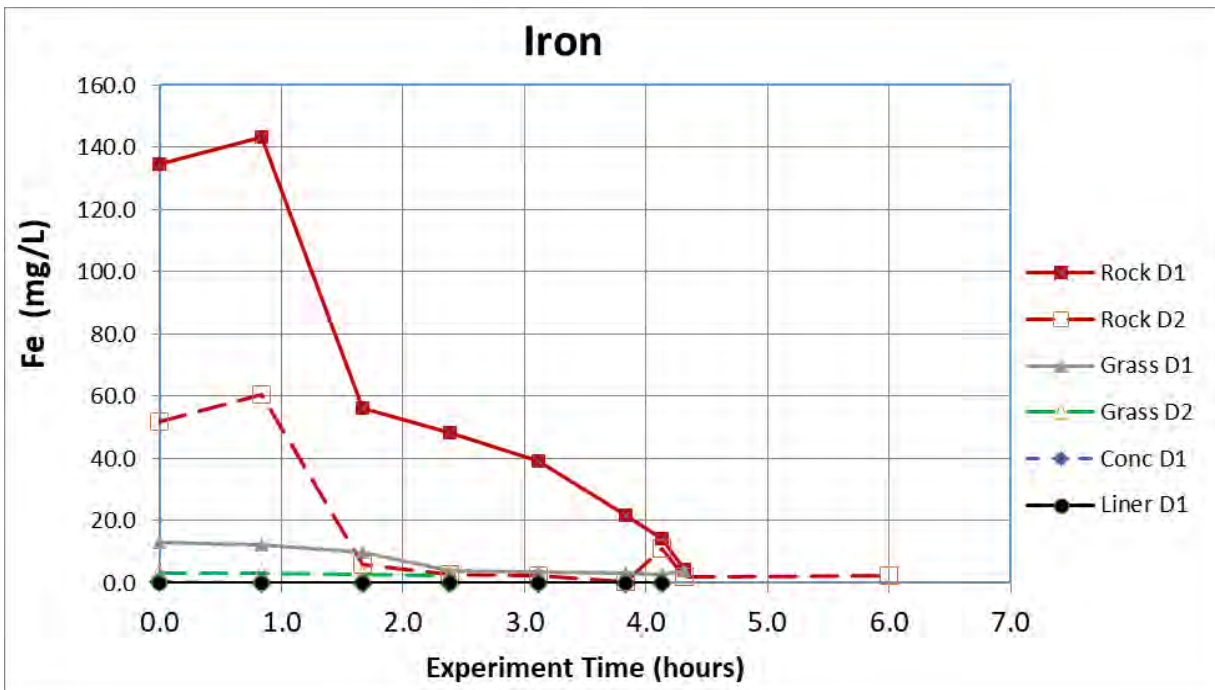


Figure 3.9. Runoff dissolved iron (mg/L) from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2).

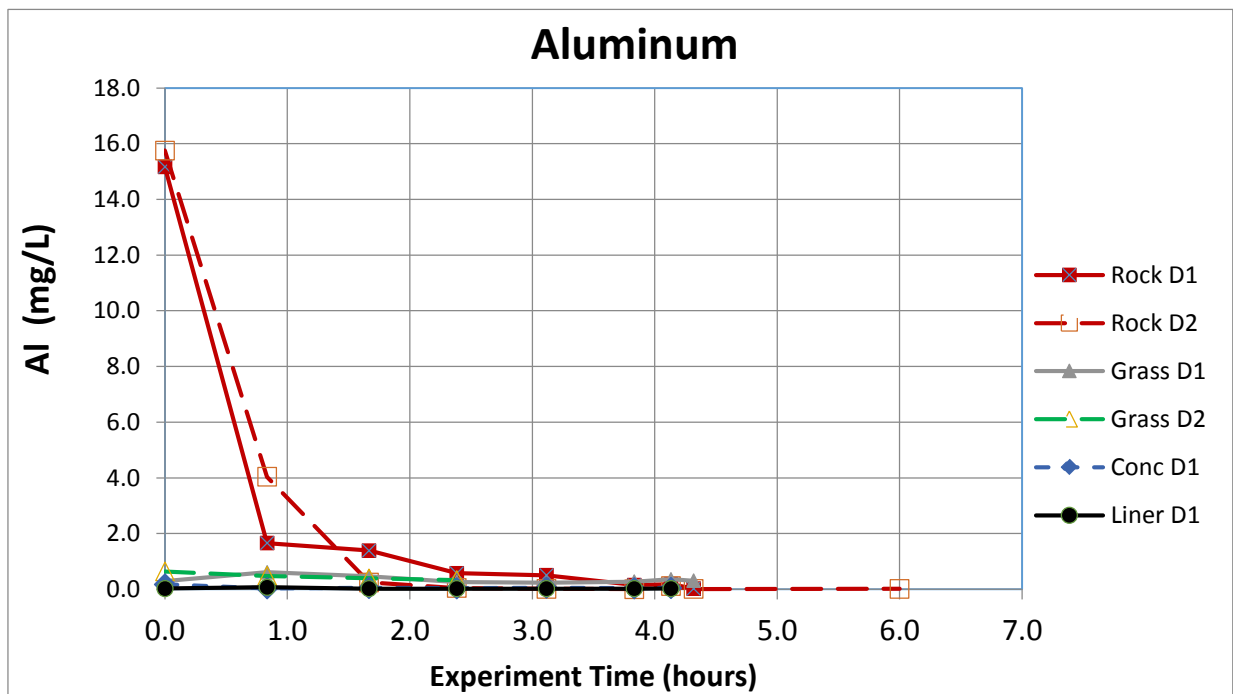


Figure 3.10. Runoff dissolved aluminum (mg/L) from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner) for day 1 (D1) and day 2 (D2).

Table 3.2. Summary of runoff chemistry from rainfall simulator experiments for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner). Water chemistry from the tank source was reported (Pool). Summary reported as the range (maximums and minimums).

| Chemical Parameter | Runoff per Control, Treatment, and Water Source | | | | |
|--|---|----------------|---------------|---------------|---------------|
| | Rock | Grass | Conc | Liner | Pool |
| Conductivity ($\mu\text{S}/\text{cm}$) | 95.5 – 2273.8 | 951.8 - 1354.2 | 266.5 – 916.8 | 225.7 – 246.1 | 216.0 -239.3 |
| pH [units] | 3.47 – 5.95 | 6.51 – 6.84 | 8.44 -9.16 | 7.58 – 7.96 | 7.74 – 7.89 |
| ANC (meq/L) | -0.25 – 0.03 | 1.72 – 3.03 | 1.28 – 1.97 | 1.28 – 1.40 | 1.21 – 1.43 |
| Cl (mg/L) | 19.2 – 32.9 | 108.6 - 250.13 | 16.5 – 22.2 | 15.9 - 18.3 | 13.1 – 18.8 |
| SO ₄ ²⁻ (mg/L) | 203.9 – 903.1 | 144.2 – 295.1 | 13.1 – 179.3 | 12.5 – 25.6 | 10.8 - 13.8 |
| NO ₃ ⁻ (mg/L) | 1.5 – 11.7 | 12.2 – 34.7 | 0.04 – 0.97 | 0.03 – 2.33 | 0.02 – 1.17 |
| Sum AA (meq/L) | 3.96 – 15.16 | 5.93 – 10.11 | 0.70 – 3.54 | 0.69 – 0.94 | 0.57 – 0.76 |
| Na ⁺ (mg/L) | 8.6 – 11.7 | 6.6 – 10.1 | 9.8 – 51.1 | 8.2 – 17.2 | 0.7 – 8.3 |
| K ⁺ (mg/L) | 2.1 – 4.2 | 172.5 -215.9 | 7.9 – 202.0 | 1.8 - 35.7 | 0.2 – 1.9 |
| Mg ²⁺ (mg/L) | 13.5 – 33.7 | 12.5 – 22.9 | 4.38 – 7.29 | 5.32 - 6.77 | 0.59 -7.08 |
| Ca ²⁺ (mg/L) | 38.8 – 72.6 | 59.4 – 90.2 | 19.1 – 26.7 | 24.7 – 27.3 | 2.35 – 25.49 |
| Sum BC (meq/L) | 3.54 – 6.88 | 8.88 – 11.77 | 2.20 – 9.10 | 2.25 – 3.33 | 0.21 – 2.20 |
| Dis Al (mg/L) | 0.03 – 15.17 | 0.24 – 0.61 | 0.03 – 0.17 | 0.02 - 0.07 | 0.00 – 0.02 |
| Dis Cu (mg/L) | 0.01 – 0.63 | 0.02 - 0.06 | 0.01 – 0.02 | 0.01 - 0.02 | 0.00 – 0.01 |
| Dis Fe (mg/L) | 4.13 – 143.50 | 2.59 – 13.04 | 0.01 – 0.22 | 0.00 – 0.04 | 0.00 – 0.01 |
| Dis Mn (mg/L) | 5.08 – 25.16 | 0.44 – 4.53 | 0.01 – 0.19 | 0.00 - 0.03 | 0.00 – 0.00 |
| Dis Si (mg/L) | 1.97 – 3.15 | 3.31 -10.36 | 3.30 – 12.51 | 2.71 – 6.58 | 0.00 - 2.60 |
| Dis Zn (mg/L) | 0.46 – 6.15 | 0.07 – 0.25 | 0.01 – 0.07 | 0.02 – 0.08 | 0.01 - 0.06 |
| Dis Cd (mg/L) | 0.000 - 0.001 | 0.001 – 0.002 | 0.001 -0.001 | 0.001 -0.001 | 0.001 – 0.001 |
| Dis Ni (mg/L) | 0.33 – 3.07 | 0.04 – 0.13 | 0.00 – 0.00 | 0.00 – 0.00 | 0.00 – 0.00 |

Sum AA = sum of acid anions; Sum BC = sum of base cations; Dis = dissolved

3.3.2 Natural Rainfall Monitoring

In general, the runoff chemistry from the pyrite rock panel (control) and three treatments: soil/vegetation (turf grass), shotcrete, and geoliner did not significantly change over the six month (June – December 2016) monitoring periods (Figures 3.11-3.19). In addition, the runoff chemistry was variable over this period (Table 3.3). The runoff pH remained low between 3.00 and 3.74 during this six-month period indicating that freshly exposed pyrite rock will take a longer time period to reach a condition where the surface pyrite has completely oxidized. The rain water pH was above 6 for most storm events. The runoff pH from the soil/vegetation (turf grass) treatment generally ranged between 5 and 6, and it was above 6 for the shotcrete and geoliner treatments. Though runoff sulfate remained constant over the six-month period, it declined significantly for the soil/vegetation treatment. Related to the decline in sulfate, conductivity and sum of acid anions also declined. Sulfate was likely a fertilizer additive to the turf grass sod (soil/vegetation treatment), and was it was depleted after about three months. Sum BC for the turf declined over the monitoring period which was likely due to calcium depletion from lime additions to the turf grass soil. As observed with the simulated rainfall experiments, complex biogeochemical processes occurred and further study is needed to assess how soil/vegetation can be used as a treatment for road cut sites through pyritic rock formations.

Statistical analysis to determine whether there was a significance difference among the pyrite rock control and three treatments: soil/vegetation (turf grass), shotcrete, and geoliner used a

Table 3.3. Summary of runoff chemistry from natural rainfall from June to December 2016 for the exposed pyrite rock control (Rock), and treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner). Water chemistry of the rain water was also reported (Rain). Summary data reported as the average, standard deviation (StDev), and range (maximums and minimums).

| Chemical Parameter | Runoff per Control, Treatment, and Rain | | | | |
|--|---|------------------------------------|-----------------------------------|-----------------------------------|-----------------------------------|
| | Rock | Grass | Conc | Liner | Rain |
| Conductivity ($\mu\text{S}/\text{cm}$) | 891.4 (436.2) 163.2 – 1740.0 | 558.8 (571.9) 115.8 – 2360.0 | 122.1 (130.5) 7.5 – 614.0 | 14.6 (15.5) 0.8 – 55.3 | 18.9 (20.8) 1.2 – 75.5 |
| pH [units] | 3.29 (0.18) 3.00 – 3.74 | 5.58 (0.62) 4.71 – 6.58 | 6.95 (0.39) 5.95 – 7.74 | 6.29 (0.39) 5.24 – 7.00 | 6.34 (0.31) 5.89 – 6.96 |
| ANC (meq/L) | 1.43 (6.25) -18.81 – 12.92 | -0.43 (4.89) -17.53 – 9.64 | 1.03 (1.19) 0.07 – 5.61 | 0.22 (0.54) -0.01 – 2.53 | 0.43 (0.90) -0.01 – 3.49 |
| Cl (mg/L) | 1.98 (4.28) 0.10 – 20.68 | 45.88 (57.92) 2.53 – 219.19 | 6.26 (26.69) 0.03 – 131.39 | 1.55 (5.48) 0.06 – 27.00 | 1.22 (3.13) 0.03 – 12.15 |
| SO ₄ ²⁻ (me/L) | 1.95 (1.86) 0.19 – 6.86 | 69.11 (89.89) 7.53 – 420.42 | 2.49 (6.19) 0.05 – 30.81 | 1.06 (1.03) 0.04 – 4.17 | 7.64 (22.71) 0.21 – 88.78 |
| NO ₃ ⁻ (mg/L) | 99.20 (88.29) 35.61 – 351.79 | 395.76 (337.29) 48.97 – 1580.58 | 22.50 (27.76) 1.83 (138.27) | 1.50 (2.34) 0.14 – 9.64 | 4.92 (12.18) 0.23 – 47.60 |
| PO ₄ ³⁻ (mg/L) | 0.25 (0.32) 0.03 – 1.13 | 1.66 (4.25) 0.00 – 16.31 | 0.07 (0.21) 0.00 – 0.74 | 0.02 (0.03) 0.00 – 0.09 | 0.02 (0.02) 0.00 – 0.06 |
| Sum AA (meq/L) | 6.45 (5.50) 0.28 – 25.70 | 4.40 (4.98) 1.05 – 20.75 | 0.58 (1.39) 0.03 – 6.58 | 0.10 (0.20) 0.01 – 0.91 | 0.27 (0.76) 0.01 – 2.96 |
| Na ⁺ (mg/L) | 1.93 (1.49) 0.02 – 3.84 | 2.26 (4.51) 0.27 – 21.02 | 3.00 (3.38) 0.00 – 15.72 | 0.41 (0.54) 0.04 – 2.65 | 0.55 (0.80) 0.00 – 3.20 |
| K ⁺ (mg/L) | 0.54 (0.85) 0.04 – 3.97 | 13.15 (11.37) 9.83 – 32.94 | 11.23 (9.90) 0.11 – 44.89 | 0.51 (1.09) 0.01 – 5.03 | 0.88 (1.87) 0.03 – 5.51 |
| Mg ²⁺ (mg/L) | 17.74 (13.58) 2.77 – 51.33 | 14.64 (13.27) 3.99 (61.20) | 4.09 (7.85) 0.04 – 31.39 | 0.24 (0.36) 0.05 – 1.59 | 0.71 (1.45) 0.01 – 5.58 |
| Ca ²⁺ (mg/L) | 19.18 (23.55) 3.92 – 101.32 | 37.38 (35.79) 10.84 – 148.27 | 13.39 (21.74) 0.88 – 95.90 | 5.08 (13.30) 0.46 – 62.11 | 11.28 (28.30) 0.10 – 110.63 |
| Sum BC (meq/L) | 2.51 (1.71) 0.43 – 5.84 | 3.51 (2.84) 1.36 – 13.36 | 1.42 (1.86) 0.05 – 8.10 | 0.30 (0.73) 0.05 – 3.40 | 0.67 (1.61) 0.01 – 3.49 |
| Dis Al (mg/L) | 18.18 (15.60) 0.00 – 59.21 | 0.63 (0.38) 0.07 – 1.58 | 0.03 (0.02) 0.01 – 0.05 | 0.01 (0.01) 0.00 – 0.06 | 0.04 (0.05) 0.00 – 0.13 |
| Dis Cu (mg/L) | 0.58 (0.62) 0.07 – 2.56 | 0.02 (0.02) 0.01 – 0.11 | 0.01 (0.01) 0.00 – 0.04 | 0.01 (0.01) 0.00 – 0.02 | 0.01 (0.02) 0.00 – 0.06 |
| Dis Fe (mg/L) | 76.38 (70.12) 10.41 – 284.27 | 2.64 (1.99) 0.49 – 8.71 | 0.29 (0.31) 0.02 – 1.04 | 0.06 (0.10) 0.00 – 0.40 | 0.22 (0.34) 0.00 – 0.99 |
| Dis Mn (mg/L) | 12.25 (11.25) 1.23 – 45.40 | 2.86 (2.36) 1.16 – 10.78 | 1.28 (4.43) 0.01 – 20.49 | 0.01 (0.02) 0.00 – 0.06 | 0.02 (0.03) 0.00 – 0.08 |
| Dis Si (mg/L) | 0.84 (0.65) 0.20 – 3.08 | 2.17 (1.14) 0.86 – 5.20 | 1.87 (1.38) 0.59 – 7.05 | 0.06 (0.09) 0.00 – 0.41 | 0.23 (0.64) 0.00 – 2.51 |
| Dis Zn (mg/L) | 1.55 (1.61) 0.19 – 6.30 | 0.16 (0.15) 0.05 – 0.62 | 0.03 (0.06) 0.00 – 0.24 | 0.02 (0.04) 0.00 – 0.13 | 0.01 (0.01) 0.00 – 0.04 |
| Dis Cd (mg/L) | 0.00 (0.00) 0.00 – 0.00 | 0.00 (0.00) 0.00 – 0.00 | 0.00 (0.00) 0.00 – 0.00 | 0.00 (0.00) 0.00 – 0.00 | 0.00 (0.00) 0.00 – 0.00 |
| Dis Ni (mg/L) | 1.04 (0.94) 0.14 – 3.85 | 0.06 (0.06) 0.02 – 0.26 | 0.02 (0.06) 0.00 – 0.27 | 0.00 (0.00) 0.00 – 0.00 | 0.00 (0.00) 0.00 – 0.01 |

Sum AA = sum of acid anions; Sum BC = sum of base cations; Dis = dissolved

single-factor ANOVA for the tested parameters, pH, conductivity, ANC, sum of acid anions, sum of base cations, sulfate, dissolved iron, and dissolved aluminum. All ANOVA tests were significantly different ($p < 0.001$).

The chemical parameters pH, conductivity, ANC, sum of acid anions, sum of base cations, sulfate, dissolved iron, and dissolved aluminum were significantly different between the pyrite rock control and the three treatments using a statistical t-test with no exceptions (Table 3.4). Conductivity and sum of base cations between pyrite rock control and soil/vegetation were not significantly different. Both the control and the treatment had high concentrations of calcium for the control and soil/vegetation treatment compared with the shotcrete and geoliner treatments and rain water (Table 3.3). Runoff from the geoliner treatment and rain water were essentially the same water therefore no significant difference occurred for most chemical parameters, except dissolved metals were found to be different. These two dissolved metals were likely found to be significantly different due to the low concentrations of the metals and low variance. There also was a lack of significant differences for sulfate and sum of acid anions for treatments pairs: shotcrete and rain water, and shotcrete and geoliner. Overall, the results indicate that shotcrete and geoliner are effective treatment measures to prevent runoff acidification, and the soil/vegetation (turf grass sod) needs further study.

Table 3.4. Statistical results for runoff treatment/control/rain pairs using a t-test for chemical parameters from the natural rainfall events from June – December 2016. Control = pyrite rock (Rock), Treatments = soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain).

| Control/Treatment/ Rain Pairs | | Chemical Parameter: Statistical Test Significance Level | | | | | | |
|----------------------------------|-------|---|---------|---------|---------|---------|---------|---------|
| | | pH | Cond | Sum AA | Sum BC | Sulfate | Dis Fe | Dis Al |
| Rock | Grass | < 0.001 | 0.016 | 0.003 | 0.206 | < 0.001 | < 0.001 | < 0.001 |
| Rock | Conc | < 0.001 | < 0.001 | < 0.001 | 0.004 | 0.008 | < 0.001 | < 0.001 |
| Rock | Liner | < 0.001 | < 0.001 | < 0.001 | < 0.001 | 0.025 | < 0.001 | < 0.001 |
| Rock | Rain | < 0.001 | < 0.001 | < 0.001 | < 0.001 | 0.011 | < 0.001 | < 0.001 |
| Grass | Conc | < 0.001 | 0.001 | < 0.001 | < 0.001 | < 0.001 | < 0.001 | < 0.001 |
| Grass | Liner | < 0.001 | < 0.001 | < 0.001 | < 0.001 | < 0.001 | < 0.001 | < 0.001 |
| Grass | Rain | < 0.001 | < 0.001 | < 0.001 | < 0.001 | < 0.001 | < 0.001 | < 0.001 |
| Conc | Liner | < 0.001 | < 0.001 | 0.063 | 0.001 | 0.285 | 0.002 | 0.006 |
| Conc | Rain | < 0.001 | < 0.001 | 0.197 | 0.003 | 0.159 | 0.286 | 0.162 |
| Liner | Rain | 0.396 | 0.256 | 0.195 | 0.174 | 0.299 | 0.047 | 0.034 |

Cond = Conductivity; Sum AA = sum of acid anions; Sum BC = sum of base cations; Dis Fe = dissolved iron; Dis Al = dissolved aluminum

3.3.3 Relationships between Runoff Chemistry and Natural Rainfall Volumes

Because of the importance of understanding the potential acidification of runoff from exposed pyrite rock two additional analyses were conducted with water chemistry from the control panel. The two analyses included 1) examining the influence of rainfall magnitude on runoff chemistry for pH, conductivity, sulfate and iron; and 2) the influence of dry days between storm events on runoff chemistry. A summary of rainfall depth and runoff volumes for the control panel and treatments is in Table 3.5. As the magnitude of the storm event increases, there were general trends of increased pH, decreased conductivity and dissolved iron, and

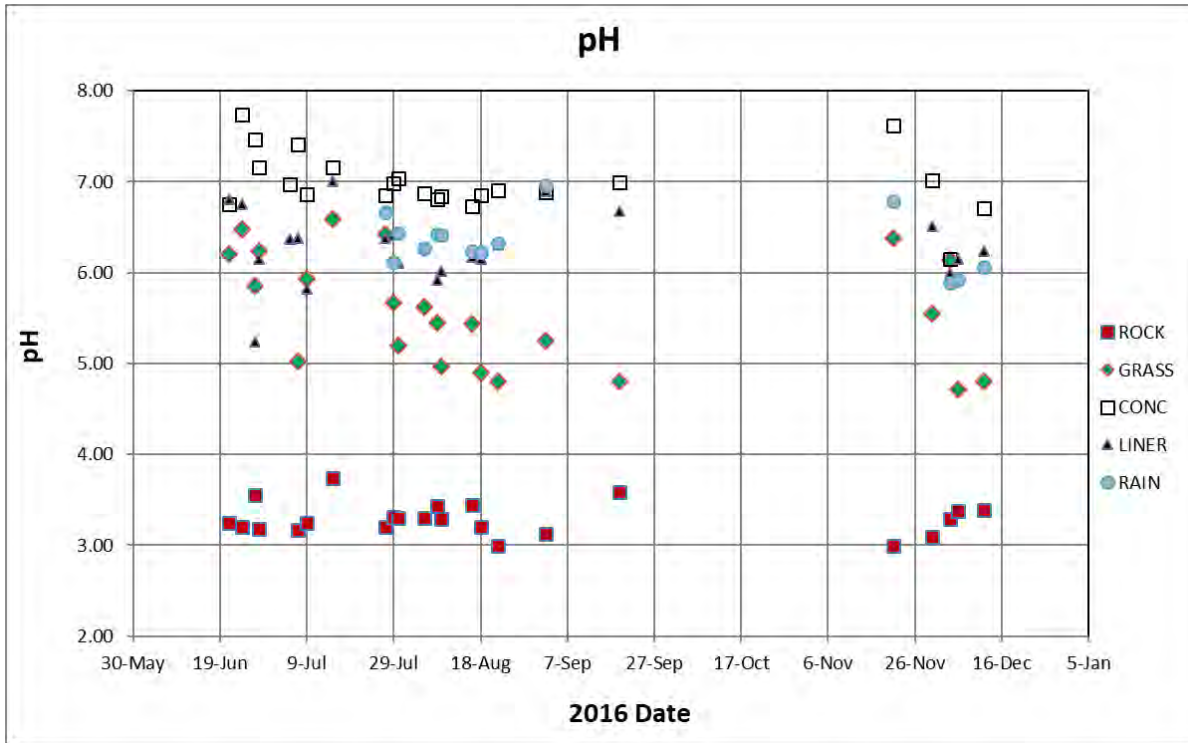


Figure 3.11. Runoff pH from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain).

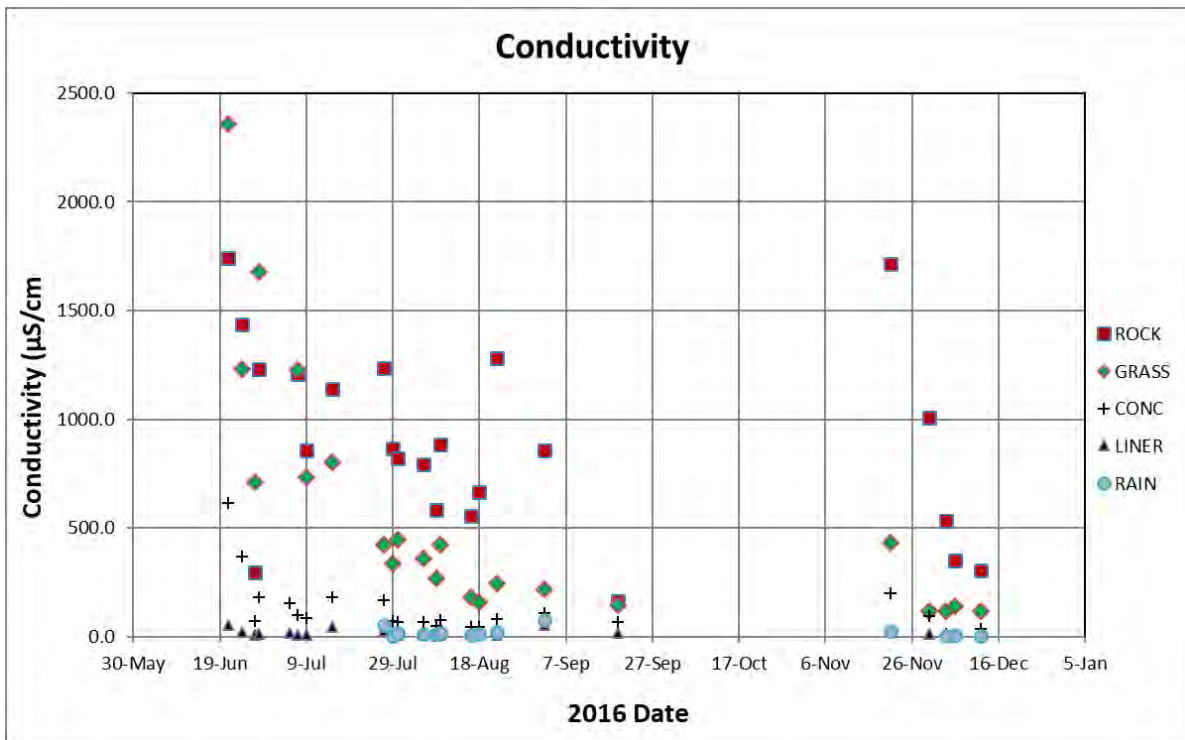


Figure 3.12. Runoff conductivity (µS/cm) from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain).

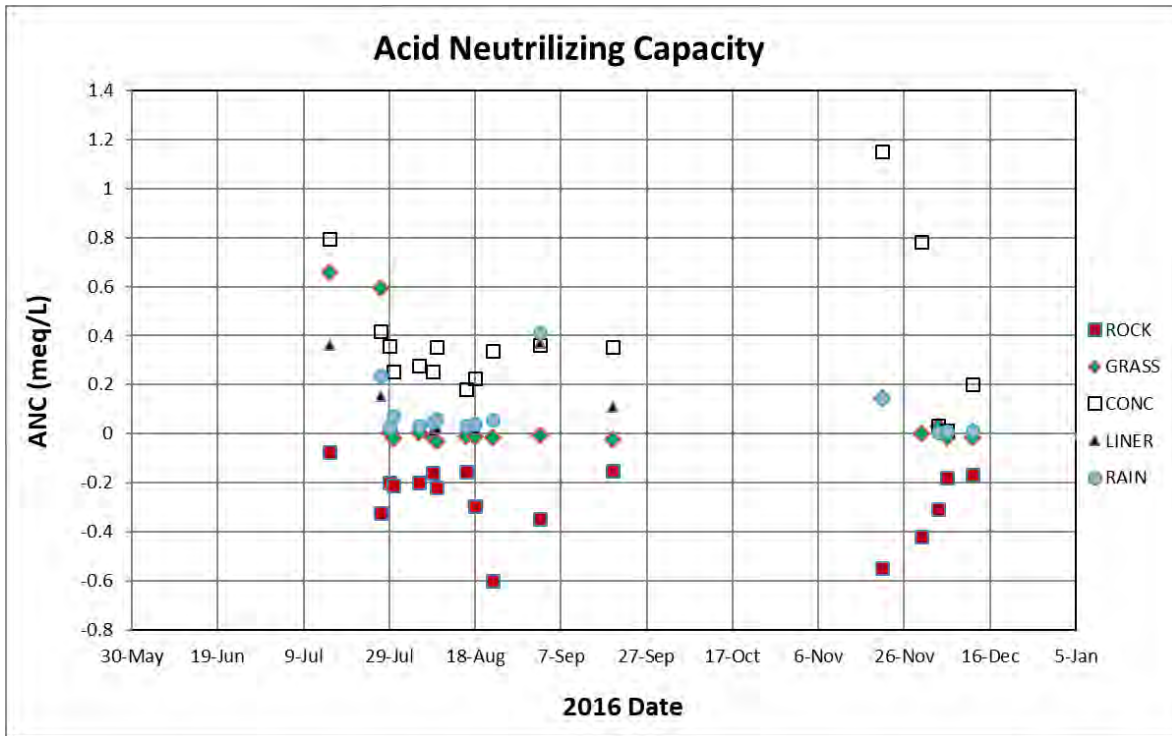


Figure 3.13. Runoff ANC (meq/L) from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain).

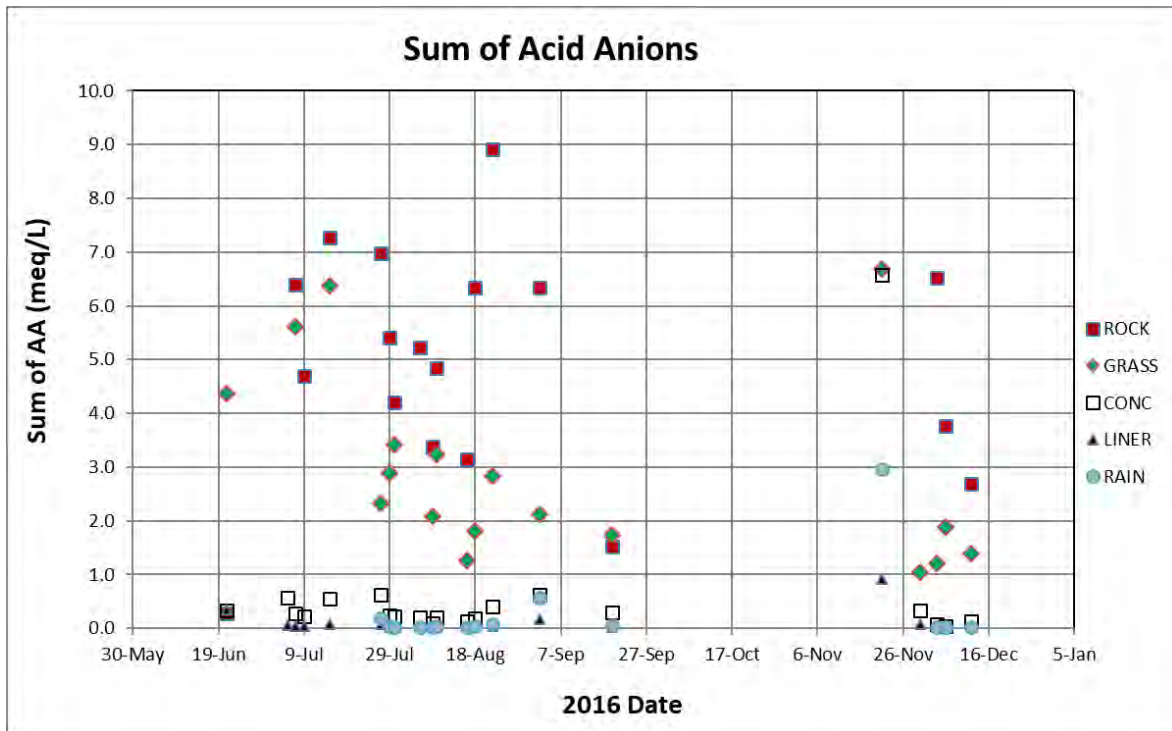


Figure 3.14. Runoff sum of acid anions (meq/L) from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain).

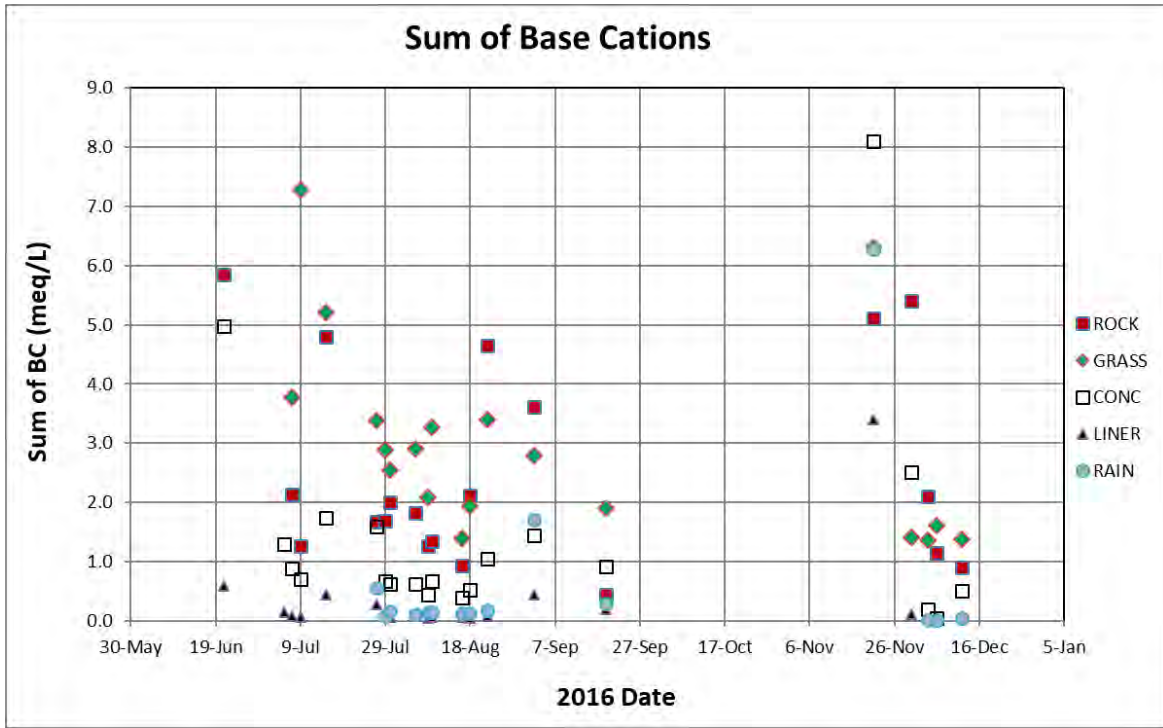


Figure 3.15. Runoff sum of base cations (meq/L) from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain).

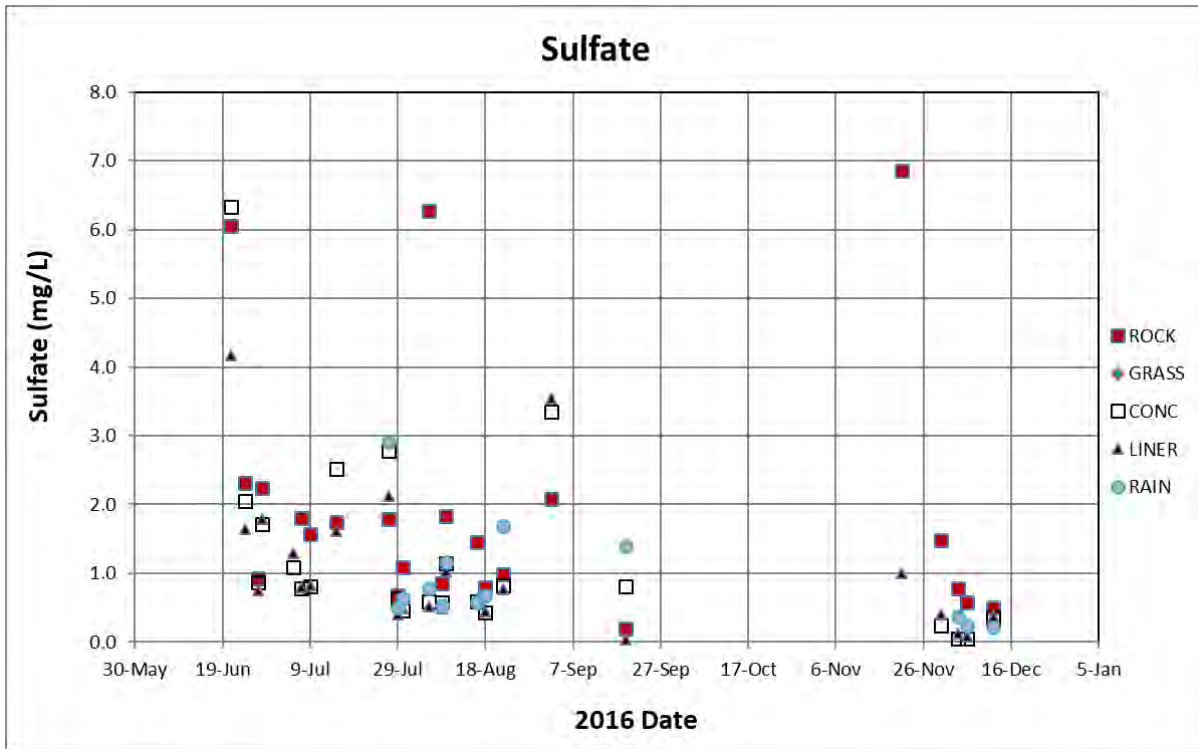


Figure 3.16. Runoff sulfate (mg/L) from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: shotcrete (Conc), and geoliner (Liner), and rain water (Rain).

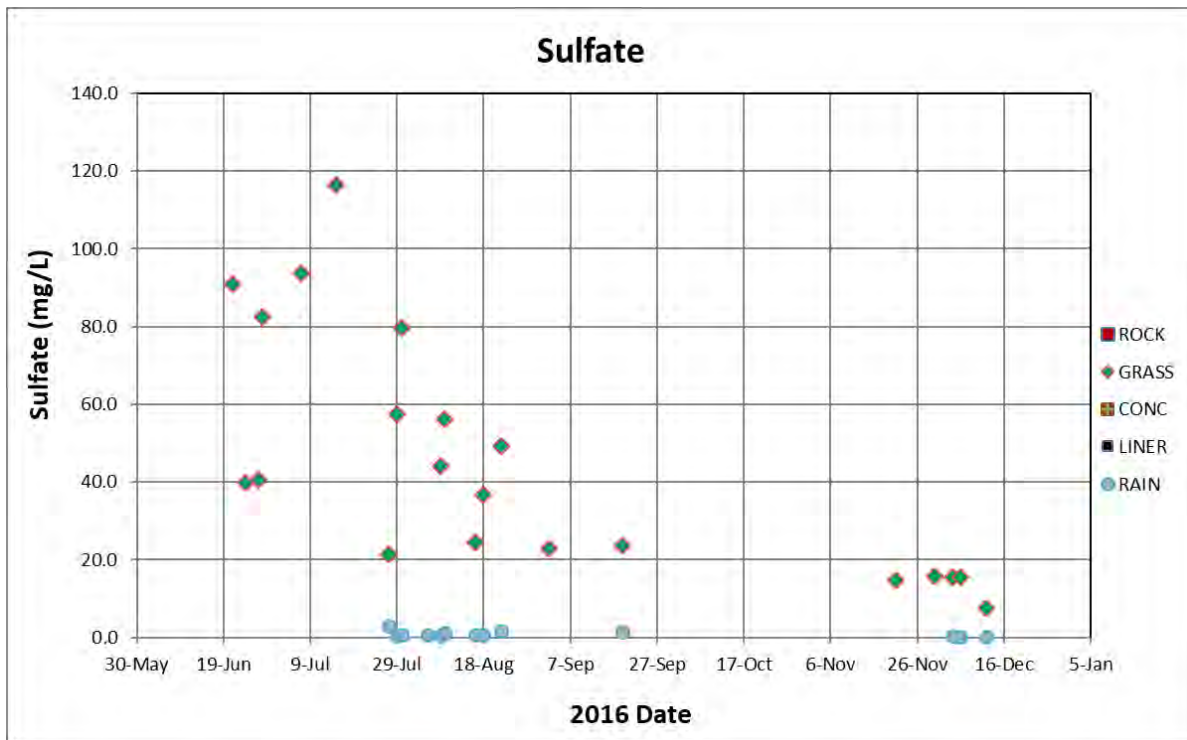


Figure 3.17. Runoff sulfate (mg/L) from natural rainfall (June – December 2016) for the treatment soil/vegetation (Grass), and rain water (Rain).

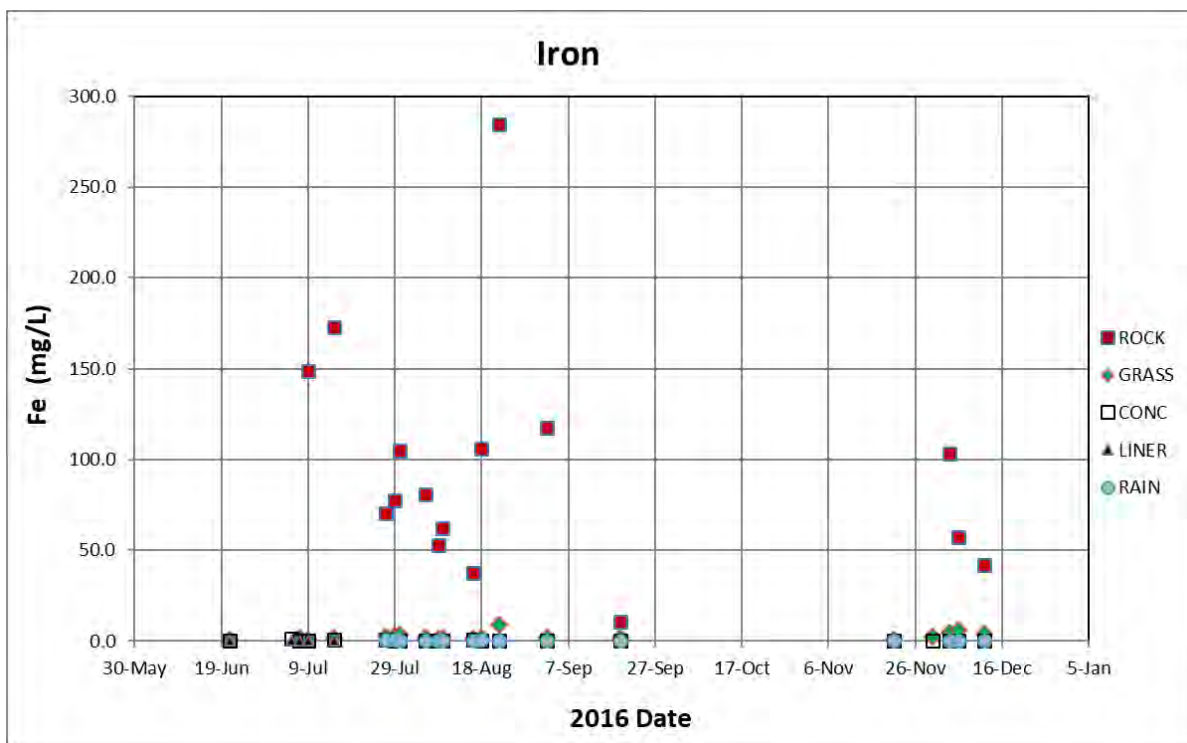


Figure 3.18. Runoff dissolved iron (mg/L) from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain).

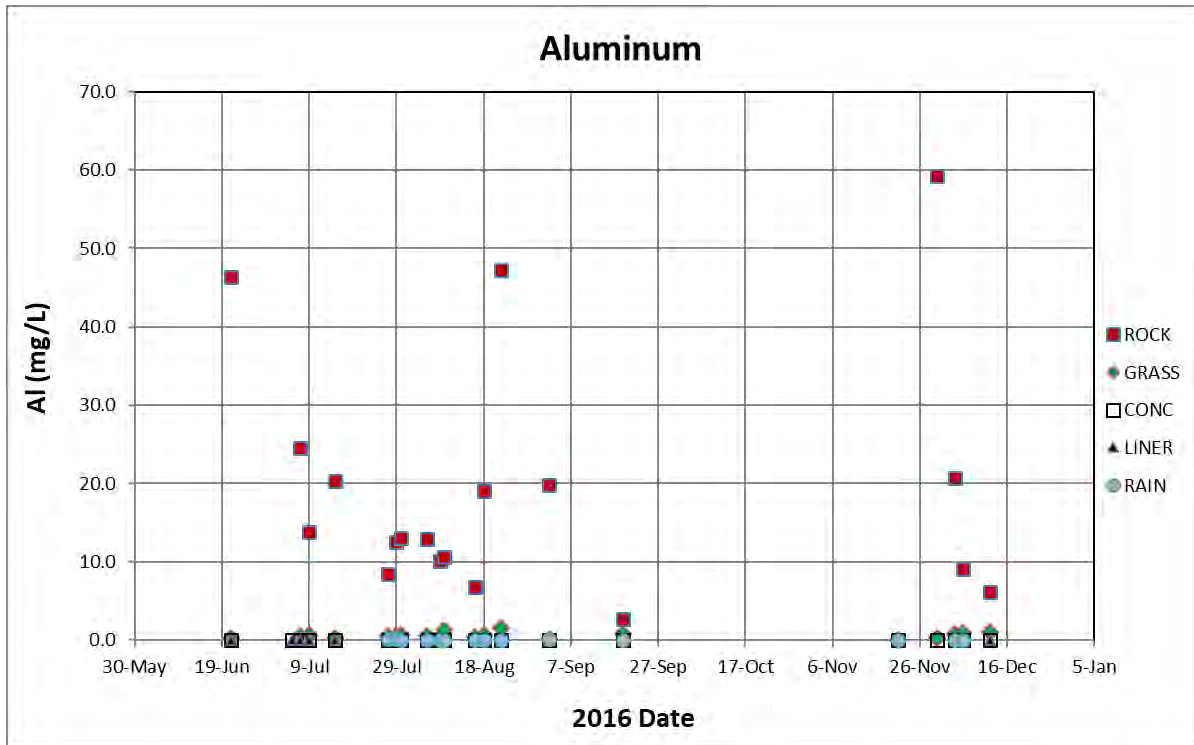


Figure 3.19. Runoff dissolved aluminum (mg/L) from natural rainfall (June – December 2016) for the exposed pyrite rock control (Rock), treatments: soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner), and rain water (Rain).

variable trend for sulfate (Figure 3.20). No correlations were found between number of dry days between storm events and runoff chemistry. In general, there appears to be a limit to the amount of ion mass loading that can be washed off for a storm event; therefore a dilution effect occurs for larger storm events.

3.4 Discussion and Conclusion

In this study, runoff from simulated and natural rainfall onto exposed pyrite rock and three material “treatment” covers were collected and chemically analyzed to assess their performance to reduce to eliminate acidification. The exposed pyrite rock was used as the experimental control, but also provided valuable information as to how long would exposed “freshly cut” pyrite rock would generate acidity and changes based on different magnitude storm events. The “freshly cut” pyrite rock represents a road cut with newly exposed surfaces. In this experiment, Anakeesta rock from a landslide on Clingman’s Dome Road in the Great Smoky Mountains National Park was collected the day of the landslide on May 28, 2015. The rock was stored in a cold, dry indoor location at the University of Tennessee until the experimental set-up could be constructed, and the experiments implemented. The material covers selected for this study included soil/vegetation, shotcrete, and a geosynthetic membrane liner, which are three of the feasible covers that could be incorporated into roadway design if necessary. An experiment using constructed panel containers testing these covers for effects on runoff water quality had not been conducted before by others so it offers new information to better understand how runoff from road cut sites through pyritic rock could be managed for protection of receiving surface waters.

Table 3.5. A summary of storm events for the natural rainfall monitoring between June – December 2016 consisting of storm event depth (inch) and estimated runoff volumes for the pyrite rock control panel (Rock) and three treatments soil/vegetation (Grass), shotcrete (Conc), and geoliner (Liner).

| Storm Event | | Rainfall Depth (inch) | Runoff Volumes (ft ³) for Control/Treatments | | | |
|-------------|-----------|--------------------------|--|-------|-------|-------|
| No. | 2016 Date | | Rock | Grass | Conc | Liner |
| 1 | June 21 | 0.05 | 0.031 | 0.015 | 0.022 | 0.048 |
| 2 | June 24 | 0.32 | 0.292 | 0.171 | 0.142 | 0.242 |
| 3 | June 27 | 1.54 | 1.470 | 0.876 | 0.684 | 1.118 |
| 4 | June 28 | 0.20 | 0.176 | 0.101 | 0.088 | 0.155 |
| 5 | July 5 | 0.54 | 0.504 | 0.298 | 0.240 | 0.400 |
| 6 | July 7 | 1.25 | 1.190 | 0.709 | 0.555 | 0.910 |
| 7 | July 9 | 0.79 | 0.746 | 0.443 | 0.351 | 0.579 |
| 8 | July 15 | 0.24 | 0.215 | 0.125 | 0.106 | 0.184 |
| 9 | July 27 | 0.27 | 0.244 | 0.142 | 0.120 | 0.206 |
| 10 | July 29 | 0.73 | 0.688 | 0.408 | 0.324 | 0.536 |
| 11 | July 30 | 0.34 | 0.311 | 0.182 | 0.151 | 0.256 |
| 12 | August 5 | 1.16 | 1.103 | 0.657 | 0.515 | 0.845 |
| 13 | August 7 | 0.07 | 0.050 | 0.026 | 0.031 | 0.062 |
| 14 | August 8 | 1.77 | 1.692 | 1.009 | 0.787 | 1.284 |
| 15 | August 9 | 0.50 | 0.466 | 0.275 | 0.222 | 0.371 |
| 16 | August 16 | 1.93 | 1.847 | 1.102 | 0.858 | 1.399 |
| 17 | August 18 | 1.14 | 1.084 | 0.645 | 0.507 | 0.831 |
| 18 | August 22 | 0.38 | 0.350 | 0.205 | 0.169 | 0.285 |
| 19 | Sept. 2 | 0.15 | 0.128 | 0.072 | 0.066 | 0.119 |
| 20 | Sept. 19 | 0.88 | 0.833 | 0.495 | 0.391 | 0.644 |
| 21 | Sept. 27 | 0.16 | 0.137 | 0.078 | 0.071 | 0.127 |
| 22 | Nov. 21 | 0.18 | 0.157 | 0.090 | 0.080 | 0.141 |
| 23 | Nov. 30 | 4.77 | 4.589 | 2.745 | 2.121 | 3.440 |
| 24 | Dec. 4 | 1.12 | 1.064 | 0.633 | 0.498 | 0.817 |
| 25 | Dec. 6 | 1.63 | 1.557 | 0.928 | 0.725 | 1.183 |
| 26 | Dec. 12 | 1.04 | 0.987 | 0.587 | 0.462 | 0.759 |

The simulated rainfall experiment applied an intensity of 0.21 inch/hour equivalent to a 1-yr, 24-hr return frequency (for the Knoxville area) in order to observe rapid changes in runoff chemistry within a few hours of a storm event start. Runoff acidification occurred from the exposed pyrite rock panel with the initial pH measured as about 3.6, and increased to about 6.0 within four hours (Figure 3.3). Because the pyrite rock lacked base cations (Figure 3.5), ANC remained about 0 meq/L throughout the simulated rainfall period (Figure 3.4). After one day rest, reapplication of simulated rainfall found that once exposed to air rapid oxidation occurred with the initial pH measured at about 3.6 as the day before. However, the day 2 simulation observed the pH over time to increase to 6.0 more quickly within two hours. Sulfate export from the rock surface was the major driver for runoff acidification, where the pattern of initially high concentration followed a declining concentration over the simulated time mirroring the inverse of the pH pattern (Figure 3.7). Conductivity in this experiment was likely controlled by sulfate

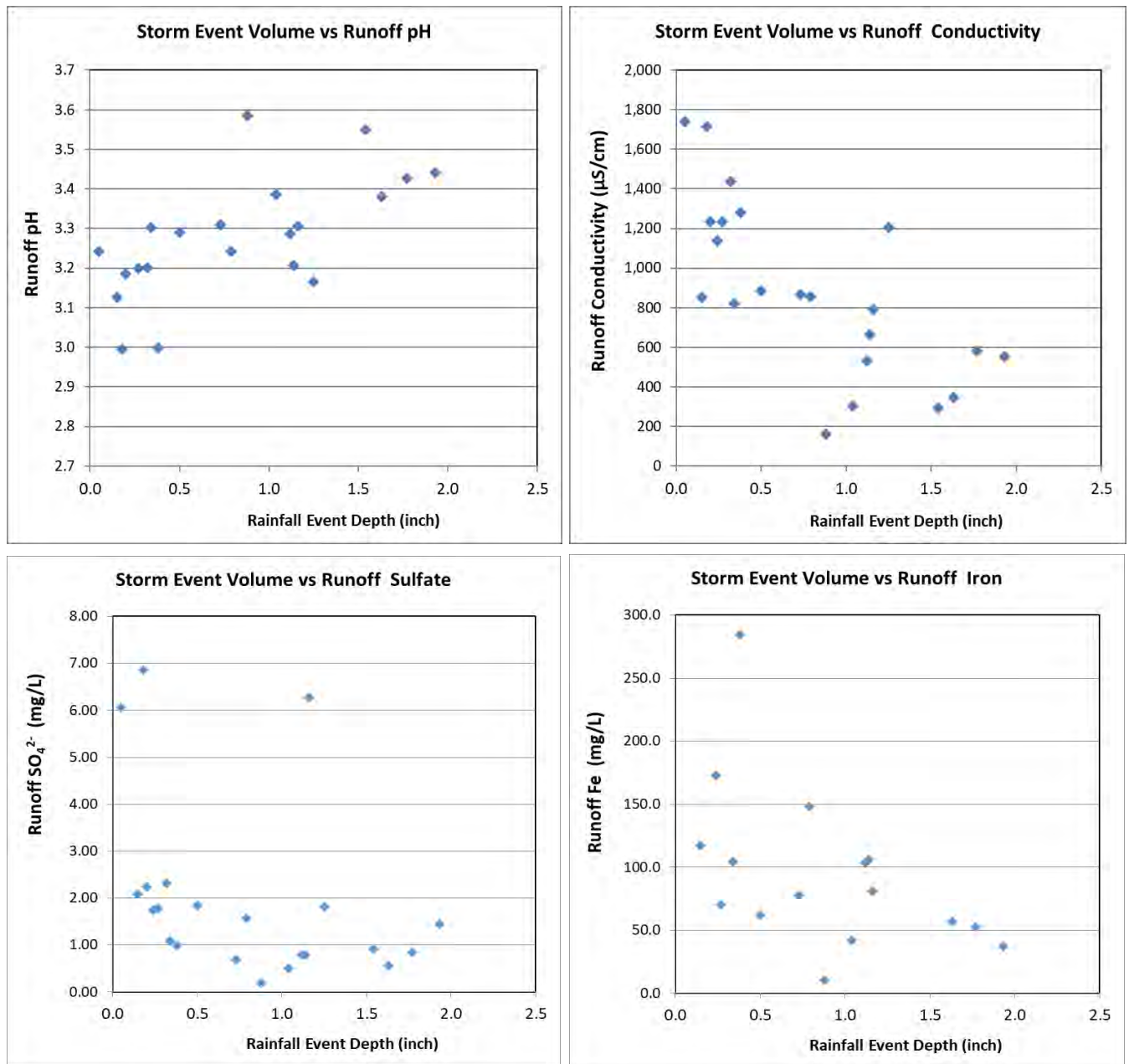


Figure 3.20. Plots of storm event magnitude (depth, inch) from the pyrite rock control panel and chemistry for pH, conductivity ($\mu\text{S}/\text{cm}$), sulfate (mg/L), and dissolved iron (mg/L).

concentrations since the same pattern of decline over the simulated rainfall period (Figure 3.8). Dissolved metals declined rapidly with the first two hours from the start of the simulated rainfall experiment (Figures 3.9 and 3.10, Table 3.2). It appears there is a “first flush” and dilution effect that occurs on the pyrite rock surface. Chemical oxidation and weathering at the surface provides for a metal ion source but that source is rapidly depleted thereby diluting the runoff.

Results of the simulated rainfall experiments demonstrated the available ions that can cause runoff acidification occurs over a short period of time, and as rainfall continues during a storm event a dilution effect occurs from the depletion of available sulfate ions. In addition, pyrite rock appears to have limited base cations to neutralize the initial acidification effect at the start of a storm event. The experiment also demonstrated that the pyrite rapidly oxidizes under exposed

air and makes available more sulfate within a day, however the availability of ions may be more limited. The statistical analysis that examined whether there was an effect on the number of days on runoff chemistry found no effect. This finding suggests pyrite oxidation and availability of sulfate for runoff acidification is surface area dependent. Once the initial rock surface has been washed of ions, there must be a period in which it is exposed to air to regenerate the ion source for the next storm. The experiment with natural rainfall monitoring provides further information on a longer term perspective to availability of sulfate and other ions. From an environmental impact perspective primarily aluminum due to its toxicity on aquatic biota, the fact that oxidation and weathering limits metal ion source and dilution occurs relatively quickly is beneficial. Within one hour, the aluminum concentration fell below the toxic levels reported at 2.0 mg/L (Figure 3.10). In addition the pH increased above 6 within four hours.

The material “treatment” covers using shotcrete and geoliner were effective preventing acidification of the runoff from the experimental panel containers (Figures 3.3-3.10). The results from the soil/vegetation cover using turf grass sod were mixed but generally demonstrated that that cover could be effective in reducing runoff acidification. However, results were mixed likely due to applied fertilizers on the sod consisting of nitrogen, phosphorous, potassium, and sulfur. In addition, the sod was likely limed with pulverized dolomite (calcium and magnesium) to control soil pH. With these additives, several biogeochemical processes co-occur including mineralization and nitrification, sulfate adsorption, cation/metal exchange, vegetation uptake of nitrate, and base cation weathering (Nodvin et al. 1986; Lindberg and Lovett 1992; Manning et al. 1996; Blake et al. 1999; Dubikova et al. 2002; Tipping et al. 2003; Cai et al. 2011a). Therefore, these results from this treatment needs to be carefully assessed and further experimentation is needed with a study design that controls for different chemical soil additives.

The monitoring of the natural rainfall for the additional six months after the simulated rainfall experiments found that the pyrite rock (experimental control) panel acidification of runoff continued for the long-term. The runoff pH remained between 3 and 4 (Figure 3.11). ANC also remained low during this monitoring period reported as less than 0 meq/L (Figure 3.13). As observed with the simulated rainfall experiment, runoff acidification was dominated by acid anions primarily sulfate (Figures 3.14-3.16). Elevated levels of dissolved iron and aluminum were observed for the runoff from the exposed rock caused by pyrite oxidation and Anakeesta shale weathering. The runoff chemistry from the exposed pyrite rock was statistically different from rainwater chemistry indicating the rock affected the runoff water quality, apparently for a period at least six months in duration (Table 3.4). A dilution effect was also observed where pH increased with storm event size (rainfall depth), and conductivity, sulfate, and dissolved iron decreased (Figure 3.20). With the information from the simulated rainfall experiments it appears that there is a limited availability of sulfate to cause runoff acidification. This dilution effect was also observed with coal-refuse deposits and exposed shales (Olyphant et al. 1991; Tuttle et al. 2009). It should be noted that that these experiments were for a condition with only pyritic rock. In contrast to the Tennessee characterization study (Chapter 2), it was observed that if other geologic formations and site soils are adjacent to the exposed pyritic formation they contribute to base cations to runoff reducing acidification. However, if a material cover is needed it appears that shotcrete and a geoliner would be effective to exposing pyritic rock to rainfall and generating acidified runoff. The use of a soil/vegetation cover needs further study, but appears it could be effective if soil fertilizers were adequately controlled and managed for the long-term. Overall, this study provides alternatives for TDOT if control of acidified runoff is needed. The potential for treatment of acidified runoff in the drainage below the road cut is reviewed in the next chapter (Chapter 4).

4. Passive Water Quality Treatment Options at Road Cuts through APM

4.1 Introduction

Guidelines on addressing the environmental impacts from acid material drainage (AMD) at road cut construction sites were developed and published by TDOT with Golder Associates (TDOT 2007). The guidance document “*Guidelines for Acid Producing Rock Investigation, Testing, Monitoring and Mitigation*” was comprehensive and focused on testing, handling, and disposal during highway construction. The TDOT (2007) document also summarized water treatment options for post-construction mitigation of the site runoff. In general, treatment of AMD has been extensively studied primarily addressing the impacts of surface mining in the Appalachian region and other regions of the world with similar geologic formations with pyrite. A list of relevant cited literature is summarized in Section 4.3 below. If runoff quality exceeds levels that may be harmful to aquatic biota (Appendix E) or impacts a public water supply, there may be a need for treatment of AMD from road cuts.

The objective of this task consists of a summary literature review of different runoff treatment options from road cut sites through pyritic rock formations, and an introductory assessment of what treatment options would perform sufficiently with the new knowledge gained from this study’s road-cut characterization of runoff chemistry and performance cover treatments (Chapters 2 and 3). The literature review focused on previous work on AMD from highway projects, which are more relevant to the challenges faced by TDOT based on typical topography and right of way constraints in regions with APM. The potential treatment options relevant to these highway construction constraints are summarized in Section 4.2 below.

4.2 Potential Treatment Options

Water treatment of AMD has been developed for both active and passive systems (TDOT 2007). Both systems utilize aggregate carbonate to neutralize the pH and encourage metals precipitation as hydroxides or sulfide minerals (Taylor et al. 2005). Active systems consisted of chemical treatment requiring mechanical infrastructure and their use generally is for highly populated acidic waters with elevated levels of dissolved metals and sulfate. Passive treatment systems are best suited for AMD with low acidity (< 800 mg CaCO₃ /L) and low flow rates (< 50 L/s). Therefore, mass loadings from road cut sites should be less than about 100 to 150 kg CaCO₃ /day. Relevant to highway runoff from road cuts through pyritic rock are passive systems. In the TDOT (2007) documents several passive treatments were identified as preferred treatment methodologies because of ease of construction, low maintenance, and treatment longevity. They were:

- Limestone beds and oxic/open channels
- Aerobic wetlands
- Settling ponds
- Bio-reactors
- Pebble quicklime

Other identified passive treatments included: vertical flow ponds, oxic limestone drains, anaerobic wetlands, and chemical addition of bases, i.e., caustic soda, soda ash, ammonia, and hydrated lime. Others have reported the addition of organic matter to provide alkalinity, enhance metals adsorption, and create redox reducing conditions which favor precipitation of sulfide minerals (Taylor et al. 2005). In order to design these passive treatment systems, a public

domain software package is available termed AMD Treat©. Version 5.0.2 Plus can be downloaded from the following web site: <https://amd.osmre.gov/downloads.htm>.

In order to design any treatment system the acidity loading must be computed (Taylor et al. 2005). Acidity is the measure of both hydrogen ion concentration and mineral acidity. Mineral or latent acidity includes the potential concentration of hydrogen ions produced by precipitation of metal hydroxides in the water, which reactions are pH dependent. Acidity load refers to the total acidity and flow rate, which forms the measure of mass per time, or more specifically mass CaCO₃ equivalent per unit time (or given volume of pass-through water). The following equations can be used to estimate acidity and acidity loads:

$$\text{Acidity (mg/L CaCO}_3\text{)} = 50 \times \{3x[\text{Fe}^{2+} + \text{Fe}^{3+}]/56 + 3x[\text{Al}^{3+}]/27 + 2x[\text{Mn}^{2+}]/55 + 1000x10^{-(\text{pH})}\}$$

$$\text{Acidity Loads (tonnes CaCO}_3\text{ /day)} = 10^{-9} \times 86,400 \text{ (conversion factor)} \times \text{flow rate (L/s)} \\ \times \text{Acidity (mg/L CaCO}_3\text{)}$$

Oxic/open Limestone Channels/Drains. Oxic/open limestone channels/drains (OLC) armor channels lined with coarse limestone and can be also designed with a series of check dams (Ziemkiewicz and Brant 1996; Cravotta and Trahan 1999; Taylor et al. 2005). Figure 4.1 shows an example of an OLC. OLC is a recommended passive treatment option if the ARD has a pH less than 5.0 and contains total iron less than 20 ppm. The coarse limestone allows for an increase in alkalinity of the water and can remove iron and aluminum via oxide and hydroxide precipitation. Calcium carbonate dissolution occurs over time, but its rate is pH dependent and a function of the precipitates that can coat or armour the limestone rock surface. The design goal is to maintain an adequate flow rate to flush precipitates through the porous spaces with continuous exposure of AMD influent with a minimum contact time of 1 to 3 hours. Channels should be constructed on the steepest possible gradient to maximize water velocity and thereby minimize armoring and plugging. However, slopes should not exceed 10° to maintain coarse aggregate stability. OLC is typically preceded by a retention pond capable of storing a 10-yr 24-hr precipitation event. The average acidity load entering an OLC should be less than 150 kg CaCO₃ /day, which can achieve a pH between 6 and 8. According to Skousen (2002), OLC can remove approximately 70% of the dissolved iron, 40-50% of the dissolved aluminum, and 10-20% of the dissolved manganese. The design life has been reported to be between 10 and 20 years (Ziemkiewicz et al. 2002). An example in Tennessee where an OLC has been used was at the Jamestown Study Sites 1 and 3 in Fentress County (Table 2.1). These treatment systems are likely the most applicable to highway construction and road cut locations because of the linear nature of the design and that of the roadway. Space permitting, a small settling pond improves overall performance by providing a location for metal precipitates to settle and makes maintenance less troublesome.

Aerobic Wetlands. Aerobic wetlands are shallow ponds that remain aerated with contact with air at the water surface (Figure 4.1). Wetlands can also be thought of as vegetated settling/retention ponds that allow for sediment and metal-oxidized precipitates to settle. Aerobic wetlands are a recommended passive treatment option if the ARD has a pH greater than 5.0 and contains total iron greater than 20 ppm. Aerobic wetlands are best suited for the treatment of water with relatively low acidity but elevated ferrous iron concentration, which could be treated by allowing the ferrous iron to oxidize in the aerobic zones of the wetlands and precipitate as iron oxides. These treatment systems are complex with a multitude of biogeochemical and plant-mediated processes occurring that naturally improve water quality (Hedin et al. 1994; Kilborn



Figure 4.1. Example photos of an open limestone channel (top) and a vertical flow wetlands (bottom). Photos from Paul Ziemkiewicz, West Virginia University, 2002.

1999; Milavec 2002; Ziemkiewicz et al. 2002; Taylor et al. 2005). These processes include: sedimentation, filtration, oxidation, reduction, precipitation, adsorption, complexation, chelation, microbial conversion/ immobilization, and plant uptake of nutrients and metals. Discharge to the aerobic wetland typically passes through a settling pond before entering the wetland. Hydraulic considerations are needed for designing the pond; typically pond sizing is based on maximum influent flows equating to a 10-yr 24-hr precipitation event. The purpose of the settling pond is to allow settling of iron precipitates and sediments before the flow enters the wetland. Vertical flow wetland design has been described by Kepler and McCleary (1997) to improve treatment performance. The average acidity load entering a wetland should be less than $1 \text{ kg CaCO}_3 / 300 \text{ m}^2$ water surface area, which can achieve a pH above 6 and dissolved iron removal between 60-95% (Taylor et al. 2005). Physical space for wetlands can be problematic for some highway projects.

Bio-reactors. Bio-reactors are anaerobic pass-through structures comprising of a substrate of organic matter and limestone aggregate. Bio-reactors are recommended passive treatment options for ARD with a total iron concentration of greater than 20 ppm. These treatment systems are typically vertical flow consisting of three layers from top to bottom: 1) ponded water; 2) limestone-buffered organic substrate; and 3) limestone. Iron, aluminum, and other heavy metals are removed via sulfide and hydroxide precipitation in addition to sorption to hydroxide surfaces

and the organic substrate. Bio-reactors rely on sulfate-reducing bacterial activity in which organic matter (carbon) is used to converted sulfate into hydrogen sulfide which is a gas and needs to be ventilated off site. Metal concentrations can be decreased by precipitation of metal sulfides in the reduced anaerobic organic matter layer of the bio-reactor (Eger and Melchert 1992). With the addition of limestone layer, greater alkalinity can be generated to assist in increasing water pH. The life expectancy of these bio-reactors is controlled by the available organic matter. Thus bio-reactors may require more maintenance than the OLC and wetlands described above. In addition, routine nutrient supplements may be needed to achieve adequate microbial growth. The organic matter / limestone aggregate mix can become clogged, requiring maintenance by replacing the bio-reactor media. The average acidity load entering a bio-reactor should be less than 1 kg CaCO₃ /200-500 m²·day, which can achieve a pH between 6 and 8 (Kilborn 1999). Dissolved oxygen concentrations should remain below 1 mg/L. The location of the bio-reactors should be well ventilated because of odor issues associated with hydrogen sulfide gas. Flow rates through the reactors should be low, approximately less than 1 L/s (Taylor et al. 2005). An additional advantage of bio-reactor is that the space needed is less than that for wetland treatment systems.

4.3 Cited Literature for Acid Rock Drainage Treatment

The following citations were found that provide for a bibliography for AMD treatment. A few of these references may overlap with the overall reference section. The literature and design criteria for treating AMD are readily available but minimally applied to road cut sites through pyritic rock formations. Relevant references are as follow:

- Akcil, A., and S. Koldas. 2006. Acid mine drainage (AMD): Causes, treatment and case studies. *Journal of Cleaner Production* 14 (12–13): 1139–1145.
- Cendrero, A., R. Anton, and J.S. De Omenaca. 1977. Geochemistry of bedrock; Its effect on the planting and maintenance of roadcuts along the Bilbao—Behobia Motorway (Northern Spain). *Landscape Planning* 4: 173–183.
- Costa, M.C., and J.C. Duarte. 2005. Bioremediation of acid mine drainage using acidic soil and organic wastes for promoting sulphate-reducing bacteria activity on a column reactor. *Water, Air, and Soil Pollution* 165 (1–4): 325–345.
- Cruz Vigg, C., F. Pagnanelli, A. Cibati, D. Uccelletti, C. Palleschi, and L. Toro. 2010. Biotreatment and bioassessment of heavy metal removal by sulphate reducing bacteria in fixed bed reactors. *Water Research* 44 (1): 151–158.
- De Vegt, A.L., Exton, P. A., Bayer, H. G. and Buisman, C. J. (1998). Biological sulphate removal and metal recovery from mine waters. *Mining Engineering*, Vol. 50 No. 11, p 67-70, 1998.
- Demchal, J., Morrow, T. and Skousen, J. (1996). *Treatment of acid mine drainage by four vertical flow wetlands in Pennsylvania*. Morgantown, WV 26506, USA Clarion University, Clarion, PA, USA. <http://www.wvu.edu/~agexten/landrec/treatment.htm>.
- Eger, P. and Melchert, G. (1992). The design of a wetland treatment system to remove trace metals from mine drainage. p 98-107. In: *Proceedings, 1992 American Society for Surface Mining and Reclamation Conference*, May 14-17, 1992, Duluth, MN. <http://www.wvu.edu/~agexten/landrec/passtr/passtr.htm>
- Enpar Technologies (2002). *Innovative Amdel Electrochemical Cover Technology Proceeds to Field Testing at the Golden Sunlight Mine, Montana, USA*. Enpar Technologies, Inc., News Release, 24 Oct/7 Nov 2002.

- EPA. 1994. *Acid Mine Drainage Prediction; Technical Document*. (EPA-530-R-94-036; available NTIS PB94-201829). Washington, D.C.: U.S. Environmental Protection Agency, Office of Solid Waste.
- Gazea, B., K. Adam, and A. Kontopoulos. 1996. A review of passive systems for the treatment of acid mine drainage. *Minerals Engineering* 9 (1): 23–42.
- Gusek, J. J. 2002. *Sulfate-Reducing Bioreactor Design and Operating Issues: Is This the Passive Treatment Technology for Your Mine Drainage?*, presented at the National Association of Abandoned Mine Land Programs, Park City, Utah, September 15-18, 2002.
- Gray, N.F. 1997. Environmental Impact and Remediation of Acid Mine Drainage: a management problem. *Environmental Geology* 30 (1–2): 62–71.
- Hallberg, K.B., and D.B. Johnson. 2004. Biological manganese removal from acid mine drainage in constructed wetlands and prototype bioreactors. *Science of The Total Environment* 338 (1–2): 115–124.
- Hard, B.C., J.P. Higgins, and A. Mattes. 2003. Bioremediation of acid rock drainage using sulphate-reducing bacteria. *Proceedings of Sudbury: Mining and Environment, Sudbury, Ontario, May, 2003*, 25–28.
- Hardeman, W. D., 1966. *State Geologic Map*, Scale 1:250,000 (1 inch = 4 miles), in 4 sheets. Tennessee Department of Environment and Conservation, Division of Geology, Maps and Publications Sales Office, 401 Church Street, 13th Floor, L&C Tower, Nashville, Tennessee 37243-0445.
- Johnson, D.B., M.A. Dziurla, A. Kolmert, and K.B. Hallberg. 2002. The microbiology of acid mine drainage: Genesis and biotreatment: review article. *South African Journal of Science* 98 (5-6): p.249.
- Johnson, D. B., and K.B. Hallberg. 2005. Acid mine drainage remediation options: a review. *Science of the Total Environment* 338 (1): 3–14.
- Johnson, D. B., and K.B. Hallberg. 2005. Biogeochemistry of the compost bioreactor components of a composite acid mine drainage passive remediation system. *Science of The Total Environment* 338 (1–2): 81–93.
- Johnson, D.B, O. Rowe, S. Kimura, and K.B. Hallberg. 2004. Development of an integrated microbiological approach for remediation of acid mine drainage and recovery of heavy metals. In *Proceedings of IMWA Symposium Mine Water*, 151–57, 2004.
- Kilborn (1999). *Review of Passive Systems for Treatment of Acid Mine Drainage*. Mine Environment Neutral Drainage (MEND) Report 3.14.1.
- Li, L. Y., M. Chen, J. R. Grace, K. Tazaki, K. Shiraki, R. Asada, and H. Watanabe. 2006. Remediation of acid rock drainage by regenerable natural clinoptilolite. *Water, Air, and Soil Pollution* 180 (1–4): 11–27.
- Luptakova, A., and M. Kusnierova. 2005. Bioremediation of acid mine drainage contaminated by SRB. *Hydrometallurgy* 77 (1–2): 97–102.
- McCauley, Craig A., Aisling D. O’Sullivan, Mark W. Milke, Paul A. Weber, and Dave A. Trumm. 2009. Sulfate and metal removal in bioreactors treating acid mine drainage dominated with iron and aluminum. *Water Research* 43 (4): 961–970.
- Milavec, P. J. (1999). *1998 Status Report, ARD Set Aside Program*. Bureau of Abandoned Mine Reclamation, Pennsylvania, USA.
- Milavec, P. J. (2002). *Abandoned mine drainage abatement projects’ success, problems and lessons learned*. The Pennsylvania Department of Environmental Protection, Bureau of Abandoned Mine Reclamation, USA. http://www.dep.state.pa.us/dep/deputate/minres/bamr/ amd/ amd_abatement_projects.htm.

- Miller, S., Smart, R., Andrina, J., Neale, A., and Richards, D. (2003). Evaluation of limestone covers and blends for long-term acid rock drainage control at the Grasberg Mine, Papua Province, Indonesia. p. 133-141. In: *Proceedings Sixth International Conference on Acid Rock Drainage*. (Cairns, QLD, July 12-18, 2003). The Australian Institute of Mining and Metallurgy, Melbourne.
- Moore, H. 1992. *The Use of Geomembranes for Mitigation of Pyritic Rock*, in 43rd Annual Highway Geology Symposium, Fayetteville, AK, August 1992. Asheville, NC: The Symposium.
- Morgan, E. L., W. F. Porak, and J. A. Arway. 1983. Controlling acidic-toxic metal leaches from southern Appalachian construction slopes: Mitigating stream damage. *Transportation Research Record*, No. 948.
- Mukhopadhyay, B., L. Bastias, and A. Mukhopadhyay. 2007. Limestone drain design parameters for acid rock drainage mitigation. *Mine Water and the Environment* 26 (1): 29–45.
- Mylona, E., A. Xenidis, and I. Paspaliaris. 2000. Inhibition of acid generation from sulphidic wastes by the addition of small amounts of limestone. *Minerals Engineering* 13 (10–11): 1161–1175.
- Nordstrom D.K and C.N. Alpers. 1998. Geochemistry of Acid Mine Waters. In *The Environmental Geochemistry of Mineral Deposits, Part A: Processes, Techniques, and Health Issues*. G.S. Plumlee and M.J. Logsdon, editors. Reviews in Economic Geology Volume 6A, Society of Economic Geologists, Inc.
- Nyquist, J., and M. Greger. 2009. A Field Study of Constructed Wetlands for Preventing and Treating Acid Mine Drainage. *Ecological Engineering* 35: 630–642.
- Peppas, A., K. Komnitsas, and I. Halikia. 2000. Use of organic covers for acid mine drainage control. *Minerals Engineering* 13 (5): 563–574.
- Ray, D., M. Clark, and T. Pitman. 2009. Treatment of an iron-rich ARD using waste carbonate rock: bench-scale reactor test results. *Mine Water and the Environment* 28 (4): 253–263.
- Sand, W., P.-Georg Jozsa, Z.M. Kovacs, N. Sásáran, and A. Schippers. 2007. Long-term evaluation of acid rock drainage mitigation measures in large lysimeters.” *Journal of Geochemical Exploration* 92 (2–3): 205–211.
- Shelp, M. L., Hayward, G. L., Seed, L. P. and Shelp. G. S. (2005). *Electrochemical Cover for the Prevention of Acid Mine Drainage – A Laboratory Test*. Golder Associates Ltd., Mississauga, Ontario. School of Engineering, The University of Guelph, Ontario. Enpar Technologies Inc., Guelph Ontario.
- Sheoran, A.S., and V. Sheoran. 2006. Heavy metal removal mechanism of acid mine drainage in wetlands: A critical review. *Minerals Engineering* 19 (2): 105–116.
- Sheoran, A.S., V. Sheoran, and R.P. Choudhary. 2010. Bioremediation of acid-rock drainage by sulphate-reducing prokaryotes: A review. *Minerals Engineering* 23 (14): 1073–1100.
- Sibilski, U. (2001). AngloGold Desalination Pilot Plant Project. In: *Proceedings of the Conference on Environmentally Responsible Mining*. Chamber of Mines of South Africa, 25-28 September, 2001.
- Shokes, T.E., and G. Möller. 1999. Removal of dissolved heavy metals from acid rock drainage using iron metal. *Environmental Science & Technology* 33 (2): 282–287.
- Sicree, A.A. 2006. Regional mineralogy of sulfide deposits north and south of the I-99 roadcut at Skytop, Centre County, Pennsylvania. In *Geological Society of America Abstracts with Programs*, Vol. 38 (2), p. 33. Geological Society of America.
- Skousen, J. 1997. Overview of passive systems for treating acid mine drainage. *Green Lands* 27 (4): 34–43.

- Skousen, J. (2002). <http://www.wvu.edu/~agexten/landrec/passtrt/passtrt.htm>. West Virginia University, September.
- Skousen J.G., J.C. Sencindiver, and R.M. Smith. 1987. *A Review of Procedures for Surface Mining and Reclamation in Areas with Acid-Producing Materials*, in cooperation with the West Virginia Surface Mine Drainage Task Force, the West Virginia University Energy and Water Research Center, and the West Virginia Mining and Reclamation Association. Morgantown, WV: The Center.
- Smoke, J.D., R.D. Neufeld, J. Monnell, and T. Gray. 2008. Remediation of acid rock discharges. *Proceedings of the Water Environment Federation* No. 11: 4790–4802.
- Smyth, D. J. A., Blowes, D. W., Ptacek, C. J. and Bain, J. G. (2003). *Removal of Dissolved Metals from Groundwater Using Permeable Reactive Barriers – Applications*. Paper presented at the 6th ICARD, July 12-18, 2003, Cairns, Australia.
- Stenzel, G. and Günther, P. (2005). *Mine Water Treatment at Landau Colliery*. Slide presentation accessed on 2/2/06 from www.sacollierymanagers.org.za.
- Taylor, J., S. Pape, and N. Murphy. 2005. A Summary of Passive and Active Treatment Technologies for Acid and metalliferous Drainage (AMD). Fifth Australian Workshop on Acid Drainage; Fremantle, Western Australia, 29-31 August 2005.
- Taylor, J., Guthrie, B., Murphy, N. and Waters, J. (2006). *Alkalinity Producing Cover Materials for Providing Sustained Improvement in Water Quality from Waste Rock Piles*. Paper submitted for the 7th ICARD, March 26-30, 2006, St Louis MO.
- Thomas, R.C., and C.S. Romanek. 2002. Passive treatment of low-pH, ferric iron-dominated acid rock drainage in a vertical flow wetland I: Acidity neutralization and alkalinity generation. In *Proc. of the 2002 National Meeting of the American Society of Mining and Reclamation, Lexington, KY*, 9–13, 2002.
<http://asmr.us/Publications/Conference%20Proceedings/2002/0723%20Thomas.pdf>.
- Tris 1980. Research Project 73-9 Final Report-Executive Summary: Application of Limestone and Lime Dust in the Abatement of Acidic Drainage in Centre County, Pennsylvania. <http://trid.trb.org/view.aspx?id=164348>.
- Valente, T, and C Lealgomes. 2009. Occurrence, properties and pollution potential of environmental minerals in acid mine drainage. *Science of The Total Environment* 407 (3): 1135–1152.
- Waters, J. C., Santomartino, S., Murphy, N., Taylor, J. R. (2003). *Acid mine drainage treatment technologies: Identifying sustainable solutions*. Paper presented at the 6th ICARD, July 12-18, 2003, Cairns, Australia.
- Ziemkiewicz, P. F. and Brant, D. L. (1996). The Casselman River Restoration Project. In: *Proceedings Eighteenth West Virginia Surface Mine Drainage Task Force Symposium 1996*, Morgantown, West Virginia.
- Ziemkiewicz, P. F., Skousen, J. G. and Simmons, J. S. (2002). Long-term performance of passive acid mine drainage treatment systems. In: *Conference Proceedings of the Twenty Third West Virginia Surface Mine Drainage Task Force Symposium*. Ramada Inn, Morgantown, West Virginia, April 16-17, 2002.
- Ziemkiewicz, P.F., J.G. Skousen, and J. Simmons. 2003. Long-term performance of passive acid mine drainage treatment systems. *Mine Water and the Environment* 22 (3): 118–129.

References: Chapter 2, 3, and 4

- Adams, C.B., C.A. Klamke, and C.L. Hollabaugh. 1999. Geochemical monitoring of Kiser Creek, near Buchanan, Haralson County, Georgia: the effects of pyrite-rich rocks on the pH, iron, and sulfate content of surface waters. *Ga. J. Sci.* 57(2):113-122.
- Ammons, J.T., C.B. Coburn Jr, and P.A. Shelton. 1990. An Application of Acid-Base Accounting for Highway Construction in East Tennessee.” In *Mining and Reclamation Conference*, 265–69. Charleston, West Virginia: West Virginia Mine Drainage Task Force, 1990. <http://wvmdtaskforce.com/proceedings/90/90AMM/90AMM.HTM>.
- Bacon, J.R. and R.P. Maas. 1979. Contamination of Great Smoky Mountains trout streams by exposed Anakeesta formation. *Journal of Environmental Quality* 8(4): 538-543.
- Baldigo, B.P., G.B. Lawrence, et al. 2009. Impacts of acidification on macroinvertebrate communities in streams of western Adirondack Mountains, New York, USA. *Ecological Indicators* 9: 226-239.
- Baker, B.J., and J.F. Banfield. Microbial communities in acid mine drainage. *FEMS Microbiology Ecology* 44, no. 2 (2003): 139–52.
- Barbour M.T., J. Gerritsen, B.D. Synder, and J.B. Stribling. 1999. *Rapid Bioassessment Protocols for Use in Wadable Streams and Rivers: Periphyton, Benthic Macroinvertebrates, and Fish*. EPA 841-B-99-002. US Environmental Protection Agency: Washington, DC.
- Barker, M., H. Van Miegroet, N.S. Nicholas, and I.F. Creed. 2002. Variation in overstory nitrogen uptake in a small, high-elevation southern Appalachian spruce-fir watershed. *Canadian Journal of Forest Research* 32: 1741–1752.
- Berube, M., J. Locat, P. Gelinat, J. Chagnon, and P. LeFrancois. 1986. Black shale heaving at Sainte-Foy, Quebec, Canada. *Can. J. Earth Sci.* 23:1774-1781.
- Bigham, J. M., and D. K. Nordstrom. 2000. Iron and Aluminum Hydroxysulfates from Acid Sulfate Waters.” *Reviews in Mineralogy and Geochemistry* 40 (1): 351–403.
- Bird, D.A. 2003. Characterization of Anthropogenic and Natural Sources of Acid Rock Drainage at the Cinnamon Gulch Abandoned Mine Land Inventory Site, Summit County, Colorado.” *Environmental Geology* 44 (8): 919–32.
- Blake, L., K.W.T. Goulding, C.J.B. Mott, and A.E. Johnston. 1999. Changes in soil chemistry accompanying acidification over more than 100 years under woodland and grass at Rothamsted Experimental Station, UK. *European Journal of Soil Biology* 50(3): 401–412.
- Booth, C.E., D.G. McDonald, B.P. Simons, and C.M. Wood. 1988. Effects of Aluminum and Low pH on Net Ion Fluxes and Ion Balance in the Brook Trout (*Salvelinus-Fontinalis*). *Can. J. Fish. Aquat. Sci.* 45: 1563-1574.
- Bradley, M.W., and S.C. Worland. 2015. *Bibliography for Acid-rock Drainage and Selected Acid-mine Drainage Issues Related to Acid-rock Drainage from Transportation Activities*. U.S. Geological Survey Open-File Report 2015-1016. 17 p.
- Byerly, D.W. 1990. *Guidelines for Handling Excavated Acid-Producing Materials*. FHWA-FL-90-007: U.S. Dept. of Transportation, Federal Highway Administration, Washington, D.C.
- Byerly, D. W. 1996. Handling acid-producing material during construction. *Env. and Eng.Geosci.* II(1):49-57.
- Cai, M., A.M. Johnson, J.S. Schwartz, S.E. Moore, and M.A. Kulp. 2012. Soil acid-base chemistry of a high-elevation forest watershed in the Great Smoky Mountains National Park watershed: Influences of acid deposition. *Water, Air, and Soil Pollution* 223: 289-303.
- Cai, M., A.M. Johnson, J.S. Schwartz, S.E. Moore, and M.A. Kulp. 2011a. Response of soil water chemistry to simulated changes in acid deposition in the Great Smoky Mountains. *ASCE Journal of Environmental Engineering* 137(7): 617-628.

- Cai, M., J.S. Schwartz, R.B. Robinson, S.E. Moore, and M.A. Kulp. 2011b. Long-term annual and season patterns of acidic deposition and stream water quality in a Great Smoky Mountains high-elevation watershed. *Water, Air and Soil Pollution* 219: 547-562.
- Cai, M. J.S. Schwartz, R.B. Robinson, S.E. Moore, and M.A. Kulp. 2010. Long-term effects of acidic deposition on water quality in a high-elevation Great Smoky Mountains National Park watershed: use of an ion input-output budget. *Water, Air, and Soil Pollution* 209: 143-156.
- Cape, J.N., L.J. Sheppard, D. Fowler, A.F. Harrison, J.A. Parkinson, P. Dao, et al. 1992. Contribution of canopy leaching to sulfate deposition in a Scots pine forest. *Environmental Pollution* 75: 229–236.
- Carline, R.F., D.R. DeWalle, W.E. Sharpe, B.A. Dempsey, C.J. Gagen and B. Swistock. 1992. Water chemistry and fish community responses to episodic stream acidification in Pennsylvania, USA. *Environmental Pollution* 78: 45-48.
- Cleveland, L., E.E. Little, C.G. Ingersoll, R.H. Wiedmeyer and J.B. Hunn. 1991. Sensitivity of brook trout to low pH, low calcium and elevated aluminum concentrations during laboratory pulse exposures. *Aquatic Toxicology* 19: 303-318.
- Cronan, C.S., and C.L. Schofield. 1990. Relationships between aqueous aluminum and acid deposition in forested watersheds of North America and Northern Europe. *Environmental Science and Technology* 24(7): 1100–1105.
- Cruz Viggli, C., F. Pagnanelli, A. Cibati, D. Uccelletti, C. Palleschi, and L. Toro. 2010. Biotreatment and bioassessment of heavy metal removal by sulphate reducing bacteria in fixed bed reactors. *Water Research* 44 (1): 151–158.
- Dahlgren, R.A., and W.J. Walker. 1993. Aluminum release rates from selected Spodosol Bs horizons: Effects of pH and solid phase aluminum pools. *Geochim. Cosmochim. Acta* 57(1): 57–66.
- DeLonay, A.J., E.E. Little, D.F. Woodward, W.G. Brumbaugh, A.M. Farag and C.F. Rabeni 1993. Sensitivity of early-life-stage golden trout to low pH and elevated aluminum. *Environmental Toxicology and Chemistry* 12: 1223-1232.
- DeNicola, D.M., and M.G. Stapleton. 2002. Impact of acid mine drainage on benthic communities in streams: The relative roles of substratum vs. aqueous effects. *Environmental Pollution* 119 (3): 303–15.
- Deyton, E.B., J.S. Schwartz, R.B. Robinson, K.J. Neff, S.E. Moore, and M.A. Kulp. 2009. Characterizing episodic stream acidity during stormflows in the Great Smoky Mountains National Park. *Water, Air, and Soil Pollution* 196: 3-18.
- Di Toro, D.M., H.E. Allen, H.L. Bergman, J.S. Meyer, P.R. Paquin, and R.C. Santore. 2001. Biotic ligand model of the acute toxicity of metals. 1. Technical basis. *Environ Toxicol Chem* 20(10): 2383–2396.
- Draaijers, G.P.J., J.W. Erisman, N.F.M. Van Leeuwen, T.E. Romer, B.H. Winkel, A.C. Veltkamp, et al. 1997. The impact of canopy exchange on differences observed between atmospheric deposition and throughfall fluxes. *Atmospheric Environment* 31(3): 387–398.
- Driscoll C.T., J.P. Baker, J.J. Bisogni, and C.L. Schofield. 1980. Effect of aluminum speciation on fish in dilute acidified waters. *Nature* 284:161–164.
- Driscoll, C.T. 1985. Aluminum in acidic surface waters: chemistry, transport, and effects. *Environ Health Perspectives* 63: 93–104.
- Driscoll C.T., and K.M. Postek. 1995. The chemistry of aluminum in surface waters. In: Sposito G (ed.). *The Environmental Chemistry of Aluminum*. Lewis, Chelsea, MI, pp 363–418.

- Driscoll C.T., G.B. Lawrence, A.J. Bulger, T.J. Butler, C.S. Cronan, C. Eagar, K.F. Lambert, G.E. Likens, J.L. Stoddard, and K.C. Weathers. 2001. Acidic deposition in the northeastern United States: Sources and inputs, ecosystem effects, and management strategies. *BioScience* 51(3):180–198.
- Dubiková, M., P. Cambier, V. Šucha, and M. Čaplovičová. 2002. Experimental soil acidification. *Appl. Geochem* 17(3): 245–257.
- EnviroSci Inquiry. 2011. Basic AMD Chemistry. Lehigh University, Lehigh Environmental Initiative. <http://www.ei.lehigh.edu/envirosci/enviroissue/amd/links/science2.html>
- Essington, M.E. 2004. *Soil and Water Chemistry: An integrative approach*. CRC, Boca Raton, Florida.
- Exley, C., J.S. Chappell, and J.D. Birchall. 1991. A mechanism for acute aluminium toxicity in fish. *Journal of Theoretical Biology* 151: 417-428.
- Fettig, D. 2018. *Handbook of Geosynthetics*. Geosynthetic Materials Association (GMA). https://carthagemills.com/pdf/handbook_of_geosynthetics.pdf.
- Fernandez, I.J., L.E. Rustad, S.A. Norton, J.S. Kahl, and B.J. Cosby. 2003. Experimental acidification causes soil base-cation depletion at the Bear Brook Watershed in Maine. *Soil Science Society of America Journal* 67(6): 1909–1919.
- Fowler, T.A, P.R Holmes, and F.K Crundwell. 2001. On the Kinetics and Mechanism of the Dissolution of Pyrite in the Presence of *Thiobacillus Ferrooxidans*. *Hydrometallurgy* 59 (2–3): 257–270. doi:10.1016/S0304-386X(00)00172-9.
- Fox, D., C. Robinson, and M. Zentilli. 1997. Pyrrhotite and associated sulphides and their relationship to acid rock drainage in the Halifax Formation, Meguma Group, Nova Scotia. *Atlantic Geology*. 33: 87-103.
- Fromm, P.O. 1980. A review of some physiological and toxicological responses of freshwater fish to acid stress. *Environmental Biology of Fishes* 5(1): 79-93.
- Gagen, C.J., W.E. Sharpe, et al. 1993. Mortality of brook trout, mottled sculpins, and slimy sculpins during acidic episodes. *Trans. Am. Fish. Soc.* 122: 616-628.
- Gazea, B., K. Adam, and A. Kontopoulos. 1996. A review of passive systems for the treatment of acid mine drainage. *Minerals Engineering* 9 (1): 23–42.
- Geosynthetica. 2018. <https://www.geosynthetica.net/geomembrane-provides-moisture-barrier-major-highway/>
- Gibert, O., J. de Pablo, J.L. Cortina, and C. Ayora. 2005. Municipal compost-based mixture for acid mine drainage bioremediation: Metal retention mechanisms. *Applied Geochemistry* 20 (9): 1648–1657.
- Gobran, G.R., H.M. Selim, H. Hultberg, and I. Andersson. 1998. Sulfate adsorption–desorption in a Swedish forest soil. *Water, Air, and Soil Pollution* 108: 411–424.
- Gonzalez, A.M. 2018. *Biogeochemistry of Sulfur in Small Forested Catchments in the Great Smoky Mountains National Park*. Doctoral Dissertation, University of Tennessee, Knoxville.
- Grande, J. A., R. Beltran, A. Sainz, J.C. Santos, M. L. Torre, and J. Borrego. 2004. Acid Mine Drainage and Acid Rock Drainage Processes in the Environment of Herrerias Mine (Iberian Pyrite Belt, Huelva-Spain) and Impact on the Andevalo Dam.” *Environmental Geology* 47 (2): 185–96. doi:10.1007/s00254-004-1142-9.
- Hallberg, K.B. 2009. New Perspectives in Acid Mine Drainage Microbiology. *Hydrometallurgy* 104 (3–4): 448–53.
- Hammarstrom, J.M., K. Brady, and C.A. Cravotta III. 2005. Acid-rock drainage at Skytop, Centre County, Pennsylvania, 2004. Open-File Report 2005-1148. U.S. Geological Survey, 954 National Center, Reston, VA 20192.

- Hammarstrom, J.M., R.R. Seal, A.L. Meier, and J.M. Kornfeld. 2005. Secondary sulfate minerals associated with acid drainage in the eastern US: recycling of metals and acidity in surficial environments. *Chem. Geol.* 215, 407-431.
- Harvey, H.H. and D.A. Jackson. 1995. Acid stress and extinction of a spring-spawning fish population. *Water, Air and Soil Pollution* 85: 383-388.
- Hayward Baker. 2018. <https://www.haywardbaker.com/uploads/solutions-techniques/sculpted-shotcrete/Hayward-Baker-Sculpted-Shotcrete-Brochure.pdf>.
- Hermann, J., E. Degerman, A. Gerhardt, C. Jonansson, P. Lingdell, and I. Muniz. 1993. Acid-stress effects on stream biology. *Ambio* 22(5): 298-307.
- Hitech Rockfall. 2018. <http://www.hitechrockfall.com/service-view/shotcrete/>
- Holzmueller, E.J., S. Jose, and M.A. Jenkins. 2007. Influence of calcium, potassium, and magnesium on *Cornus florida* L. density and resistance to dogwood anthracnose. *Plant and Soil* 290: 189–199.
- Hoomehr, S., J.S. Schwartz, D.C. Yoder, W. Wright, and E.C. Drumm. 2013. Curve numbers for hydrology on low compaction steep-sloped reclaimed mine lands in southern Appalachian. *ASCE Journal of Hydrologic Engineering* 18(12): 1627-1638.
- Houle, D., R. Ouimet, S. Couture, and C. Gagnon. 2006. Base cation reservoirs in soil control the buffering capacity of lakes in forested catchments. *Canadian Journal of Fisheries and Aquatic Sciences* 63: 471–474.
- Huckabee, J.W., C.P. Goodyear, and R.D. Jones. 1975. Acid rock in the Great Smokies: Unanticipated impact on aquatic biota of road construction in regions of sulfide mineralization. *Trans Am Fish Soc* 104, 677-684.
- Inamdar, S.P., S.F. Christopher, and M.J. Mitchell. 2004. Export mechanisms for dissolved organic carbon and nitrate during summer storm events in a glaciated forested catchment in New York, USA. *Hydrological Processes* 18: 2651–2661.
- Ji, S.W., Y.W. Cheong, G.J. Yim, and J. Bhattacharya. 2006. ARD generation and corrosion potential of exposed roadside rock mass at Boeun and Mujoo, South Korea. *Environmental Geology* 52 (6): 1033–1043.
- Johnson, D.B. and K.B. Hallberg. 2005. Acid mine drainage remediation options: A review. *Science of the Total Environment* 338 (1): 3–14.
- Johnson, D.W., and S.E. Lindberg. 1992. *Atmospheric deposition and nutrient cycling in forest ecosystems: A synthesis of the Integrated Forest Study*. Springer, New York.
- Klapper, H. and M. Schultze. 1995. Geogenically acidified mining lakes - living conditions and possibilities of restoration. *Int. Revue. Ges. Hydrobiol.* 80(4):639-653.
- Kock, D., and A. Schippers. 2008. Quantitative microbial community analysis of three different sulfidic mine tailing dumps generating acid mine drainage. *Applied and Environmental Microbiology* 74 (16): 5211–19. doi:10.1128/AEM.00649-08.
- Kucken, D.J., Davis, J.S., Petranka, J.W. and Smith, C.K. 1994. Anakeesta stream acidification and metal contamination: effects on a salamander community. *Journal of Environmental Quality* 23, 1311-1317.
- Lawrence, G.B. 2002. Persistent episodic acidification of streams linked to acid rain effects on soil. *Atmospheric Environment* 36: 1589–1598.
- Lindberg, S.E., and G.M. Lovett. 1992. Deposition and forest canopy interactions of airborne sulfur: results from the integrated forest study. *Atmospheric Environ.* 26A(8): 1477–1492.
- Igarishi, T. and T. Oyama. 1999. Deterioration of water quality in a reservoir receiving pyrite-bearing rock drainage and its geochemical modeling. *Eng. Geol.* 55:45-55.

- Ingersoll, C.G., D.D. Gully, D.R. Mount, M.E. Mueller, J.D. Fernandez, J.R. Hockett, and H.L. Bergman. 1990. Aluminum and acid toxicity to two strains of brook trout (*Salvelinus fontinalis*). *Can. J. Fish. Aquat. Sci.* 47: 164-912.
- MacAvoy, S.E., and A.J. Bulger. 2004. Sensitivity of blacknose dace (*Rhinichthys atratulus*) to moderate acidification events in Shenandoah National Park, U.S.A. *Water, Air and Soil Pollution* 153: 125-134.
- Mannings, S., S. Smith, and J.N.B Bell. 1996. Effect of acid deposition on soil acidification and metal mobilisation. *Appl. Geochem.* 11(1-2): 139-143.
- Mathews, R.C., and E.L. Morgan. 1982. Toxicity of Anakeesta formation leachates to shovel-nosed salamander, Great Smoky Mountains National Park. *J Environ Qual* 11, 102-106.
- Medla, A., R. Stangla, S.B. Kikutab, F. Florinetha. (2017). Vegetation establishment on 'Green Walls': Integrating shotcrete walls from road construction into the landscape. *Urban Forestry & Urban Greening* 25: 26-35.
- Miller W.L., C.L. Godfrey, W.G. McGully and G.W. Thomas. 1976. Formation of soil acidity in carbonaceous soil materials exposed by highway excavation in East Texas. *Soil Sci* 121(3):162-169.
- Mitchell, M. J., Mayer, B., Bailey, S. W., Hornbeck, J. W., Alewell, C., Driscoll, C.T., et al. 2001. Use of stable isotope ratios for evaluating sulfur sources and losses at the Hubbard Brook experimental forest. *Water, Air, and Soil Pollution* 130: 75-86.
- Morgan, E. L., W. F. Porak, and J. A. Arway. 1982. Controlling acidic-toxic metal leachates from southern Appalachian construction slopes: mitigating stream damage. *Trans. Res. Record.* 948:10-16.
- Mulholland, P.J. 1993. Hydrometric and stream chemistry evidence of three storm flowpaths in Walker Branch Watershed. *Journal of Hydrology* 151: 291-316.
- Neff, K.J., J.S. Schwartz, T.B. Henry, R.B. Robinson, S.E. Moore, and M.A. Kulp. 2009. Physiological stress in native brook trout during episodic stream acidification in the Great Smoky Mountains National Park. *Archives, Environ. Contamin. & Toxicology* 57: 366-376.
- Neff, K.J., J.S. Schwartz, S.E. Moore, and M.A. Kulp. 2012. Influence of basin characteristics on episodic stream acidification in the Great Smoky Mountains National Park, USA. *Hydrological Processes* 27: 2061-2074. DOI 10.1002/hyp.9366.
- Neville C.M., and P.G.C. Campbell. 1988. Possible mechanisms of aluminum toxicity in a dilute, acidic environment to fingerlings and older life stages of salmonids. *Water Air Soil Pollution* 42: 311-327.
- Newman, K., and A. Dolloff. 1995. Responses to blacknose dace (*Rhinichthys atratulus*) and brook char (*Salvelinus fontinalis*) to acidified water in a laboratory stream. *Water, Air and Soil Pollution* 85: 371-376.
- Nodvin, S.C., C.T. Driscoll, and G.E. Likens. 1986. The effect of pH on sulfate adsorption by a forest soil. *Soil Science* 142(2): 69-75.
- Nordstrom, D.K. 2000. Advances in the Hydrogeochemistry and Microbiology of Acid Mine Waters. *International Geology Review* Vol. 42 (6): 499-515.
- NRCS. 1986. *Urban Hydrology for Small Watersheds. Technical Release 55 (TR-55)*, USDA Natural Resource Conservation Service, Washington, D.C.
- Nyquist, J. and M. Greger. 2008. A field study of constructed wetlands for preventing and treating acid mine drainage. *Ecological Engineering* 35 (5): 630-642.
- Olyphant, G.A., E.R. Bayless, and D. Harper. 1991. Seasonal and weather-related controls on solute concentrations and acid drainage from a pyritic coal-refuse deposit in southwestern Indiana, USA. *Journal of Contaminant Hydrology* 7 (3): 219-236.

- Pinson W.T., D.C. Yoder, J.R. Buchanan, W.C. Wright, and J.B. Wilkerson. 2003. Design and evaluation of an improved flow divider for sampling runoff plots. *Applied Engineering in Agriculture* 20: 433-437.
- Pond, G.J., M.E. Passmore, F.A. Borsuk, L. Reynolds, and C.J. Rose. 2008. Downstream effects of mountaintop coal mining: comparing biological conditions using family- and genus level macroinvertebrate bioassessment tools. *J. N. Am. Benthol. Soc.* 27(3):717-737.
- Prometheus. 2018. <http://prometheusconstruction.com/Pali+Highway+Shotcrete+Honolulu>.
- Pye, K. and J. A. Miller. 1990. Chemical and biochemical weathering of pyritic mudrocks in a shale embankment. *Quart. J. of Eng. Geol.* 23:365-381.
- Orndorff, Z., and W.L. Daniels. 2002. *Final Contract Report – Delineation and Management of Sulfidic Materials in Virginia Highway Corridors*. Virginia Department of Transportation and the University of Virginia) Charlottesville, Virginia. September 2002; VTRC 03-CR3.
- Orndorff, Z.W., and W. L. Daniels. 2004. Evaluation of Acid-producing Sulfidic Materials in Virginia Highway Corridors. *Environmental Geology* 46 (2): 209–16.
- Palmer, S.M., C.T. Driscoll, and C.E. Johnson. 2004. Longterm trends in soil solution and stream water chemistry at the Hubbard Brook experimental forest: Relationship with landscape position. *Biogeochemistry* 68: 51–70.
- Peppas, A., K. Komnitsas, and I. Halikia. 2000. Use of Organic Covers for Acid Mine Drainage Control. *Minerals Engineering* 13 (5): 563–74.
- Qiao, P., and Z. Zhou. 2017. *Best Practices of Using Shotcrete for Wall Fascia and Slope Stabilization (Phase I Study)*. WSDOT Research Report WA-RD 870.1.
- Ray, D., M. Clark, and T. Pitman. 2009. Treatment of an iron-rich ARD using waste carbonate rock: Bench-scale reactor test results. *Mine Water and the Environment* 28 (4): 253–263.
- Rimstidt, J.D, and D.J. Vaughan. 2003. Pyrite oxidation: A state-of-the-art assessment of the reaction mechanism. *Geochimica et Cosmochimica Acta*, Vol. 67 (5): 873–880.
- Rohwerder, T., T. Gehrke, K. Kinzler, and W. Sand. 2003. Bioleaching review part A: Progress in bioleaching: fundamentals and mechanisms of bacterial metal sulfide oxidation. *Appl. Microbiol. Biotechnol.* 63: 239-248.
- Sahat, A.M. and C.W. Sum. 1990. Disintegration of road aggregates along parts of the Gurun-Alor Setar and Ipoh-Changkat Jering highways. *Warta Geologi.* 16(3):142.
- Scheetz, B.E., and C.J. Ellsworth. 2007. *Preliminary Assessment of Acid Producing Rock on Future Penn DOT Construction*. FHWA-PA-2007-002-510401-05. Pennsylvania State University, University Park, PA.
- Schwartz, J.S., Cai, M., Kulp, M.A., Moore, S.E., Nichols, B., Parker, C., 2014. Biological effects of stream water quality on aquatic macroinvertebrates and fish communities within Great Smoky Mountains National Park. US Department of Interior, National Park Service. Natural Resource Report NPS/GRSM/NRR-2014/778.
- Sheoran, A.S., and V. Sheoran. 2006. Heavy metal removal mechanism of acid mine drainage in wetlands: A critical review. *Minerals Engineering* 19 (2): 105–116.
- Sicree, A.A. 2006. Regional mineralogy of sulfide deposits north and south of the I-99 road cut at Skytop, Centre County, Pennsylvania. In *Geological Society of America Abstracts with Programs*, Vol. 38:, No. 2,;p. 33. Geological Society of America, https://gsa.confex.com/gsa/2006NE/finalprogram/abstract_100889.htm.
- Simonin, H.A., W.A. Kretser, D.W. Bath, M. Olson and J. Gallagher. 1993. In situ bioassays of brook trout (*Salvelinus fontinalis*) and blacknose dace (*Rhinichthys atratulus*) in Adirondack streams affected by episodic acidification. *Can. J. Fish. Aquat. Sci.* 50: 902-912.

- Sobek, A.A., W.A. Schuller, J.R. Freeman, and R.M. Smith. 1978. *Field and Laboratory Methods Applicable to Overburden and Mine Spoils*. U.S. Environmental Protection Agency Report, EPA-600/2-78-054. Cincinnati, Ohio.
- Sobek A.A., J.G. Skousen, S.E. Fisher Jr. 2000. Chemical and physical properties of overburdens and mine soils. *In: Barnhisel RI, Darmody RG, Lee Daniels W (eds) Reclamation of Drastically Disturbed Lands*. Am Soc Agron, Madison, WI, Monograph, No 41: 77–104.
- Solmax. 2018. <http://www.gseworld.com/Products/>
- Sokolova, T.A., and S.A. Alekseeva. 2008. Adsorption of sulfate ions by soils (a review). *Eurasian Journal of Soil Science* 41(2): 140–148.
- Spratt, H. G. (1998). Organic sulfur and the retention of nutrient cations in forest surface soils. *Water, Air, and Soil Pollution* 105: 305–317.
- Spry, D. J. and J. G. Wiener (1991). Metal bioavailability and toxicity to fish in low-alkalinity lakes: a critical review. *Environmental Pollution* 71: 243-304.
- Sracek, O., M. Choquette, P. Gélinas, R. Lefebvre, and R.V. Nicholson. 2004. Geochemical Characterization of Acid Mine Drainage from a Waste Rock Pile, Mine Doyon, Quebec, Canada. *Journal of Contaminant Hydrology* 69 (1): 45–71.
- Stevens, C.J., N.B. Dise, and D.J. Gowing. 2009. Regional trends in soil acidification and exchangeable metal concentrations in relation to acid deposition rates. *Environmental Pollution* 157(1): 313–319.
- Sullivan T.J., and B.J. Cosby. 1998. Modeling the concentration of aluminum in surface waters. *Water Air Soil Pollution* 105:643–659.
- TDOT. 2007. *Guidelines for Acid Producing Rock Investigation, Testing, Monitoring and Mitigation*. Prepared for the Tennessee Department of Transportation, Nashville, and prepared by Golder Associates, Inc., Lakewood, CO.
- Tipping, E., et al. 2003. The solid-solution partitioning of heavy metals (Cu, Zn, Cd, Pb) in upland soils of England and Wales. *Environmental Pollution* 125(2): 213–225.
- Tuttle, M.L.W., and G.N. Breit. 2009. Weathering of the New Albany Shale, Kentucky, USA: I. Weathering zones defined by mineralogy and major-element composition. *Applied Geochemistry* 24 (8): 1549–64.
- U.S. EPA. *A Field-Based Aquatic Life Benchmark for Conductivity in Central Appalachian Streams (2010)* (External Review Draft). U.S. Environmental Protection Agency, Washington, DC, EPA/600/R-10/023A, 2010.
- Vear, A., and C. Curtis. 1981. A Quantitative Evaluation of Pyrite Weathering.” *Earth Surface Processes and Landforms* 6 (2): 191–198.
- Webb, J.S., S. McGinness, and H.M. Lappin-Scott. 1998. Metal removal by sulphate-reducing bacteria from natural and constructed wetlands. *J. Appl. Microbiol.* 84 (2): 240–48.
- Wigington, P.J., D.R. DeWalle, et al. 1996. Episodic acidification of small streams in the northeastern US: ionic controls of episodes. *Ecological Applications* 6(2): 389-407.
- Wren, C.D., and G.L. Stephenson. 1991. The effect of acidification on the accumulation and toxicity of metals to freshwater invertebrates. *Environmental Pollution* 71: 205-241.

Appendices

Appendix A: Photo Record of Road Cut Study Sites.

Appendix B: Surficial Pyritic Rock Formations in Tennessee

Appendix C: Monitoring Station Storm Event Precipitation Depths and Runoff Volumes

Appendix D: Water Chemistry for Monitoring Station Storm Events

Appendix E: Literature Review of Toxicity Thresholds to Acidification for Aquatic Biota

Appendix F: Pollutant Export Mass Loadings from Monitoring Station Storm Events

Appendix G: EPT XTRAM® HPL Specifications

Appendix H: Water Chemistry for Middlebrook Pike Experiment: Simulated Rainfall

Appendix I: Water Chemistry for Middlebrook Pike Experiment: Natural Rainfall

Appendix A: Photo Record of Road Cut Study Sites

Grainger 1

Grainger County, 36°21'1.52"N, 83°20'16.76"W
SR-32, Chattanooga Shale



Pinson et al. (2003) devices under construction

Jamestown 1

Fentress County, 36°29'49.48"N, 36°29'49.48"N
SR-28, Fentress Formation



Google Images, 2017



Pinson et al. (2003) devices under construction

Jamestown 2

Fentress County, 36°29'42.95"N, 84°58'05.10"W
SR-28, Fentress Formation



Google Images, 2017



Pinson et al. (2003) devices under construction

Jamestown 3

Fentress County, 36°26'28.27"N, 84°57'46.52"W
SR-52, Fentress Formation

Google Image, 2017



Pinson et al. (2003) devices under construction

Nashville 1

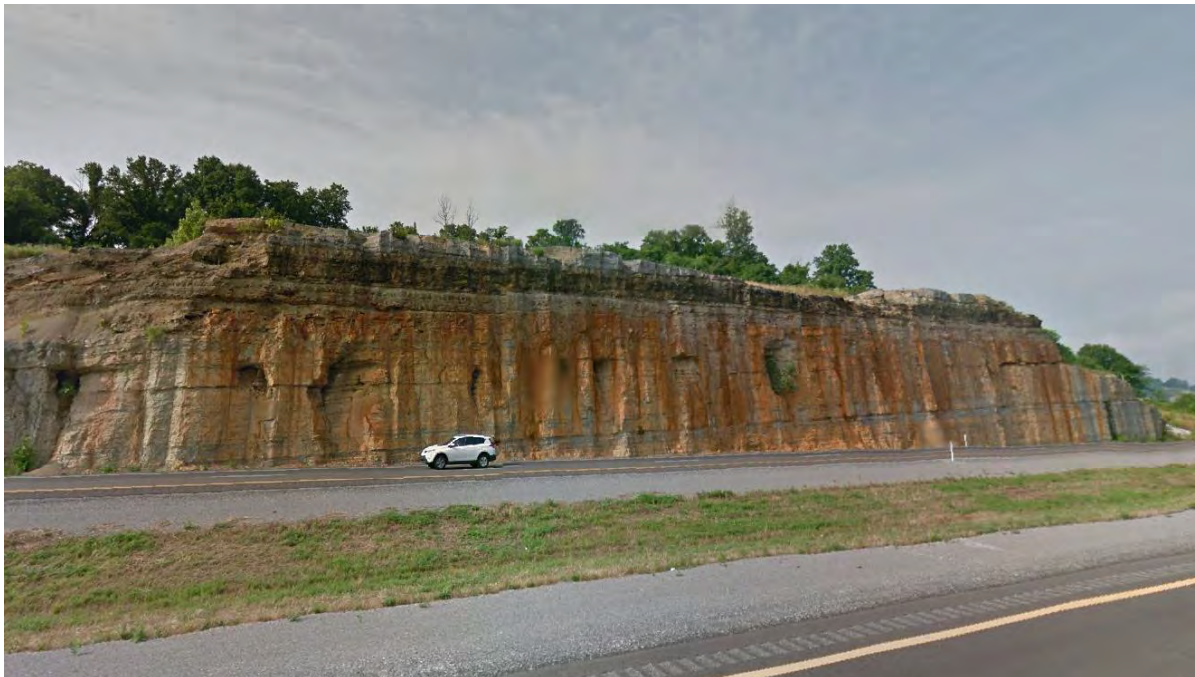
Williamson County, 35°48'52.44"N, 86°58'30.49"W
I-840, Chattanooga Shale

Google Image, 2017



Nashville 2
Williamson County, 35°49'17.21"N, 86°57'11.81"W
I-840, Chattanooga Shale

Google Image, 2017



Ocoee 1

Polk County, 35° 7'13.81"N, 84°34'5.49"W
SR-30, Sandsuck Formation



Google Image, 2017
Winter photo



Ocoee 1 Site, monitoring equipment
on the right of photo.
Summer photo



Ocoee 2

Polk County, 35° 7'13.90"N, 84°34'4.71"W
SR-30, Sandsuck Formation



Google Image, 2017
Winter photo



Ocoee 2 Site, monitoring equipment
on the left of photo.
Summer photo



Ocoee 3

Polk County, 35° 2'35.90"N, 84°26'59.61"W

US-64, Great Smoky Group: Anakeesta



Sevierville 1

Sevier County, 35°49'1.80"N, 83°27'39.95"W

Pitman Center Road (CR 416), Snowbird Group: Roaring Fork



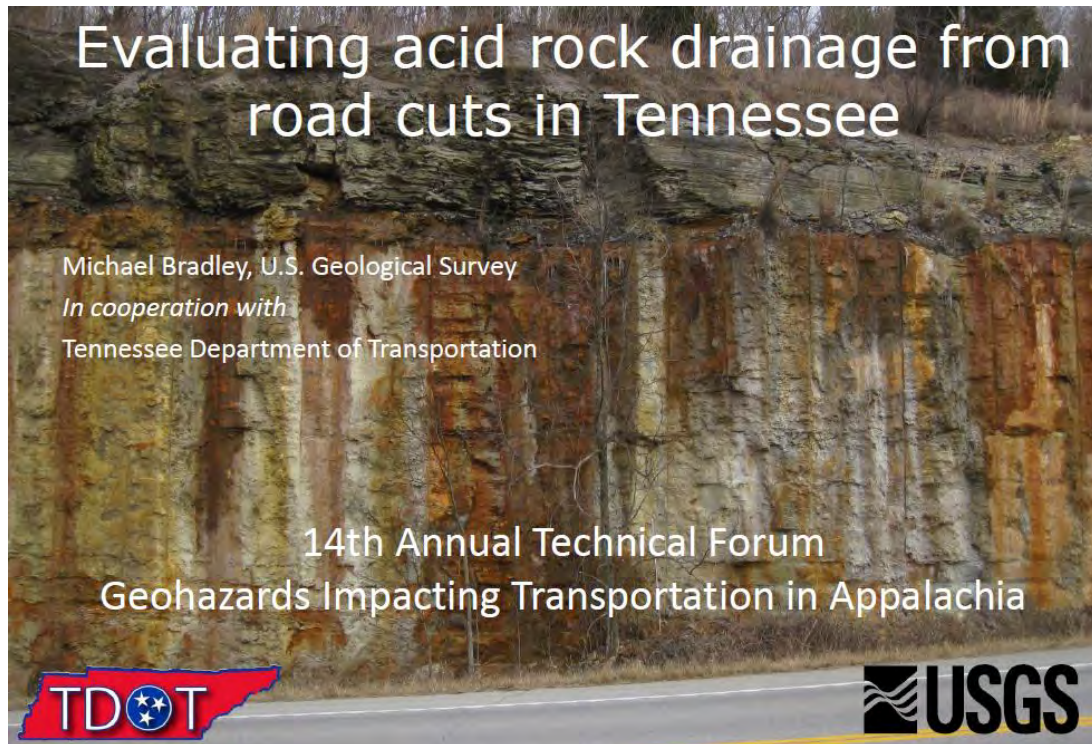
Google Image, 2017



Appendix B: Surficial Pyritic Rock Formations in Tennessee

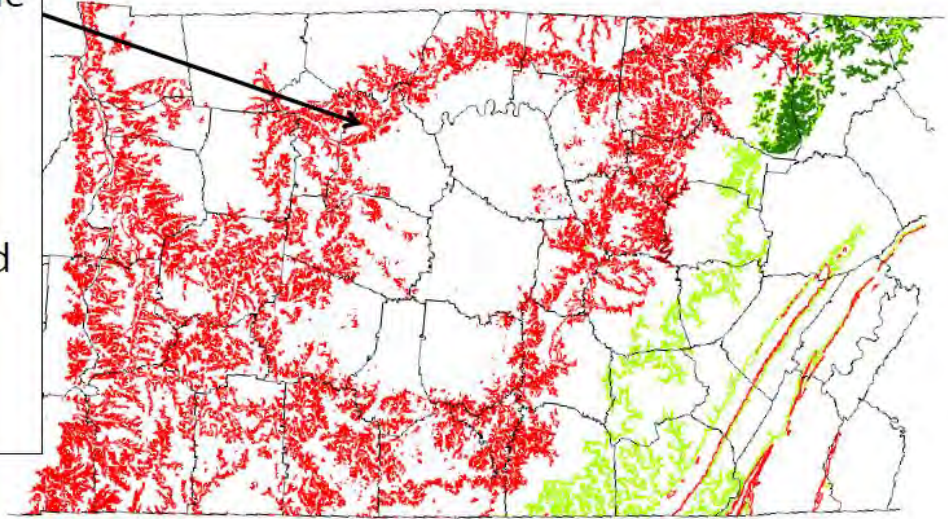
Material obtained from USGS, Nashville Office.

Reference: Michael Bradley



Chattanooga Shale

15-30 ft thick
Trace metals
Radionuclides
Disseminated and
clustered pyrite

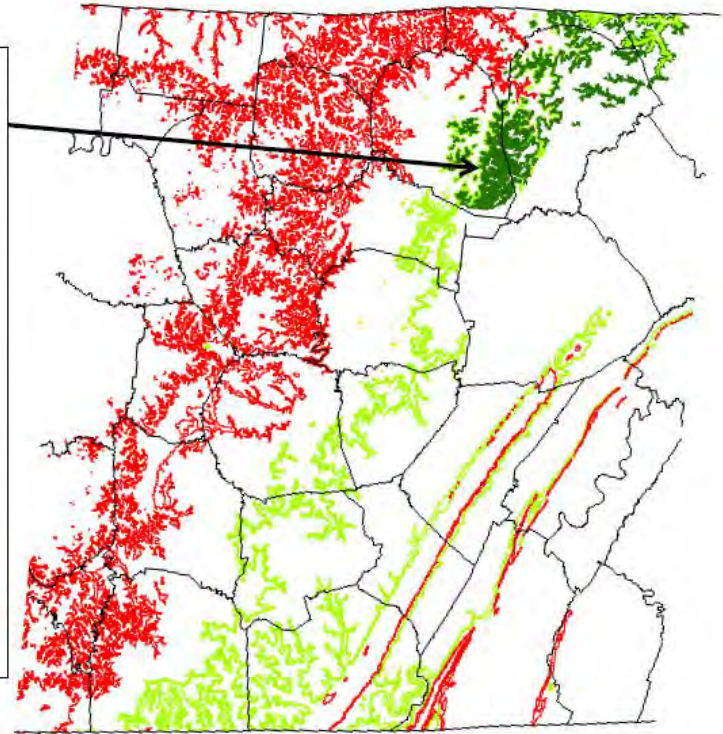


Fentress Formation

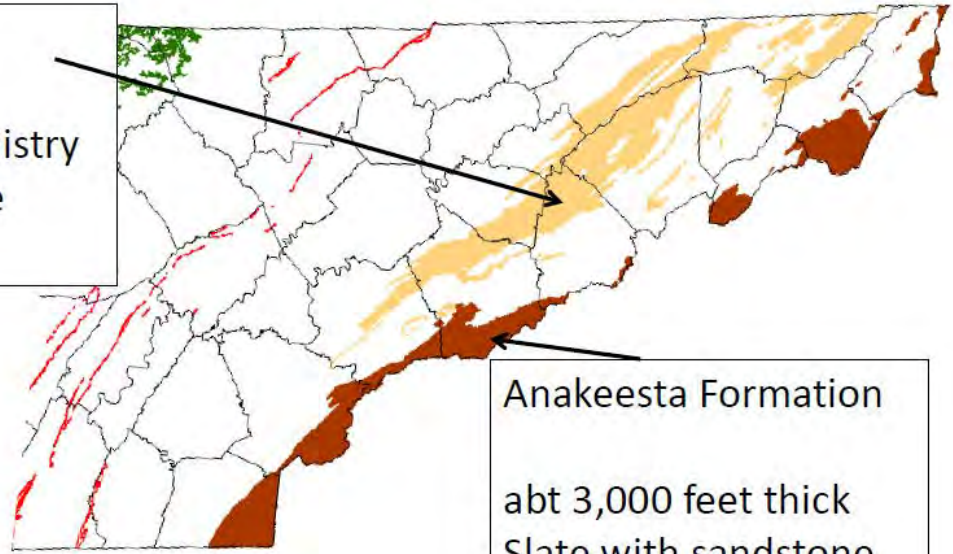
Up to 300 feet thick

Lower unit has multiple coal/shale sequences with high pyrite -- variable

Problem primarily in NW Cumberland Plateau



Sevier Shale
Variable geochemistry
Locally high pyrite



Anakeesta Formation
abt 3,000 feet thick
Slate with sandstone
Adversely affected
ecosystems in GRSM
stream

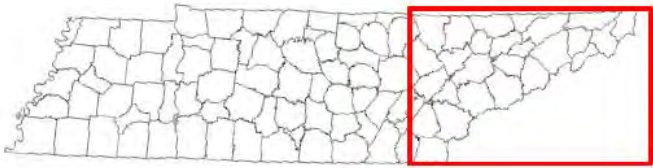
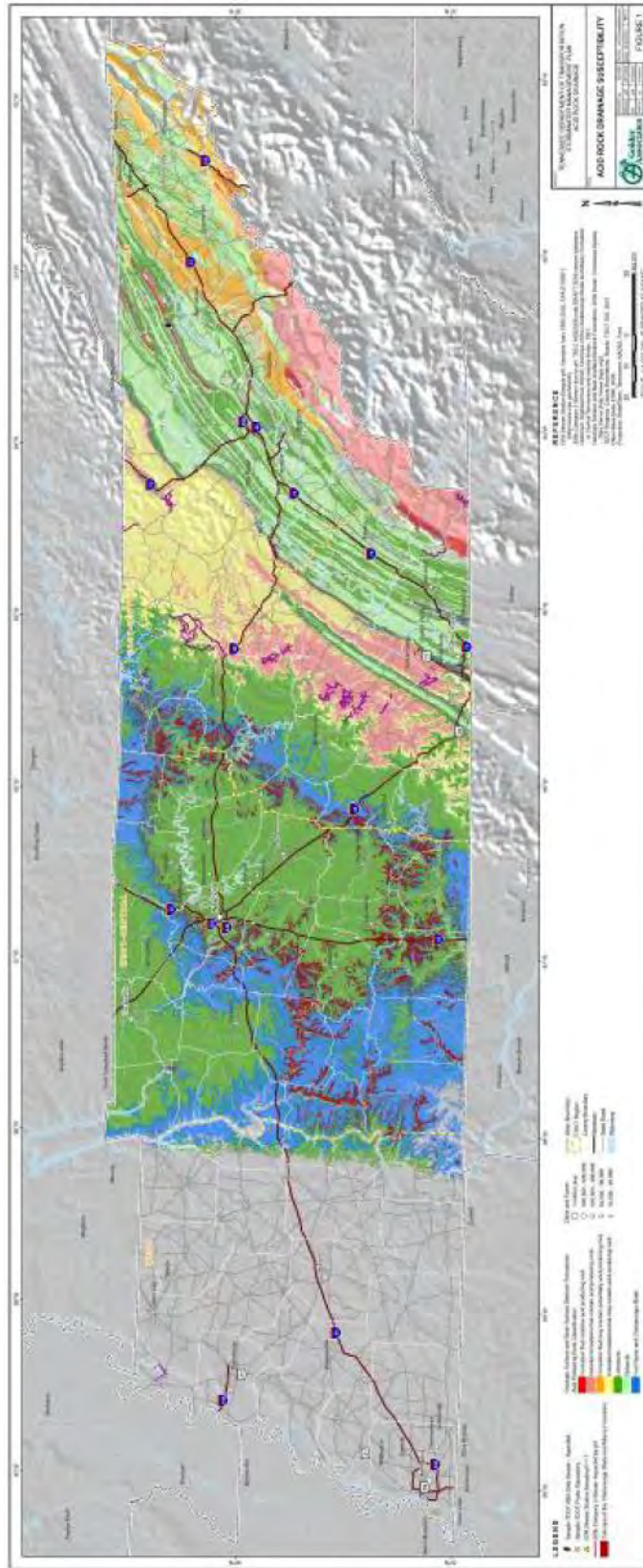


Figure from TDOT (2007), Risk map of acid producing rock.



**Appendix C: Road Cut Monitoring Stations: Storm Event Rainfall
Characteristics and Runoff Volumes.**

APPENDIX C: ROAD CUT SITE RAINFALL AND RUNOFF ESTIMATES

| Site | Date Collected | Rainfall (RF) (days) | RF Average Intensity (cm/hr) | RF Max. Intensity (cm/hr) | Precipitation Depth (cm) | Runoff Depth (cm) | Runoff Volume (m ³) | Runoff Volume * |
|-----------------------|----------------|----------------------|------------------------------|---------------------------|--------------------------|-------------------|---------------------------------|-----------------|
| 1. Grainger | | | | | | | | |
| G1 | 7/21/2014 | 3 | 0.18 | 0.81 | 4.750 | 3.922 | 1.86 | 10.16 |
| G1 | 10/10/2014 | 1 | 0.30 | 0.91 | 3.353 | 2.565 | 1.22 | 6.65 |
| G1 | 10/23/2014 | 6 | 0.18 | 0.66 | 7.518 | 6.651 | 3.15 | 17.23 |
| G1 | 1/29/2015 | 3 | 0.08 | 0.48 | 1.905 | 1.206 | 0.57 | 3.12 |
| G1 | 4/16/2015 | 3 | 0.20 | 0.97 | 3.658 | 2.859 | 1.35 | 7.41 |
| G1 | 5/18/2015 | 3 | 0.10 | 0.23 | 0.610 | 0.165 | 0.08 | 0.43 |
| G1 | 6/9/2015 | 2 | 0.33 | 0.89 | 3.251 | 2.468 | 1.17 | 6.39 |
| G1 | 7/6/2015 | 4 | 0.18 | 1.12 | 4.877 | 4.046 | 1.92 | 10.48 |
| 2. Jamestown 1 | | | | | | | | |
| J1 | 7/22/2014 | 2 | 0.08 | 0.33 | 2.235 | 0.413 | 1.04 | 2.02 |
| J1 | 7/31/2014 | 1 | 0.33 | 2.11 | 5.182 | 2.380 | 6.01 | 11.61 |
| J1 | 9/15/2014 | 1 | 0.18 | 1.17 | 2.286 | 0.438 | 1.11 | 2.14 |
| J1 | 12/10/2014 | 1 | 0.15 | 0.51 | 2.540 | 0.569 | 1.44 | 2.77 |
| J1 | 1/2/2015 | 3 | 0.05 | 0.33 | 2.057 | 0.331 | 0.84 | 1.61 |
| J1 | 1/26/2015 | 1 | 0.05 | 0.13 | 0.762 | 0.000 | 0.00 | 0.00 |
| J1 | 4/16/2015 | 2 | 0.05 | 0.25 | 2.184 | 0.389 | 0.98 | 1.90 |
| J1 | 6/20/2015 | 1 | 0.10 | 0.46 | 1.905 | 0.266 | 0.67 | 1.30 |
| J1 | 7/24/2015 | 1 | 0.28 | 0.97 | 2.286 | 0.438 | 1.11 | 2.14 |
| 3. Jamestown 2 | | | | | | | | |
| J2 | 7/15/2014 | 1 | 0.23 | 0.69 | 2.388 | 0.489 | 1.78 | 3.43 |
| J2 | 9/15/2014 | 1 | 0.18 | 1.17 | 2.286 | 0.438 | 1.59 | 3.07 |
| J2 | 12/10/2014 | 1 | 0.15 | 0.51 | 2.540 | 0.569 | 2.07 | 3.99 |
| J2 | 1/2/2015 | 3 | 0.05 | 0.33 | 2.057 | 0.331 | 1.20 | 2.32 |
| J2 | 1/26/2015 | 1 | 0.05 | 0.13 | 0.762 | 0.000 | 0.00 | 0.00 |
| J2 | 4/16/2015 | 2 | 0.05 | 0.25 | 2.184 | 0.389 | 1.41 | 2.73 |
| J2 | 6/20/2015 | 1 | 0.10 | 0.46 | 1.905 | 0.266 | 0.97 | 1.86 |
| 4. Jamestown 3 | | | | | | | | |
| J3 | 7/4/2014 | 3 | 0.05 | 0.28 | 0.711 | 0.048 | 0.02 | 0.12 |
| J3 | 7/15/2014 | 1 | 0.15 | 0.38 | 1.473 | 0.396 | 0.18 | 0.97 |
| J3 | 7/22/2014 | 2 | 0.08 | 0.36 | 2.489 | 1.105 | 0.49 | 2.69 |
| J3 | 7/31/2014 | 1 | 0.36 | 2.29 | 5.842 | 4.044 | 1.80 | 9.86 |
| J3 | 9/15/2014 | 1 | 0.20 | 1.30 | 2.565 | 1.164 | 0.52 | 2.84 |
| J3 | 12/10/2014 | 1 | 0.18 | 0.51 | 2.692 | 1.264 | 0.56 | 3.08 |
| J3 | 1/2/2015 | 3 | 0.05 | 0.38 | 2.184 | 0.874 | 0.39 | 2.13 |
| J3 | 1/22/2015 | 1 | 0.08 | 0.30 | 1.245 | 0.268 | 0.12 | 0.65 |
| J3 | 1/26/2015 | 1 | 0.05 | 0.15 | 0.787 | 0.071 | 0.03 | 0.17 |
| J3 | 4/16/2015 | 2 | 0.05 | 0.23 | 2.032 | 0.764 | 0.34 | 1.86 |
| J3 | 6/20/2015 | 1 | 0.10 | 0.53 | 2.032 | 0.764 | 0.34 | 1.86 |
| J3 | 7/24/2015 | 1 | 0.25 | 1.19 | 2.362 | 1.007 | 0.45 | 2.46 |

| Site | Date Collected | Rainfall (RF) (days) | RF Average Intensity (cm/hr) | RF Max. Intensity (cm/hr) | Precipitation Depth (cm) | Runoff Depth (cm) | Runoff Volume (m³) | Runoff Volume * m ³ |
|-----------------------|-----------------------|-----------------------------|-------------------------------------|----------------------------------|---------------------------------|--------------------------|--------------------------------------|--|
| 5. Nashville 1 | | | | | | | | |
| N1 | 6/8/2014 | 1 | 0.30 | 1.07 | 2.210 | 1.690 | 0.61 | 3.35 |
| N1 | 7/1/2014 | 2 | 0.15 | 0.51 | 1.600 | 1.112 | 0.40 | 2.20 |
| N1 | 7/22/2014 | 1 | 0.48 | 2.11 | 7.671 | 7.082 | 2.57 | 14.03 |
| N1 | 8/26/2014 | 1 | 0.05 | 0.23 | 0.610 | 0.250 | 0.09 | 0.50 |
| N1 | 10/11/2014 | 2 | 0.38 | 1.30 | 5.436 | 4.859 | 1.76 | 9.63 |
| N1 | 1/9/2015 | 2 | 0.15 | 0.61 | 3.277 | 2.727 | 0.99 | 5.40 |
| N1 | 1/30/2015 | 2 | 0.05 | 0.25 | 1.168 | 0.716 | 0.26 | 1.42 |
| N1 | 3/23/2015 | 2 | 0.05 | 0.15 | 0.737 | 0.348 | 0.13 | 0.69 |
| N1 | 6/1/2015 | 2 | 0.25 | 1.14 | 5.715 | 5.137 | 1.86 | 10.18 |
| N1 | 7/24/2015 | 1 | 0.18 | 0.46 | 1.575 | 1.088 | 0.39 | 2.16 |
| N1 | 8/21/2015 | 2 | 0.23 | 0.71 | 3.505 | 2.952 | 1.07 | 5.85 |
| 6. Nashville 2 | | | | | | | | |
| N2 | 7/1/2014 | 2 | 0.15 | 0.51 | 1.600 | 1.112 | 0.31 | 1.69 |
| N2 | 8/26/2014 | 1 | 0.05 | 0.23 | 0.610 | 0.250 | 0.07 | 0.38 |
| N2 | 10/11/2014 | 2 | 0.38 | 1.30 | 5.436 | 4.859 | 1.35 | 7.41 |
| N2 | 1/9/2015 | 2 | 0.15 | 0.61 | 3.277 | 2.727 | 0.76 | 4.16 |
| N2 | 3/23/2015 | 2 | 0.05 | 0.15 | 0.737 | 0.348 | 0.10 | 0.53 |
| N2 | 6/1/2015 | 2 | 0.25 | 1.14 | 5.715 | 5.137 | 1.43 | 7.83 |
| N2 | 7/24/2015 | 1 | 0.18 | 0.46 | 1.575 | 1.088 | 0.30 | 1.66 |
| N2 | 8/21/2015 | 2 | 0.23 | 0.71 | 3.505 | 2.952 | 0.82 | 4.50 |
| 7. Ocoee 1 | | | | | | | | |
| O1 | 6/12/2014 | 3 | 0.10 | 0.79 | 2.642 | 1.224 | 0.55 | 1.49 |
| O1 | 6/19/2014 | 2 | 0.15 | 0.86 | 0.940 | 0.126 | 0.06 | 0.15 |
| O1 | 6/26/2014 | 4 | 0.15 | 2.03 | 3.962 | 2.334 | 1.04 | 2.85 |
| O1 | 7/1/2014 | 3 | 0.23 | 2.21 | 6.248 | 4.424 | 1.97 | 5.39 |
| O1 | 7/17/2014 | 6 | 0.08 | 0.79 | 2.007 | 0.746 | 0.33 | 0.91 |
| O1 | 7/20/2014 | 2 | 0.08 | 0.61 | 2.438 | 1.065 | 0.48 | 1.30 |
| O1 | 9/28/2014 | 1 | 0.10 | 0.33 | 1.092 | 0.192 | 0.09 | 0.23 |
| O1 | 11/2/2014 | 4 | 0.03 | 0.30 | 0.787 | 0.071 | 0.03 | 0.09 |
| O1 | 1/3/2015 | 1 | 0.03 | 0.10 | 0.483 | 0.005 | 0.00 | 0.01 |
| O1 | 1/25/2015 | 2 | 0.08 | 2.26 | 2.388 | 1.027 | 0.46 | 1.25 |
| O1 | 2/10/2015 | 1 | 0.13 | 0.64 | 1.245 | 0.268 | 0.12 | 0.33 |
| O1 | 4/8/2015 | 1 | 0.10 | 0.76 | 1.651 | 0.506 | 0.23 | 0.62 |
| O1 | 5/18/2015 | 4 | 0.08 | 0.69 | 3.886 | 2.267 | 1.01 | 2.76 |
| O1 | 6/9/2015 | 1 | 0.18 | 0.84 | 3.023 | 1.531 | 0.68 | 1.87 |
| O1 | 7/6/2015 | 4 | 0.18 | 1.47 | 13.411 | 11.362 | 5.07 | 13.85 |
| 8. Ocoee 2 | | | | | | | | |
| O2 | 6/26/2014 | 4 | 0.15 | 2.03 | 3.962 | 2.334 | 0.87 | 2.85 |
| O2 | 7/1/2014 | 3 | 0.23 | 2.21 | 6.248 | 4.424 | 1.64 | 5.39 |
| O2 | 7/17/2014 | 6 | 0.08 | 0.79 | 2.007 | 0.746 | 0.28 | 0.91 |
| O2 | 7/20/2014 | 2 | 0.08 | 0.61 | 2.438 | 1.065 | 0.40 | 1.30 |

8. Ocoee 2 continued

| Site | Date Collected | Rainfall (RF) (days) | RF Average Intensity (cm/hr) | RF Max. Intensity (cm/hr) | Precipitation Depth (cm) | Runoff Depth (cm) | Runoff Volume (m³) | Runoff Volume * |
|-------------|-----------------------|-----------------------------|-------------------------------------|----------------------------------|---------------------------------|--------------------------|--------------------------------------|------------------------|
| O2 | 9/28/2014 | 1 | 0.10 | 0.33 | 1.092 | 0.192 | 0.07 | 0.23 |
| O2 | 11/2/2014 | 4 | 0.03 | 0.30 | 0.787 | 0.071 | 0.03 | 0.09 |
| O2 | 1/3/2015 | 1 | 0.03 | 0.10 | 0.483 | 0.005 | 0.00 | 0.01 |
| O2 | 1/25/2015 | 2 | 0.08 | 2.26 | 2.388 | 1.027 | 0.38 | 1.25 |
| O2 | 2/10/2015 | 1 | 0.13 | 0.64 | 1.245 | 0.268 | 0.10 | 0.33 |
| O2 | 4/8/2015 | 1 | 0.10 | 0.76 | 1.651 | 0.506 | 0.19 | 0.62 |
| O2 | 5/18/2015 | 4 | 0.08 | 0.69 | 3.886 | 2.267 | 0.84 | 2.76 |
| O2 | 6/9/2015 | 1 | 0.18 | 0.84 | 3.023 | 1.531 | 0.57 | 1.87 |
| O2 | 7/6/2015 | 4 | 0.18 | 1.47 | 13.411 | 11.362 | 4.22 | 13.85 |

9. Ocoee 3

| | | | | | | | | |
|----|-----------|---|------|------|--------|--------|------|-------|
| O3 | 7/1/2014 | 3 | 0.33 | 1.52 | 8.001 | 6.090 | 3.23 | 17.64 |
| O3 | 7/17/2014 | 6 | 0.13 | 0.86 | 3.023 | 1.531 | 0.81 | 4.43 |
| O3 | 7/20/2014 | 2 | 0.08 | 0.38 | 2.667 | 1.244 | 0.66 | 3.60 |
| O3 | 9/28/2014 | 1 | 0.13 | 0.38 | 1.295 | 0.295 | 0.16 | 0.85 |
| O3 | 11/2/2014 | 4 | 0.05 | 0.36 | 1.041 | 0.169 | 0.09 | 0.49 |
| O3 | 1/3/2015 | 1 | 0.03 | 0.18 | 0.686 | 0.042 | 0.02 | 0.12 |
| O3 | 1/22/2015 | 1 | 0.05 | 0.36 | 1.194 | 0.242 | 0.13 | 0.70 |
| O3 | 1/25/2015 | 2 | 0.10 | 0.23 | 2.362 | 1.007 | 0.53 | 2.92 |
| O3 | 2/10/2015 | 1 | 0.03 | 0.13 | 0.330 | 0.001 | 0.00 | 0.00 |
| O3 | 4/8/2015 | 1 | 0.10 | 0.36 | 1.524 | 0.427 | 0.23 | 1.24 |
| O3 | 5/18/2015 | 4 | 0.13 | 0.33 | 5.131 | 3.385 | 1.79 | 9.80 |
| O3 | 6/9/2015 | 1 | 0.28 | 1.17 | 3.962 | 2.334 | 1.24 | 6.76 |
| O3 | 7/6/2015 | 4 | 0.23 | 1.19 | 16.078 | 13.992 | 7.41 | 40.51 |
| O3 | 8/8/2015 | 2 | 0.05 | 0.23 | 0.406 | 0.000 | 0.00 | 0.00 |

10. Seiverville

| | | | | | | | | |
|----|------------|---|------|------|-------|-------|------|-------|
| S1 | 7/17/2014 | 3 | 0.15 | 1.45 | 2.311 | 0.969 | 0.38 | 2.07 |
| S1 | 7/21/2014 | 2 | 0.10 | 0.51 | 3.505 | 1.937 | 0.76 | 4.13 |
| S1 | 7/29/2014 | 1 | 0.91 | 2.92 | 7.315 | 5.434 | 2.12 | 11.59 |
| S1 | 8/10/2014 | 3 | 0.13 | 0.69 | 3.404 | 1.850 | 0.72 | 3.95 |
| S1 | 8/30/2014 | 6 | 0.25 | 0.79 | 1.803 | 0.606 | 0.24 | 1.29 |
| S1 | 11/16/2014 | 1 | 0.13 | 0.61 | 1.753 | 0.572 | 0.22 | 1.22 |
| S1 | 1/22/2015 | 2 | 0.08 | 0.25 | 1.600 | 0.474 | 0.18 | 1.01 |
| S1 | 1/29/2015 | 4 | 0.08 | 0.25 | 2.540 | 1.144 | 0.45 | 2.44 |
| S1 | 4/16/2015 | 2 | 0.10 | 0.48 | 2.794 | 1.345 | 0.52 | 2.87 |
| S1 | 5/18/2015 | 3 | 0.13 | 0.76 | 1.295 | 0.295 | 0.12 | 0.63 |
| S1 | 6/9/2015 | 1 | 0.18 | 0.66 | 2.184 | 0.874 | 0.34 | 1.87 |
| S1 | 7/6/2015 | 4 | 0.13 | 0.43 | 3.658 | 2.068 | 0.81 | 4.41 |

* Per 10 m of road cut

APPENDIX C: ROAD CUT RUNOFF ESTIMATES

Road Dimensions and Hydrologic Parameters

1. Grainger

Width = 1.8288 m
Height = 25.908 m
Area = 47.38 ft²
CN = 97
S = 7.86

Ratio for 10 m

5.468066

2. Jamestown 1

Width = 5.1816 m
Height = 48.768 m
Area = 252.70 m²
CN = 87
S = 37.95

Ratio for 10 m

1.929906

3. Jamestown 2

Width = 5.1816 m
Height = 70.104 m
Area = 363.25 m²
CN = 87
S = 37.95

Ratio for 10 m

1.929906

4. Jamestown 3

Width = 1.8288 m
Height = 24.384 m
Area = 44.59 m²
CN = 93
S = 19.12

Ratio for 10 m

5.468066

5. Nashville 1

Width = 1.8288 m
Height = 19.812 m
Area = 36.23 m²
CN = 98
S = 5.18

Ratio for 10 m

5.468066

6. Nashville 2

Width = 1.8288 m
Height = 15.240 m
Area = 27.87 m²
CN = 98
S = 5.18

Ratio for 10 m

5.468066

Road Dimensions and Hydrologic Parameters

7. Ocoee 1

Width = 3.6576 m
Height = 12.192 m
Area = 44.59 m²
CN = 93
S = 19.12

Ratio for 10 m
2.734033

8. Ocoee 2

Width = 3.048 m
Height = 12.192 m
Area = 37.16 m²
CN = 93
S = 19.12

Ratio for 10 m
3.28084

9. Ocoee 3

Width = 1.8288 m
Height = 28.956 m
Area = 52.96 m²
CN = 93
S = 19.12

Ratio for 10 m
5.468066

10. Seiverville

Width = 1.8288 m
Height = 21.336 m
Area = 39.02 m²
CN = 93
S = 19.12

Ratio for 10 m
5.468066

Review of Runoff Calculations

Precipitation depth (cm) = P (per storm event)

Hydrology Curve Number = CN (same per site)

Drainage Area (Area) is the exposed surface to precipitation

Potential Maximum Retention = S

$S = 25400/CN - 254$ (metric equation, same per site)

Runoff Depth = P_e (cm) $P_e = (P - 0.2 * S) / (P + 0.08 * S)$

Runoff Depth (P_e) converted from cm to m by dividing by 100

Runoff Volume = P_e (m) * Area (m²)

Runoff volume per 10 m of road cut = 10 m / (width, m)

Appendix D: Water Chemistry for Monitoring Station Storm Events.

ROAD CUT RUNOFF WATER CHEMISTRY: SUMMARY SHEET

| Site | Date Collected | TSS (mg/L) | Conductivity (uS/cm) | pH units | Hardness (mg/L) | ANC (meq/L) | Acid Anions | | | | Base Cations | | |
|-----------------------|----------------|------------|----------------------|----------|-----------------|-------------|-------------|---------------|-------------|---------------|--------------|--|--|
| | | | | | | | Cl- (mg/L) | SO4^2- (mg/L) | NO3- (mg/L) | PO4^3- (mg/L) | Na+ | | |
| 1. Grainger | | | | | | | | | | | | | |
| G1 | 7/21/2014 | 2309 | 209 | 4.93 | 19.03 | -0.240 | 5.975 | 92.972 | 13.911 | 12.866 | 5.494 | | |
| G1 | 10/10/2014 | 2282 | 168 | 4.77 | 19.35 | 0.017 | 6.185 | 84.156 | 14.619 | 11.912 | 5.934 | | |
| G1 | 10/23/2014 | 2168 | 157 | 4.27 | 18.20 | -0.038 | 5.916 | 81.993 | 15.351 | 12.481 | 5.998 | | |
| G1 | 1/29/2015 | 100 | 117 | 4.93 | 16.53 | -0.406 | 6.154 | 97.299 | 14.410 | 12.360 | 6.562 | | |
| G1 | 4/16/2015 | 2423 | 182 | 5.09 | 15.73 | -0.638 | 5.695 | 87.349 | 14.464 | 12.282 | 0.958 | | |
| G1 | 5/18/2015 | 2533 | 174 | 4.85 | 15.11 | -0.730 | 6.036 | 88.412 | 14.182 | 12.840 | 0.820 | | |
| G1 | 6/9/2015 | 2645 | 170 | 5.50 | 23.38 | 0.181 | 5.699 | 95.658 | 16.124 | 13.946 | 6.894 | | |
| G1 | 7/6/2015 | 2449 | 206 | 4.93 | 17.86 | -0.326 | 5.928 | 92.808 | 14.599 | 12.539 | 5.655 | | |
| 2. Jamestown 1 | | | | | | | | | | | | | |
| J1 | 7/22/2014 | 185 | 110 | 6.60 | 14.97 | -1.139 | 6.012 | 80.013 | 14.239 | 12.018 | 3.486 | | |
| J1 | 7/31/2014 | 535 | 133 | 6.60 | 12.55 | -1.358 | 5.786 | 83.960 | 15.366 | 12.235 | 3.874 | | |
| J1 | 9/15/2014 | 153 | 105 | 6.72 | 13.92 | -1.049 | 5.967 | 72.261 | 14.499 | 12.694 | 3.794 | | |
| J1 | 12/10/2014 | 136 | 109 | 6.60 | 15.76 | -0.841 | 6.587 | 65.152 | 14.688 | 12.623 | 3.642 | | |
| J1 | 1/2/2015 | 39 | 95 | 6.97 | 17.06 | -0.865 | 6.017 | 70.402 | 14.687 | 12.716 | 3.664 | | |
| J1 | 1/26/2015 | 39 | 93 | 7.37 | 13.38 | 0.074 | 6.418 | 16.835 | 14.892 | 11.955 | 3.244 | | |
| J1 | 4/16/2015 | 152 | 123 | 6.63 | 12.73 | -1.234 | 6.001 | 68.359 | 14.248 | 12.371 | 0.000 | | |
| J1 | 6/20/2015 | 165 | 126 | 6.57 | 21.77 | 0.227 | 6.433 | 41.789 | 12.926 | 10.865 | 5.098 | | |
| J1 | 7/24/2015 | 141 | 118 | 6.75 | 13.10 | -1.162 | 5.520 | 76.006 | 14.719 | 12.362 | 3.278 | | |
| 3. Jamestown 2 | | | | | | | | | | | | | |
| J2 | 7/15/2014 | 81 | 371 | 3.73 | 44.34 | -13.877 | 17.776 | 788.005 | 39.599 | 29.375 | 13.649 | | |
| J2 | 9/15/2014 | 79 | 300 | 4.13 | 26.91 | -7.483 | 21.694 | 384.187 | 25.215 | 35.717 | 0.000 | | |
| J2 | 12/10/2014 | 80 | 297 | 3.97 | 28.56 | -8.227 | 23.477 | 424.863 | 28.073 | 32.453 | 0.000 | | |
| J2 | 1/2/2015 | 39 | 428 | 6.80 | 27.16 | -2.060 | 7.557 | 136.213 | 40.982 | 28.982 | 0.000 | | |
| J2 | 1/26/2015 | 96 | 22 | 7.07 | 34.44 | -0.259 | 7.408 | 89.922 | 11.132 | 37.169 | 0.000 | | |
| J2 | 4/16/2015 | 63 | 292 | 4.75 | 41.39 | -11.749 | 14.166 | 646.935 | 33.037 | 38.827 | 0.000 | | |
| J2 | 6/20/2015 | 68 | 382 | 3.45 | 45.47 | -10.848 | 21.119 | 629.866 | 58.604 | 38.686 | 23.579 | | |
| 4. Jamestown 3 | | | | | | | | | | | | | |
| J3 | 7/4/2014 | 51 | 33 | 6.07 | 10.28 | 0.168 | 3.660 | 15.583 | 13.737 | 7.608 | 1.606 | | |
| J3 | 7/15/2014 | 63 | 33 | 6.20 | 9.81 | 0.033 | 4.352 | 23.822 | 10.992 | 7.680 | 1.697 | | |

| Site | Date Collected | TSS (mg/L) | Conductivity (uS/cm) | pH units | Hardness (mg/L) | ANC (meq/L) | Acid Anions | | | Base Cations | | |
|-----------------------|----------------|------------|----------------------|----------|-----------------|-------------|-------------|---------------|-------------|---------------|------------|--|
| | | | | | | | Cl- (mg/L) | SO4^2- (mg/L) | NO3- (mg/L) | PO4^3- (mg/L) | Na+ (mg/L) | |
| 4. Jamestown 3 | | | | | | | | | | | | |
| J3 | 7/22/2014 | 47 | 34 | 5.97 | 9.58 | 0.154 | 4.414 | 20.258 | 6.503 | 7.569 | 1.757 | |
| J3 | 7/31/2014 | 47 | 36 | 6.17 | 9.25 | 0.244 | 5.206 | 16.425 | 9.122 | 5.641 | 1.857 | |
| J3 | 9/15/2014 | 39 | 36 | 6.20 | 9.14 | 0.214 | 4.446 | 13.292 | 9.408 | 6.363 | 1.535 | |
| J3 | 12/10/2014 | 42 | 34 | 6.05 | 9.54 | 0.156 | 5.239 | 13.656 | 11.175 | 7.065 | 1.542 | |
| J3 | 1/2/2015 | 40 | 24 | 6.50 | 9.43 | 0.085 | 4.849 | 19.794 | 8.074 | 6.821 | 1.546 | |
| J3 | 1/22/2015 | 59 | 22 | 6.27 | 9.10 | 0.116 | 3.579 | 19.084 | 10.652 | 5.627 | 1.531 | |
| J3 | 1/26/2015 | 39 | 20 | 6.50 | 9.55 | 0.165 | 3.250 | 15.683 | 9.412 | 8.208 | 1.552 | |
| J3 | 4/16/2015 | 50 | 36 | 6.39 | 10.04 | 0.133 | 4.683 | 17.101 | 7.363 | 7.832 | 0.416 | |
| J3 | 6/20/2015 | 35 | 30 | 6.45 | 11.29 | 0.059 | 6.440 | 16.085 | 15.070 | 11.755 | 2.170 | |
| J3 | 7/24/2015 | 46 | 38 | 6.46 | 10.31 | 0.184 | 4.816 | 20.690 | 9.624 | 5.444 | 1.708 | |
| 5. Nashville 1 | | | | | | | | | | | | |
| N1 | 6/8/2014 | 3045 | 4808 | 6.17 | 617.80 | -9.242 | 14.517 | 2213.109 | 33.961 | 29.570 | 29.034 | |
| N1 | 7/1/2014 | 3014 | 4991 | 6.08 | 621.11 | -9.774 | 16.081 | 2247.811 | 32.651 | 24.138 | 26.713 | |
| N1 | 7/22/2014 | 2011 | 5638 | 5.32 | 615.86 | -10.924 | 16.365 | 2281.844 | 33.219 | 27.262 | 25.142 | |
| N1 | 8/26/2014 | 1603 | 5199 | 6.37 | 604.70 | -0.265 | 11.882 | 1766.905 | 27.283 | 9.880 | 21.933 | |
| N1 | 10/11/2014 | 1724 | 4891 | 6.53 | 644.49 | 2.509 | 14.704 | 1764.356 | 18.045 | 9.025 | 27.402 | |
| N1 | 1/9/2015 | 601 | 5226 | 6.45 | 616.26 | 1.081 | 11.551 | 1744.212 | 18.077 | 9.039 | 28.125 | |
| N1 | 1/30/2015 | 39 | 5164 | 6.45 | 624.69 | 1.659 | 11.553 | 1743.903 | 24.603 | 8.876 | 25.114 | |
| N1 | 3/23/2015 | 3136 | 5328 | 5.78 | 798.83 | -3.203 | 13.369 | 2381.333 | 30.245 | 8.854 | 1.022 | |
| N1 | 6/1/2015 | 2211 | 5425 | 5.55 | 826.63 | 2.877 | 11.534 | 2210.103 | 28.754 | 9.530 | 17.700 | |
| N1 | 7/24/2015 | 2693 | 4686 | 6.23 | 610.67 | -10.828 | 17.776 | 2286.072 | 36.574 | 22.486 | 28.431 | |
| N1 | 8/21/2015 | 2545 | 6031 | 6.37 | 616.64 | -11.373 | 18.809 | 2294.043 | 33.567 | 26.889 | 25.484 | |
| 6. Nashville 2 | | | | | | | | | | | | |
| N2 | 7/1/2014 | 384 | 2045 | 4.71 | 662.65 | 28.028 | 16.612 | 861.531 | 40.245 | 34.068 | 30.839 | |
| N2 | 8/26/2014 | 367 | 1958 | 4.92 | 684.20 | 28.832 | 18.293 | 836.095 | 43.454 | 33.285 | 29.287 | |
| N2 | 10/11/2014 | 320 | 1451 | 4.50 | 664.58 | 27.177 | 14.124 | 871.569 | 46.985 | 19.064 | 34.751 | |
| N2 | 1/9/2015 | 135 | 1455 | 6.40 | 540.05 | 31.844 | 12.669 | 734.006 | 20.465 | 42.538 | 274.410 | |
| N2 | 3/23/2015 | 442 | 3188 | 5.55 | 750.67 | 24.951 | 16.477 | 767.379 | 34.451 | 23.437 | 0.000 | |
| N2 | 6/1/2015 | 361 | 3355 | 4.75 | 736.33 | 32.996 | 14.662 | 839.855 | 34.543 | 19.444 | 47.308 | |
| N2 | 7/24/2015 | 408 | 3352 | 3.91 | 659.18 | 28.340 | 11.567 | 859.750 | 14.662 | 31.038 | 28.695 | |
| N2 | 8/21/2015 | 385 | 3460 | 4.75 | 667.89 | 36.652 | 14.933 | 456.263 | 34.230 | 29.433 | 31.386 | |

| Site | Date Collected | TSS (mg/L) | Conductivity (uS/cm) | pH units | Hardness (mg/L) | Acid Anions | | | | Base Cations | | |
|-------------------|----------------|------------|----------------------|----------|-----------------|-------------|------------|---------------|-------------|---------------|------------|--|
| | | | | | | ANC (meq/L) | Cl- (mg/L) | SO4^2- (mg/L) | NO3- (mg/L) | PO4^3- (mg/L) | Na+ (mg/L) | |
| 7. Ocoee 1 | | | | | | | | | | | | |
| O1 | 6/12/2014 | 422 | 131 | 6.83 | 62.22 | 3.554 | 4.446 | 49.754 | 22.911 | 9.634 | 12.185 | |
| O1 | 6/19/2014 | 520 | 230 | 6.60 | 59.34 | 4.022 | 7.380 | 29.702 | 19.044 | 5.952 | 15.042 | |
| O1 | 6/26/2014 | 419 | 116 | 6.60 | 67.87 | 3.181 | 5.744 | 72.578 | 33.202 | 10.920 | 12.964 | |
| O1 | 7/1/2014 | 560 | 92 | 6.92 | 62.38 | 3.625 | 5.104 | 51.091 | 19.804 | 9.318 | 11.783 | |
| O1 | 7/17/2014 | 221 | 348 | 6.97 | 64.41 | 4.121 | 5.953 | 28.850 | 12.539 | 10.779 | 11.554 | |
| O1 | 7/20/2014 | 77 | 113 | 6.53 | 66.29 | 4.146 | 4.194 | 20.010 | 32.505 | 10.863 | 12.064 | |
| O1 | 9/28/2014 | 475 | 184 | 4.68 | 57.68 | 3.874 | 3.360 | 24.478 | 15.865 | 7.191 | 11.128 | |
| O1 | 11/2/2014 | 482 | 183 | 4.63 | 57.48 | 3.451 | 4.059 | 48.931 | 14.508 | 5.633 | 11.986 | |
| O1 | 1/3/2015 | 45 | 136 | 7.57 | 56.92 | 3.710 | 3.829 | 38.396 | 11.883 | 5.557 | 11.438 | |
| O1 | 1/25/2015 | 45 | 155 | 7.33 | 57.64 | 3.874 | 5.628 | 21.151 | 14.315 | 6.935 | 11.379 | |
| O1 | 2/10/2015 | 38 | 96 | 7.23 | 54.46 | 3.764 | 2.873 | 20.455 | 16.595 | 6.631 | 11.488 | |
| O1 | 4/8/2015 | 674 | 174 | 4.60 | 48.00 | 1.840 | 6.238 | 44.745 | 21.419 | 9.469 | 2.154 | |
| O1 | 5/18/2015 | 343 | 215 | 4.61 | 48.10 | 1.928 | 5.428 | 47.303 | 22.214 | 7.486 | 1.725 | |
| O1 | 6/9/2015 | 802 | 264 | 4.84 | 61.44 | 3.771 | 6.280 | 34.955 | 14.275 | 12.415 | 13.942 | |
| O1 | 7/6/2015 | 146 | 181 | 6.51 | 61.52 | 3.396 | 5.612 | 45.683 | 29.564 | 10.743 | 12.581 | |
| 8. Ocoee 2 | | | | | | | | | | | | |
| O2 | 6/26/2014 | 460 | 282 | 7.35 | 43.81 | 1.299 | 6.355 | 58.387 | 12.964 | 10.223 | 5.641 | |
| O2 | 7/1/2014 | 248 | 150 | 7.17 | 43.75 | 1.416 | 6.489 | 50.470 | 13.627 | 11.120 | 5.372 | |
| O2 | 7/17/2014 | 53 | 256 | 7.20 | 38.01 | 0.741 | 6.342 | 55.559 | 12.490 | 10.479 | 1.746 | |
| O2 | 7/20/2014 | 37 | 264 | 7.17 | 42.60 | 1.373 | 6.556 | 47.355 | 12.631 | 10.588 | 5.095 | |
| O2 | 9/28/2014 | 182 | 248 | 7.38 | 36.45 | 1.020 | 5.637 | 47.070 | 11.969 | 9.782 | 4.721 | |
| O2 | 11/2/2014 | 216 | 224 | 7.10 | 34.76 | 1.001 | 5.688 | 42.763 | 11.560 | 9.341 | 3.734 | |
| O2 | 1/3/2015 | 40 | 174 | 7.70 | 35.77 | 1.068 | 5.516 | 49.380 | 10.951 | 9.590 | 7.137 | |
| O2 | 1/25/2015 | 39 | 140 | 7.40 | 35.20 | 1.024 | 5.671 | 46.295 | 11.709 | 9.612 | 5.574 | |
| O2 | 2/10/2015 | 38 | 170 | 7.47 | 36.35 | 1.090 | 5.716 | 43.616 | 11.723 | 9.798 | 5.172 | |
| O2 | 4/8/2015 | 110 | 233 | 4.53 | 44.72 | 1.525 | 6.372 | 53.614 | 12.087 | 10.146 | 5.677 | |
| O2 | 5/18/2015 | 162 | 281 | 5.47 | 40.82 | 1.176 | 6.282 | 51.982 | 14.359 | 11.127 | 5.369 | |
| O2 | 6/9/2015 | 47 | 309.3 | 5.45 | 42.15 | 1.502 | 6.750 | 55.284 | 12.888 | 8.836 | 6.329 | |
| O2 | 7/6/2015 | 72 | 189 | 7.18 | 46.20 | 1.886 | 6.188 | 47.320 | 13.117 | 10.344 | 9.261 | |

| Site | Date Collected | TSS (mg/L) | Conductivity (uS/cm) | pH units | Hardness (mg/L) | ANC (meq/L) | Acid Anions | | | Base Cations | | |
|------------------------|----------------|------------|----------------------|----------|-----------------|-------------|-------------|---------------|-------------|---------------|------------|--|
| | | | | | | | Cl- (mg/L) | SO4^2- (mg/L) | NO3- (mg/L) | PO4^3- (mg/L) | Na+ (mg/L) | |
| 9. Ocoee 3 | | | | | | | | | | | | |
| O3 | 7/1/2014 | 72 | 327 | 3.30 | 31.88 | 2.364 | 5.736 | 80.733 | 13.666 | 10.481 | 33.537 | |
| O3 | 7/17/2014 | 56 | 401 | 4.37 | 26.65 | 1.859 | 6.357 | 94.253 | 14.160 | 11.682 | 32.109 | |
| O3 | 7/20/2014 | 42 | 237 | 4.17 | 25.91 | 1.479 | 6.603 | 75.448 | 13.672 | 11.952 | 23.716 | |
| O3 | 9/28/2014 | 50 | 283 | 5.32 | 23.90 | 1.473 | 5.580 | 76.963 | 12.492 | 9.833 | 23.148 | |
| O3 | 11/2/2014 | 56 | 312 | 5.43 | 23.77 | 1.379 | 5.533 | 84.135 | 12.475 | 9.786 | 23.260 | |
| O3 | 1/3/2015 | 39 | 1008 | 3.00 | 41.19 | 2.501 | 9.112 | 134.589 | 24.302 | 14.898 | 42.824 | |
| O3 | 1/22/2015 | 39 | 1147 | 2.97 | 44.12 | 1.785 | 8.984 | 138.688 | 19.902 | 13.822 | 33.837 | |
| O3 | 1/25/2015 | 67 | 678 | 3.30 | 36.92 | 3.358 | 7.814 | 139.225 | 18.983 | 16.827 | 61.085 | |
| O3 | 2/10/2015 | 39 | 787 | 3.52 | 36.86 | 2.867 | 8.840 | 120.498 | 18.942 | 14.559 | 42.692 | |
| O3 | 4/8/2015 | 59 | 316 | 5.43 | 21.96 | -0.220 | 6.009 | 90.043 | 14.378 | 9.993 | 0.000 | |
| O3 | 5/18/2015 | 51 | 243 | 4.55 | 29.06 | 0.431 | 6.316 | 96.179 | 13.674 | 11.325 | 0.000 | |
| O3 | 6/9/2015 | 42 | 385 | 4.30 | 29.81 | 2.886 | 6.613 | 65.092 | 13.357 | 12.450 | 40.779 | |
| O3 | 7/6/2015 | 48 | 260 | 6.13 | 26.41 | 1.296 | 6.412 | 96.508 | 14.595 | 10.901 | 35.295 | |
| O3 | 8/8/2015 | 55 | 329 | 4.21 | 30.63 | 2.219 | 6.092 | 85.790 | 13.887 | 10.918 | 20.714 | |
| 10. Seiverville | | | | | | | | | | | | |
| S1 | 7/17/2014 | 60 | 429 | 6.63 | 80.43 | 3.190 | 6.074 | 122.287 | 12.872 | 9.836 | 17.256 | |
| S1 | 7/21/2014 | 46 | 433 | 6.03 | 79.63 | 3.010 | 6.410 | 117.202 | 13.871 | 11.745 | 13.576 | |
| S1 | 7/29/2014 | 185 | 225 | 6.07 | 79.28 | 3.238 | 5.994 | 116.526 | 10.940 | 11.045 | 17.202 | |
| S1 | 8/10/2014 | 82 | 257 | 5.90 | 77.24 | 3.007 | 6.600 | 114.249 | 10.332 | 10.998 | 13.984 | |
| S1 | 8/30/2014 | 98 | 287 | 6.23 | 72.15 | 2.882 | 5.726 | 103.272 | 9.970 | 9.481 | 12.786 | |
| S1 | 11/16/2014 | 106 | 233 | 6.40 | 72.35 | 2.869 | 5.700 | 104.425 | 11.232 | 9.620 | 13.337 | |
| S1 | 1/22/2015 | 43 | 1104 | 4.20 | 111.41 | 3.269 | 10.376 | 221.674 | 14.055 | 14.051 | 24.239 | |
| S1 | 1/29/2015 | 38 | 1089 | 4.37 | 121.42 | 3.683 | 10.537 | 213.948 | 18.431 | 13.270 | 16.710 | |
| S1 | 4/16/2015 | 92 | 469 | 6.32 | 82.12 | 2.289 | 6.494 | 123.745 | 11.010 | 10.892 | 0.000 | |
| S1 | 5/18/2015 | 89 | 306 | 6.25 | 80.99 | 2.144 | 6.417 | 126.632 | 14.427 | 10.462 | 0.000 | |
| S1 | 6/9/2015 | 58 | 374 | 6.14 | 84.79 | 3.616 | 6.763 | 108.030 | 14.468 | 12.254 | 22.281 | |
| S1 | 7/6/2015 | 95 | 339 | 6.06 | 79.85 | 3.021 | 6.332 | 122.643 | 11.722 | 10.655 | 14.660 | |

Dissolved Metals

| Site | Date Collected | K + (mg/L) | Mg ²⁺ (mg/L) | Ca ²⁺ (mg/L) | Al (mg/L) | Cu (mg/L) | Fe (mg/L) | Mn (mg/L) | Si (mg/L) | Zn (mg/L) | Ba (mg/L) | Cd (mg/L) |
|-----------------------|----------------|---------------|----------------------------|----------------------------|--------------|--------------|--------------|--------------|--------------|--------------|--------------|--------------|
| 1. Grainger | | | | | | | | | | | | |
| G1 | 7/21/2014 | 1.652 | 10.204 | 8.578 | 0.920 | 0.071 | 0.245 | 0.736 | 5.525 | 0.470 | 0.118 | 0.020 |
| G1 | 10/10/2014 | 1.869 | 9.428 | 9.759 | 0.823 | 0.072 | 0.167 | 0.800 | 5.886 | 0.469 | 0.129 | 0.025 |
| G1 | 10/23/2014 | 1.935 | 9.048 | 8.943 | 0.969 | 0.064 | 0.205 | 0.801 | 5.694 | 0.476 | 0.131 | 0.017 |
| G1 | 1/29/2015 | 2.044 | 8.133 | 8.170 | 0.950 | 0.070 | 0.224 | 0.712 | 5.899 | 0.502 | 0.117 | 0.017 |
| G1 | 4/16/2015 | 1.582 | 7.997 | 7.681 | 0.273 | 0.056 | 0.050 | 0.410 | 5.477 | 0.194 | 0.099 | 0.005 |
| G1 | 5/18/2015 | 1.555 | 7.754 | 7.306 | 0.248 | 0.037 | 0.048 | 0.396 | 5.500 | 0.187 | 0.090 | 0.008 |
| G1 | 6/9/2015 | 2.012 | 11.824 | 11.432 | 0.699 | 0.063 | 0.126 | 0.641 | 7.143 | 0.301 | 0.138 | 0.026 |
| G1 | 7/6/2015 | 1.815 | 9.053 | 8.604 | 0.877 | 0.073 | 0.205 | 0.747 | 5.509 | 0.464 | 0.119 | 0.021 |
| 2. Jamestown 1 | | | | | | | | | | | | |
| J1 | 7/22/2014 | 6.344 | 2.942 | 11.829 | 0.416 | 0.068 | 0.195 | 0.327 | 0.575 | 0.274 | 0.069 | 0.016 |
| J1 | 7/31/2014 | 6.425 | 2.513 | 9.819 | 0.413 | 0.067 | 0.215 | 0.329 | 0.561 | 0.269 | 0.064 | 0.003 |
| J1 | 9/15/2014 | 6.537 | 2.840 | 10.979 | 0.307 | 0.055 | 0.097 | 0.334 | 0.580 | 0.272 | 0.065 | 0.005 |
| J1 | 12/10/2014 | 6.110 | 3.133 | 12.509 | 0.291 | 0.064 | 0.120 | 0.338 | 0.546 | 0.288 | 0.067 | 0.007 |
| J1 | 1/2/2015 | 6.374 | 3.212 | 13.727 | 0.284 | 0.054 | 0.117 | 0.331 | 0.535 | 0.291 | 0.067 | 0.017 |
| J1 | 1/26/2015 | 6.806 | 3.031 | 10.256 | 0.284 | 0.067 | 0.097 | 0.341 | 0.608 | 0.271 | 0.063 | 0.013 |
| J1 | 4/16/2015 | 5.759 | 2.324 | 10.360 | 0.000 | 0.044 | 0.047 | 0.003 | 0.829 | 0.016 | 0.053 | 0.016 |
| J1 | 6/20/2015 | 6.398 | 4.238 | 17.472 | 0.557 | 0.041 | 0.058 | 0.026 | 1.071 | 0.077 | 0.055 | 0.014 |
| J1 | 7/24/2015 | 7.083 | 2.362 | 10.524 | 0.440 | 0.064 | 0.218 | 0.332 | 0.557 | 0.265 | 0.067 | 0.003 |
| 3. Jamestown 2 | | | | | | | | | | | | |
| J2 | 7/15/2014 | 3.252 | 15.047 | 27.440 | 0.751 | 0.161 | 1.856 | 2.256 | 7.508 | 0.128 | 0.289 | 0.064 |
| J2 | 9/15/2014 | 2.950 | 11.136 | 15.677 | 0.450 | 0.194 | 0.100 | 1.505 | 5.387 | 0.105 | 0.179 | 0.065 |
| J2 | 12/10/2014 | 2.579 | 14.455 | 14.022 | 0.323 | 0.161 | 0.087 | 2.354 | 4.696 | 0.074 | 0.212 | 0.087 |
| J2 | 1/2/2015 | 2.748 | 13.999 | 13.106 | 0.438 | 0.154 | 0.050 | 2.177 | 3.819 | 0.114 | 0.214 | 0.074 |
| J2 | 1/26/2015 | 1.581 | 11.269 | 23.093 | 0.285 | 0.195 | 0.081 | 2.031 | 6.510 | 0.138 | 0.195 | 0.064 |
| J2 | 4/16/2015 | 3.318 | 14.600 | 26.678 | 0.626 | 0.171 | 0.117 | 2.315 | 7.678 | 0.118 | 0.250 | 0.047 |
| J2 | 6/20/2015 | 3.056 | 15.294 | 27.532 | 1.380 | 0.070 | 2.648 | 2.086 | 6.716 | 0.117 | 0.249 | 0.076 |
| 4. Jamestown 3 | | | | | | | | | | | | |
| J3 | 7/4/2014 | 1.455 | 2.131 | 8.102 | 0.130 | 0.020 | 0.047 | 0.087 | 2.439 | 0.080 | 0.057 | 0.010 |
| J3 | 7/15/2014 | 1.580 | 1.999 | 7.768 | 0.131 | 0.020 | 0.044 | 0.087 | 2.678 | 0.084 | 0.057 | 0.010 |

Dissolved Metals

| Site | Date Collected | K + (mg/L) | Mg ²⁺ (mg/L) | Ca ²⁺ (mg/L) | Al (mg/L) | Cu (mg/L) | Fe (mg/L) | Mn (mg/L) | Si (mg/L) | Zn (mg/L) | Ba (mg/L) | Cd (mg/L) |
|-----------------------|----------------|------------|-------------------------|-------------------------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|
| 4. Jamestown 3 | | | | | | | | | | | | |
| J3 | 7/22/2014 | 1.555 | 2.155 | 7.367 | 0.141 | 0.020 | 0.054 | 0.081 | 2.512 | 0.091 | 0.057 | 0.010 |
| J3 | 7/31/2014 | 1.476 | 1.981 | 7.217 | 0.124 | 0.020 | 0.054 | 0.087 | 2.760 | 0.087 | 0.050 | 0.010 |
| J3 | 9/15/2014 | 1.381 | 1.851 | 7.250 | 0.121 | 0.020 | 0.042 | 0.080 | 2.317 | 0.080 | 0.050 | 0.010 |
| J3 | 12/10/2014 | 1.382 | 1.846 | 7.645 | 0.121 | 0.020 | 0.045 | 0.081 | 2.341 | 0.081 | 0.050 | 0.010 |
| J3 | 1/2/2015 | 1.375 | 1.847 | 7.537 | 0.121 | 0.020 | 0.047 | 0.080 | 2.297 | 0.080 | 0.050 | 0.010 |
| J3 | 1/22/2015 | 1.383 | 1.862 | 7.195 | 0.121 | 0.020 | 0.044 | 0.080 | 2.298 | 0.080 | 0.050 | 0.010 |
| J3 | 1/26/2015 | 1.394 | 1.847 | 7.655 | 0.121 | 0.020 | 0.044 | 0.081 | 2.351 | 0.081 | 0.050 | 0.010 |
| J3 | 4/16/2015 | 1.537 | 1.846 | 8.163 | 0.018 | 0.011 | 0.026 | 0.004 | 2.563 | 0.031 | 0.052 | 0.003 |
| J3 | 6/20/2015 | 1.583 | 2.191 | 9.049 | 0.082 | 0.012 | 0.053 | 0.005 | 2.875 | 0.029 | 0.048 | 0.010 |
| J3 | 7/24/2015 | 1.506 | 2.051 | 8.214 | 0.131 | 0.020 | 0.042 | 0.084 | 2.545 | 0.087 | 0.053 | 0.005 |
| 5. Nashville 1 | | | | | | | | | | | | |
| N1 | 6/8/2014 | 0.726 | 154.407 | 462.806 | 3.250 | 0.555 | 0.592 | 2.611 | 7.294 | 2.384 | 0.492 | 0.107 |
| N1 | 7/1/2014 | 0.704 | 153.742 | 466.763 | 3.396 | 0.522 | 0.608 | 2.625 | 7.215 | 2.232 | 0.569 | 0.144 |
| N1 | 7/22/2014 | 0.659 | 153.583 | 461.681 | 3.318 | 0.518 | 0.593 | 2.598 | 7.356 | 2.286 | 0.557 | 0.096 |
| N1 | 8/26/2014 | 0.703 | 151.540 | 452.646 | 2.499 | 0.568 | 0.513 | 2.622 | 7.428 | 2.337 | 0.563 | 0.181 |
| N1 | 10/11/2014 | 0.702 | 165.100 | 478.937 | 2.532 | 0.548 | 0.449 | 2.537 | 7.202 | 2.203 | 0.475 | 0.000 |
| N1 | 1/9/2015 | 0.703 | 147.154 | 468.586 | 2.416 | 0.633 | 0.522 | 2.617 | 7.373 | 2.300 | 0.522 | 0.146 |
| N1 | 1/30/2015 | 0.703 | 159.702 | 464.522 | 2.355 | 0.534 | 0.468 | 2.653 | 7.228 | 2.315 | 0.544 | 0.127 |
| N1 | 3/23/2015 | 0.593 | 179.179 | 619.533 | 0.036 | 0.243 | 0.114 | 1.062 | 11.704 | 0.718 | 0.340 | 0.096 |
| N1 | 6/1/2015 | 0.585 | 198.829 | 627.718 | 0.403 | 0.215 | 0.081 | 0.959 | 9.526 | 0.263 | 0.365 | 0.143 |
| N1 | 7/24/2015 | 0.689 | 161.323 | 448.727 | 3.202 | 0.570 | 0.620 | 2.614 | 7.415 | 2.239 | 0.535 | 0.090 |
| N1 | 8/21/2015 | 0.728 | 150.706 | 465.383 | 3.119 | 0.602 | 0.552 | 2.598 | 6.867 | 2.151 | 0.510 | 0.147 |
| 6. Nashville 2 | | | | | | | | | | | | |
| N2 | 7/1/2014 | 2.069 | 146.711 | 513.416 | 48.616 | 1.090 | 2.518 | 5.685 | 20.624 | 9.031 | 0.476 | 0.137 |
| N2 | 8/26/2014 | 1.923 | 144.556 | 537.467 | 41.493 | 1.070 | 2.181 | 6.091 | 21.979 | 8.919 | 0.503 | 0.116 |
| N2 | 10/11/2014 | 1.860 | 130.541 | 530.590 | 40.622 | 0.955 | 3.452 | 6.067 | 21.171 | 8.802 | 0.523 | 0.164 |
| N2 | 1/9/2015 | 6.411 | 42.774 | 495.750 | 44.505 | 1.093 | 1.523 | 5.773 | 22.168 | 8.998 | 0.569 | 0.232 |
| N2 | 3/23/2015 | 1.745 | 127.750 | 622.817 | 0.015 | 0.261 | 0.104 | 1.223 | 6.492 | 0.158 | 0.318 | 0.089 |
| N2 | 6/1/2015 | 1.930 | 153.450 | 582.705 | 40.357 | 0.966 | 0.178 | 3.666 | 22.715 | 7.411 | 0.346 | 0.075 |
| N2 | 7/24/2015 | 1.917 | 138.741 | 519.682 | 50.029 | 1.065 | 0.755 | 5.749 | 20.556 | 8.511 | 0.470 | 0.137 |
| N2 | 8/21/2015 | 1.960 | 137.758 | 527.829 | 46.393 | 1.027 | 2.305 | 5.822 | 21.589 | 9.027 | 0.487 | 0.074 |

Dissolved Metals

| Site | Date Collected | K + (mg/L) | Mg ²⁺ (mg/L) | Ca ²⁺ (mg/L) | Al (mg/L) | Cu (mg/L) | Fe (mg/L) | Mn (mg/L) | Si (mg/L) | Zn (mg/L) | Ba (mg/L) | Cd (mg/L) |
|-------------------|----------------|------------|-------------------------|-------------------------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|
| 7. Ocoee 1 | | | | | | | | | | | | |
| O1 | 6/12/2014 | 0.321 | 14.944 | 47.221 | 0.321 | 0.044 | 0.057 | 0.020 | 8.553 | 0.050 | 0.084 | 0.013 |
| O1 | 6/19/2014 | 0.392 | 18.561 | 40.752 | 0.362 | 0.050 | 0.030 | 0.020 | 7.501 | 0.050 | 0.090 | 0.010 |
| O1 | 6/26/2014 | 0.335 | 17.671 | 50.143 | 0.305 | 0.044 | 0.057 | 0.020 | 8.149 | 0.050 | 0.084 | 0.013 |
| O1 | 7/1/2014 | 0.350 | 17.959 | 44.377 | 0.306 | 0.047 | 0.047 | 0.020 | 8.341 | 0.050 | 0.087 | 0.015 |
| O1 | 7/17/2014 | 0.315 | 16.544 | 47.815 | 0.328 | 0.044 | 0.050 | 0.020 | 7.910 | 0.050 | 0.084 | 0.013 |
| O1 | 7/20/2014 | 0.332 | 16.084 | 50.163 | 0.342 | 0.044 | 0.040 | 0.020 | 7.997 | 0.050 | 0.087 | 0.013 |
| O1 | 9/28/2014 | 0.290 | 15.430 | 42.205 | 0.278 | 0.040 | 0.045 | 0.020 | 7.397 | 0.050 | 0.080 | 0.010 |
| O1 | 11/2/2014 | 0.298 | 16.128 | 41.300 | 0.288 | 0.040 | 0.054 | 0.020 | 7.284 | 0.050 | 0.080 | 0.010 |
| O1 | 1/3/2015 | 0.298 | 15.972 | 40.906 | 0.275 | 0.040 | 0.040 | 0.020 | 7.607 | 0.050 | 0.080 | 0.010 |
| O1 | 1/25/2015 | 0.297 | 15.668 | 41.914 | 0.277 | 0.040 | 0.053 | 0.020 | 7.011 | 0.050 | 0.080 | 0.010 |
| O1 | 2/10/2015 | 0.305 | 14.588 | 39.834 | 0.291 | 0.040 | 0.033 | 0.020 | 7.097 | 0.050 | 0.080 | 0.010 |
| O1 | 4/8/2015 | 0.234 | 12.792 | 35.182 | 0.010 | 0.050 | 0.029 | 0.005 | 4.742 | 0.018 | 0.074 | 0.011 |
| O1 | 5/18/2015 | 0.250 | 13.148 | 34.932 | 0.009 | 0.056 | 0.016 | 0.002 | 5.234 | 0.043 | 0.073 | 0.008 |
| O1 | 6/9/2015 | 0.381 | 16.580 | 44.845 | 0.240 | 0.034 | 0.014 | 0.002 | 7.338 | 0.052 | 0.066 | 0.020 |
| O1 | 7/6/2015 | 0.328 | 16.456 | 45.018 | 0.309 | 0.048 | 0.047 | 0.019 | 7.864 | 0.052 | 0.083 | 0.014 |
| 8. Ocoee 2 | | | | | | | | | | | | |
| O2 | 6/26/2014 | 1.265 | 10.501 | 33.200 | 0.459 | 0.108 | 0.106 | 0.151 | 2.465 | 0.252 | 0.093 | 0.020 |
| O2 | 7/1/2014 | 1.209 | 9.566 | 34.082 | 0.293 | 0.097 | 0.107 | 0.197 | 2.888 | 0.179 | 0.093 | 0.035 |
| O2 | 7/17/2014 | 0.976 | 7.954 | 30.015 | 0.000 | 0.042 | 0.037 | 0.059 | 2.433 | 0.024 | 0.078 | 0.011 |
| O2 | 7/20/2014 | 1.177 | 9.189 | 33.384 | 0.081 | 0.047 | 0.023 | 0.002 | 2.747 | 0.032 | 0.087 | 0.009 |
| O2 | 9/28/2014 | 0.950 | 7.768 | 28.628 | 0.254 | 0.053 | 0.053 | 0.234 | 2.214 | 0.143 | 0.071 | 0.010 |
| O2 | 11/2/2014 | 0.935 | 7.521 | 27.180 | 0.366 | 0.054 | 0.060 | 0.235 | 2.175 | 0.164 | 0.070 | 0.010 |
| O2 | 1/3/2015 | 0.951 | 8.298 | 27.408 | 0.120 | 0.070 | 0.060 | 0.020 | 2.242 | 0.160 | 0.070 | 0.000 |
| O2 | 1/25/2015 | 0.953 | 8.173 | 26.959 | 0.168 | 0.060 | 0.067 | 0.315 | 2.190 | 0.191 | 0.070 | 0.017 |
| O2 | 2/10/2015 | 0.887 | 7.861 | 28.435 | 0.110 | 0.047 | 0.050 | 0.171 | 2.189 | 0.231 | 0.070 | 0.017 |
| O2 | 4/8/2015 | 1.381 | 11.828 | 32.773 | 0.302 | 0.060 | 0.117 | 0.342 | 2.679 | 0.238 | 0.091 | 0.023 |
| O2 | 5/18/2015 | 1.105 | 9.524 | 31.170 | 0.320 | 0.054 | 0.128 | 0.161 | 2.500 | 0.299 | 0.094 | 0.034 |
| O2 | 6/9/2015 | 1.023 | 12.477 | 29.547 | 0.020 | 0.070 | 0.130 | 0.381 | 3.500 | 0.050 | 0.100 | 0.010 |
| O2 | 7/6/2015 | 1.318 | 10.407 | 35.664 | 0.612 | 0.121 | 0.133 | 0.487 | 2.795 | 0.435 | 0.110 | 0.046 |

Dissolved Metals

| Site | Date Collected | K + (mg/L) | Mg ²⁺ (mg/L) | Ca ²⁺ (mg/L) | Al (mg/L) | Cu (mg/L) | Fe (mg/L) | Mn (mg/L) | Si (mg/L) | Zn (mg/L) | Ba (mg/L) | Cd (mg/L) |
|------------------------|----------------|---------------|----------------------------|----------------------------|--------------|--------------|--------------|--------------|--------------|--------------|--------------|--------------|
| 9. Ocoee 3 | | | | | | | | | | | | |
| O3 | 7/1/2014 | 2.256 | 2.722 | 28.450 | 5.138 | 0.907 | 0.703 | 3.546 | 5.243 | 2.732 | 0.429 | 0.151 |
| O3 | 7/17/2014 | 2.509 | 2.697 | 23.303 | 6.251 | 0.626 | 0.653 | 3.188 | 5.649 | 2.790 | 0.445 | 0.081 |
| O3 | 7/20/2014 | 1.866 | 3.050 | 22.114 | 5.013 | 0.627 | 0.742 | 3.318 | 4.085 | 2.581 | 0.453 | 0.091 |
| O3 | 9/28/2014 | 1.477 | 2.578 | 20.713 | 4.512 | 0.597 | 0.613 | 2.895 | 5.083 | 2.385 | 0.367 | 0.064 |
| O3 | 11/2/2014 | 2.046 | 2.730 | 20.451 | 5.366 | 0.491 | 0.587 | 2.833 | 4.697 | 2.366 | 0.377 | 0.050 |
| O3 | 1/3/2015 | 2.900 | 4.946 | 35.132 | 8.905 | 0.813 | 1.115 | 3.175 | 7.296 | 3.269 | 0.517 | 0.296 |
| O3 | 1/22/2015 | 3.447 | 4.408 | 38.705 | 6.947 | 0.959 | 1.008 | 3.955 | 4.922 | 4.046 | 0.678 | 0.321 |
| O3 | 1/25/2015 | 3.620 | 4.347 | 31.565 | 6.537 | 1.139 | 1.004 | 4.661 | 10.728 | 3.819 | 0.540 | 0.273 |
| O3 | 2/10/2015 | 2.832 | 5.676 | 30.137 | 9.016 | 0.939 | 1.043 | 4.792 | 7.781 | 3.643 | 0.801 | 0.278 |
| O3 | 4/8/2015 | 0.954 | 1.448 | 19.772 | 2.877 | 0.252 | 0.741 | 0.690 | 5.905 | 0.339 | 0.348 | 0.178 |
| O3 | 5/18/2015 | 1.073 | 1.889 | 26.962 | 3.386 | 0.248 | 0.204 | 0.728 | 8.537 | 0.471 | 0.389 | 0.085 |
| O3 | 6/9/2015 | 1.068 | 1.858 | 27.804 | 4.610 | 0.355 | 0.145 | 0.732 | 7.809 | 0.500 | 0.337 | 0.000 |
| O3 | 7/6/2015 | 2.930 | 3.314 | 22.462 | 3.496 | 0.687 | 0.633 | 3.064 | 2.988 | 2.647 | 0.433 | 0.181 |
| O3 | 8/8/2015 | 1.073 | 2.596 | 27.279 | 7.208 | 0.579 | 0.753 | 3.277 | 8.316 | 2.534 | 0.454 | 0.115 |
| 10. Seiverville | | | | | | | | | | | | |
| S1 | 7/17/2014 | 2.893 | 25.307 | 54.725 | 2.289 | 0.288 | 0.399 | 1.830 | 2.698 | 1.512 | 0.352 | 0.134 |
| S1 | 7/21/2014 | 2.880 | 25.964 | 53.263 | 2.308 | 0.325 | 0.405 | 1.812 | 2.522 | 1.534 | 0.375 | 0.151 |
| S1 | 7/29/2014 | 2.854 | 25.355 | 53.523 | 2.294 | 0.292 | 0.403 | 1.880 | 2.619 | 1.548 | 0.339 | 0.154 |
| S1 | 8/10/2014 | 2.890 | 25.720 | 51.132 | 2.189 | 0.307 | 0.391 | 1.855 | 2.419 | 1.467 | 0.348 | 0.090 |
| S1 | 8/30/2014 | 2.595 | 23.377 | 48.436 | 1.971 | 0.295 | 0.335 | 1.716 | 2.353 | 1.348 | 0.295 | 0.111 |
| S1 | 11/16/2014 | 2.600 | 23.410 | 48.613 | 1.967 | 0.285 | 0.322 | 1.709 | 2.362 | 1.347 | 0.298 | 0.074 |
| S1 | 1/22/2015 | 3.652 | 34.767 | 76.123 | 2.278 | 0.578 | 0.524 | 2.030 | 4.186 | 1.983 | 0.544 | 0.235 |
| S1 | 1/29/2015 | 4.348 | 39.209 | 81.579 | 2.966 | 0.527 | 0.627 | 2.956 | 3.198 | 2.106 | 0.567 | 0.275 |
| S1 | 4/16/2015 | 2.857 | 26.018 | 56.013 | 0.000 | 0.145 | 0.088 | 0.349 | 3.667 | 0.245 | 0.248 | 0.050 |
| S1 | 5/18/2015 | 2.863 | 25.929 | 55.015 | 0.000 | 0.178 | 0.046 | 0.355 | 3.749 | 0.231 | 0.245 | 0.080 |
| S1 | 6/9/2015 | 2.761 | 26.773 | 57.941 | 0.839 | 0.174 | 0.076 | 0.367 | 2.892 | 0.258 | 0.208 | 0.019 |
| S1 | 7/6/2015 | 2.870 | 25.867 | 53.615 | 2.190 | 0.323 | 0.369 | 1.893 | 2.593 | 1.480 | 0.325 | 0.129 |

Total Ions/ Metals

| Site | Date Collected | Ni (mg/L) | Mg ²⁺ (mg/L) | Ca ²⁺ (mg/L) | Al (mg/L) | Cu (mg/L) | Fe (mg/L) | Mn (mg/L) | Si (mg/L) | Zn (mg/L) | Cd (mg/L) | Ni (mg/L) |
|-----------------------|----------------|-----------|-------------------------|-------------------------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|
| | | | | | | | | | | | | |
| 1. Grainger | | | | | | | | | | | | |
| G1 | 7/21/2014 | 0.104 | 11.841 | 10.446 | 8.266 | 0.026 | 9.906 | 0.867 | 8.543 | 0.185 | 0.011 | 0.122 |
| G1 | 10/10/2014 | 0.099 | | | | | | | | | | |
| G1 | 10/23/2014 | 0.087 | | | | | | | | | | |
| G1 | 1/29/2015 | 0.097 | 7.954 | 8.237 | 9.190 | 0.034 | 11.719 | 0.640 | 9.560 | 0.144 | 0.003 | 0.063 |
| G1 | 4/16/2015 | 0.086 | | | | | | | | | | |
| G1 | 5/18/2015 | 0.069 | 7.872 | 6.753 | 7.695 | 0.039 | 10.287 | 0.576 | 9.147 | 0.216 | 0.000 | 0.085 |
| G1 | 6/9/2015 | 0.098 | | | | | | | | | | |
| G1 | 7/6/2015 | 0.105 | 7.033 | 5.781 | 10.481 | 0.055 | 14.064 | 0.744 | 10.167 | 0.222 | 0.010 | 0.066 |
| 2. Jamestown 1 | | | | | | | | | | | | |
| J1 | 7/22/2014 | 0.020 | 1.728 | 8.429 | 0.947 | 0.004 | 0.624 | 0.069 | 1.816 | 0.029 | 0.002 | 0.009 |
| J1 | 7/31/2014 | 0.017 | | | | | | | | | | |
| J1 | 9/15/2014 | 0.020 | | | | | | | | | | |
| J1 | 12/10/2014 | 0.020 | | | | | | | | | | |
| J1 | 1/2/2015 | 0.020 | | | | | | | | | | |
| J1 | 1/26/2015 | 0.013 | 2.723 | 14.173 | 1.248 | 0.000 | 1.157 | 0.121 | 2.635 | 0.188 | 0.001 | 0.061 |
| J1 | 4/16/2015 | 0.009 | 2.765 | 11.633 | 2.306 | 0.000 | 2.035 | 0.144 | 4.753 | 0.287 | 0.435 | 0.247 |
| J1 | 6/20/2015 | 0.021 | | | | | | | | | | |
| J1 | 7/24/2015 | 0.030 | 2.958 | 12.158 | 2.216 | 0.000 | 2.684 | 0.168 | 4.641 | 0.342 | 0.091 | 0.301 |
| 3. Jamestown 2 | | | | | | | | | | | | |
| J2 | 7/15/2014 | 0.168 | | | | | | | | | | |
| J2 | 9/15/2014 | 0.119 | | | | | | | | | | |
| J2 | 12/10/2014 | 0.091 | | | | | | | | | | |
| J2 | 1/2/2015 | 0.104 | 18.410 | 37.360 | 0.905 | 0.000 | 0.907 | 0.230 | 2.470 | 0.049 | 0.005 | 0.023 |
| J2 | 1/26/2015 | 0.124 | 0.586 | 1.667 | 0.321 | 0.000 | 0.350 | 0.024 | 0.555 | 0.010 | 0.000 | 0.003 |
| J2 | 4/16/2015 | 0.065 | 14.013 | 26.403 | 2.008 | 0.000 | 5.763 | 2.288 | 6.712 | 0.102 | 0.021 | 0.076 |
| J2 | 6/20/2015 | 0.208 | 13.796 | 26.453 | 2.392 | 0.004 | 5.710 | 2.276 | 6.929 | 0.114 | 0.063 | 0.038 |
| 4. Jamestown 3 | | | | | | | | | | | | |
| J3 | 7/4/2014 | 0.000 | | | | | | | | | | |
| J3 | 7/15/2014 | 0.000 | 2.206 | 8.907 | 0.485 | 0.019 | 0.472 | 0.215 | 1.536 | 0.079 | 0.002 | 0.005 |

Total Ions/ Metals

| Site | Date Collected | Ni (mg/L) | Mg ²⁺ (mg/L) | Ca ²⁺ (mg/L) | Al (mg/L) | Cu (mg/L) | Fe (mg/L) | Mn (mg/L) | Si (mg/L) | Zn (mg/L) | Cd (mg/L) | Ni (mg/L) |
|-------------------|----------------|-----------|-------------------------|-------------------------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|
| | | | | | | | | | | | | |
| 7. Ocoee 1 | | | | | | | | | | | | |
| O1 | 6/12/2014 | 0.003 | | | | | | | | | | |
| O1 | 6/19/2014 | 0.000 | | | | | | | | | | |
| O1 | 6/26/2014 | 0.007 | | | | | | | | | | |
| O1 | 7/1/2014 | 0.002 | | | | | | | | | | |
| O1 | 7/17/2014 | 0.003 | | | | | | | | | | |
| O1 | 7/20/2014 | 0.003 | | | | | | | | | | |
| O1 | 9/28/2014 | 0.003 | | | | | | | | | | |
| O1 | 11/2/2014 | 0.000 | | | | | | | | | | |
| O1 | 1/3/2015 | 0.000 | | | | | | | | | | |
| O1 | 1/25/2015 | 0.003 | 7.405 | 20.875 | 0.516 | 0.001 | 0.096 | 0.018 | 5.374 | 0.034 | 0.001 | 0.000 |
| O1 | 2/10/2015 | 0.000 | 7.269 | 20.445 | 0.318 | 0.003 | 0.100 | 0.024 | 5.352 | 0.039 | 0.001 | 0.000 |
| O1 | 4/8/2015 | 0.000 | 13.560 | 36.896 | 2.304 | 0.000 | 3.582 | 0.288 | 10.312 | 0.451 | 0.183 | 0.037 |
| O1 | 5/18/2015 | 0.004 | 13.616 | 38.047 | 0.478 | 0.009 | 2.022 | 0.151 | 9.840 | 0.111 | 0.001 | 0.000 |
| O1 | 6/9/2015 | 0.000 | | | | | | | | | | |
| O1 | 7/6/2015 | 0.006 | 13.043 | 35.025 | 2.142 | 0.000 | 3.233 | 0.268 | 12.658 | 0.349 | 0.276 | 0.236 |
| 8. Ocoee 2 | | | | | | | | | | | | |
| O2 | 6/26/2014 | 0.005 | 9.671 | 25.595 | 0.800 | 0.000 | 0.719 | 0.138 | 4.432 | 0.051 | 0.000 | 0.015 |
| O2 | 7/1/2014 | 0.005 | 3.614 | 10.410 | 0.872 | 0.000 | 0.523 | 0.075 | 4.739 | 0.029 | 0.005 | 0.000 |
| O2 | 7/17/2014 | 0.017 | 8.324 | 22.439 | 0.750 | 0.000 | 0.286 | 0.043 | 5.709 | 0.050 | 0.000 | 0.045 |
| O2 | 7/20/2014 | 0.009 | | | | | | | | | | |
| O2 | 9/28/2014 | 0.003 | | | | | | | | | | |
| O2 | 11/2/2014 | 0.013 | | | | | | | | | | |
| O2 | 1/3/2015 | 0.010 | | | | | | | | | | |
| O2 | 1/25/2015 | 0.003 | | | | | | | | | | |
| O2 | 2/10/2015 | 0.000 | 8.128 | 25.817 | 0.921 | 0.000 | 0.565 | 0.181 | 4.976 | 0.069 | 0.000 | 0.004 |
| O2 | 4/8/2015 | 0.013 | | | | | | | | | | |
| O2 | 5/18/2015 | 0.003 | | | | | | | | | | |
| O2 | 6/9/2015 | 0.020 | | | | | | | | | | |
| O2 | 7/6/2015 | 0.017 | 5.561 | 22.194 | 2.034 | 0.018 | 5.321 | 0.592 | 3.271 | 0.170 | 0.002 | 0.014 |

Total Ions/Metals

| Site | Date Collected | Ni (mg/L) | Mg ²⁺ (mg/L) | Ca ²⁺ (mg/L) | Al (mg/L) | Cu (mg/L) | Fe (mg/L) | Mn (mg/L) | Si (mg/L) | Zn (mg/L) | Cd (mg/L) | Ni (mg/L) |
|------------------------|----------------|-----------|-------------------------|-------------------------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|
| | | | | | | | | | | | | |
| 9. Ocoee 3 | | | | | | | | | | | | |
| O3 | 7/1/2014 | 0.107 | 1.038 | 14.497 | 1.985 | 0.019 | 4.397 | 0.492 | 3.601 | 0.239 | 0.034 | 0.029 |
| O3 | 7/17/2014 | 0.010 | | | | | | | | | | |
| O3 | 7/20/2014 | 0.174 | 0.963 | 10.354 | 1.648 | 0.000 | 2.774 | 0.526 | 2.003 | 0.317 | 0.000 | 0.032 |
| O3 | 9/28/2014 | 0.072 | | | | | | | | | | |
| O3 | 11/2/2014 | 0.090 | | | | | | | | | | |
| O3 | 1/3/2015 | 0.044 | | | | | | | | | | |
| O3 | 1/22/2015 | 0.149 | 6.005 | 72.580 | 22.755 | 0.000 | 20.054 | 3.497 | 21.294 | 1.433 | 0.000 | 0.057 |
| O3 | 1/25/2015 | 0.200 | | | | | | | | | | |
| O3 | 2/10/2015 | 0.291 | 4.030 | 60.318 | 12.723 | 0.000 | 6.821 | 2.369 | 14.493 | 0.889 | 0.010 | 0.081 |
| O3 | 4/8/2015 | 0.069 | | | | | | | | | | |
| O3 | 5/18/2015 | 0.137 | | | | | | | | | | |
| O3 | 6/9/2015 | 0.189 | | | | | | | | | | |
| O3 | 7/6/2015 | 0.026 | | | | | | | | | | |
| O3 | 8/8/2015 | 0.142 | 1.725 | 25.434 | 3.800 | 0.000 | 2.618 | 0.709 | 5.784 | 0.394 | 0.004 | 0.045 |
| 10. Seiverville | | | | | | | | | | | | |
| S1 | 7/17/2014 | 0.027 | | | | | | | | | | |
| S1 | 7/21/2014 | 0.060 | 19.849 | 40.867 | 0.635 | 0.006 | 1.441 | 0.066 | 3.907 | 0.120 | 0.000 | 0.006 |
| S1 | 7/29/2014 | 0.024 | | | | | | | | | | |
| S1 | 8/10/2014 | 0.033 | 8.084 | 16.859 | 0.526 | 0.004 | 2.107 | 0.157 | 1.980 | 0.096 | 0.022 | 0.043 |
| S1 | 8/30/2014 | 0.054 | | | | | | | | | | |
| S1 | 11/16/2014 | 0.060 | | | | | | | | | | |
| S1 | 1/22/2015 | 0.084 | | | | | | | | | | |
| S1 | 1/29/2015 | 0.037 | 66.069 | 145.245 | 2.063 | 0.000 | 5.843 | 0.636 | 5.171 | 0.319 | 0.000 | 0.015 |
| S1 | 4/16/2015 | 0.067 | 24.547 | 54.314 | 1.404 | 0.023 | 5.911 | 0.384 | 3.785 | 0.290 | 0.015 | 0.038 |
| S1 | 5/18/2015 | 0.053 | | | | | | | | | | |
| S1 | 6/9/2015 | 0.010 | | | | | | | | | | |
| S1 | 7/6/2015 | 0.063 | 24.418 | 54.002 | 1.001 | 0.000 | 5.649 | 0.412 | 3.478 | 0.287 | 0.030 | 0.068 |

| Site | Date Collected | Acid Anions | | | | | Base Cations | | | | | Dissolved Metals | | |
|-----------------------|----------------|-------------|---------------------------|--------------|---------------------------|----------------|--------------|------------|--------------------------|--------------------------|----------------|--------------------------|--|--|
| | | Cl- (meq/L) | SO4 ²⁻ (meq/L) | NO3- (meq/L) | PO4 ³⁻ (meq/L) | Sum AA (meq/L) | Na+ (meq/L) | K+ (meq/L) | Mg ²⁺ (meq/L) | Ca ²⁺ (meq/L) | Sum BC (meq/L) | Al ³⁺ (meq/L) | | |
| 1. Grainger | | | | | | | | | | | | | | |
| G1 | 7/21/2014 | 0.169 | 1.937 | 0.224 | 0.406 | 2.736 | 0.239 | 0.042 | 0.840 | 0.429 | 1.550 | 0.102 | | |
| G1 | 10/10/2014 | 0.175 | 1.753 | 0.236 | 0.376 | 2.540 | 0.258 | 0.048 | 0.776 | 0.488 | 1.570 | 0.092 | | |
| G1 | 10/23/2014 | 0.167 | 1.708 | 0.248 | 0.394 | 2.517 | 0.261 | 0.050 | 0.745 | 0.447 | 1.502 | 0.108 | | |
| G1 | 1/29/2015 | 0.174 | 2.027 | 0.232 | 0.390 | 2.824 | 0.285 | 0.052 | 0.669 | 0.409 | 1.416 | 0.106 | | |
| G1 | 4/16/2015 | 0.161 | 1.820 | 0.233 | 0.388 | 2.602 | 0.042 | 0.041 | 0.658 | 0.384 | 1.124 | 0.030 | | |
| G1 | 5/18/2015 | 0.171 | 1.842 | 0.229 | 0.405 | 2.647 | 0.036 | 0.040 | 0.638 | 0.365 | 1.079 | 0.028 | | |
| G1 | 6/9/2015 | 0.161 | 1.993 | 0.260 | 0.440 | 2.854 | 0.300 | 0.052 | 0.973 | 0.572 | 1.896 | 0.078 | | |
| G1 | 7/6/2015 | 0.167 | 1.934 | 0.235 | 0.396 | 2.732 | 0.246 | 0.047 | 0.745 | 0.430 | 1.468 | 0.098 | | |
| 2. Jamestown 1 | | | | | | | | | | | | | | |
| J1 | 7/22/2014 | 0.170 | 1.667 | 0.230 | 0.380 | 2.446 | 0.152 | 0.163 | 0.242 | 0.591 | 1.148 | 0.046 | | |
| J1 | 7/31/2014 | 0.163 | 1.749 | 0.248 | 0.386 | 2.547 | 0.168 | 0.165 | 0.207 | 0.491 | 1.031 | 0.046 | | |
| J1 | 9/15/2014 | 0.169 | 1.505 | 0.234 | 0.401 | 2.309 | 0.165 | 0.168 | 0.234 | 0.549 | 1.115 | 0.034 | | |
| J1 | 12/10/2014 | 0.186 | 1.357 | 0.237 | 0.399 | 2.179 | 0.158 | 0.157 | 0.258 | 0.625 | 1.198 | 0.032 | | |
| J1 | 1/2/2015 | 0.170 | 1.467 | 0.237 | 0.402 | 2.275 | 0.159 | 0.163 | 0.264 | 0.686 | 1.273 | 0.032 | | |
| J1 | 1/26/2015 | 0.181 | 0.351 | 0.240 | 0.378 | 1.150 | 0.141 | 0.175 | 0.249 | 0.513 | 1.078 | 0.032 | | |
| J1 | 4/16/2015 | 0.170 | 1.424 | 0.230 | 0.391 | 2.214 | 0.000 | 0.148 | 0.191 | 0.518 | 0.857 | 0.000 | | |
| J1 | 6/20/2015 | 0.182 | 0.871 | 0.208 | 0.343 | 1.604 | 0.222 | 0.164 | 0.349 | 0.874 | 1.608 | 0.062 | | |
| J1 | 7/24/2015 | 0.156 | 1.583 | 0.237 | 0.390 | 2.367 | 0.143 | 0.182 | 0.194 | 0.526 | 1.045 | 0.049 | | |
| 3. Jamestown 2 | | | | | | | | | | | | | | |
| J2 | 7/15/2014 | 0.502 | 16.417 | 0.639 | 0.928 | 18.485 | 0.593 | 0.083 | 1.238 | 1.372 | 3.287 | 0.084 | | |
| J2 | 9/15/2014 | 0.613 | 8.004 | 0.407 | 1.128 | 10.151 | 0.000 | 0.076 | 0.917 | 0.784 | 1.776 | 0.050 | | |
| J2 | 12/10/2014 | 0.663 | 8.851 | 0.453 | 1.025 | 10.992 | 0.000 | 0.066 | 1.190 | 0.701 | 1.957 | 0.036 | | |
| J2 | 1/2/2015 | 0.213 | 2.838 | 0.661 | 0.915 | 4.627 | 0.000 | 0.070 | 1.152 | 0.655 | 1.878 | 0.049 | | |
| J2 | 1/26/2015 | 0.209 | 1.873 | 0.180 | 1.174 | 3.436 | 0.000 | 0.041 | 0.927 | 1.155 | 2.123 | 0.032 | | |
| J2 | 4/16/2015 | 0.400 | 13.478 | 0.533 | 1.226 | 15.637 | 0.000 | 0.085 | 1.202 | 1.334 | 2.621 | 0.070 | | |
| J2 | 6/20/2015 | 0.597 | 13.122 | 0.945 | 1.222 | 15.886 | 1.025 | 0.078 | 1.259 | 1.377 | 3.739 | 0.153 | | |
| 4. Jamestown 3 | | | | | | | | | | | | | | |
| J3 | 7/4/2014 | 0.103 | 0.325 | 0.222 | 0.240 | 0.890 | 0.070 | 0.037 | 0.175 | 0.405 | 0.688 | 0.015 | | |
| J3 | 7/15/2014 | 0.123 | 0.496 | 0.177 | 0.243 | 1.039 | 0.074 | 0.041 | 0.164 | 0.388 | 0.667 | 0.015 | | |

| Site | Date Collected | Acid Anions | | | | | Base Cations | | | | | Dissolved Metals | | |
|-----------------------|----------------|-------------|----------------|--------------|----------------|----------------|--------------|------------|----------------|----------------|----------------|------------------|--|--|
| | | Cl- (meq/L) | SO4^2- (meq/L) | NO3- (meq/L) | PO4^3- (meq/L) | Sum AA (meq/L) | Na+ (meq/L) | K+ (meq/L) | Mg ^2+ (meq/L) | Ca ^2+ (meq/L) | Sum BC (meq/L) | Al^3+ (meq/L) | | |
| 4. Jamestown 3 | | | | | | | | | | | | | | |
| J3 | 7/22/2014 | 0.125 | 0.422 | 0.105 | 0.239 | 0.891 | 0.076 | 0.040 | 0.177 | 0.368 | 0.662 | 0.016 | | |
| J3 | 7/31/2014 | 0.147 | 0.342 | 0.147 | 0.178 | 0.814 | 0.081 | 0.038 | 0.163 | 0.361 | 0.642 | 0.014 | | |
| J3 | 9/15/2014 | 0.126 | 0.277 | 0.152 | 0.201 | 0.755 | 0.067 | 0.035 | 0.152 | 0.363 | 0.617 | 0.013 | | |
| J3 | 12/10/2014 | 0.148 | 0.284 | 0.180 | 0.223 | 0.836 | 0.067 | 0.035 | 0.152 | 0.382 | 0.637 | 0.013 | | |
| J3 | 1/2/2015 | 0.137 | 0.412 | 0.130 | 0.215 | 0.895 | 0.067 | 0.035 | 0.152 | 0.377 | 0.631 | 0.013 | | |
| J3 | 1/22/2015 | 0.101 | 0.398 | 0.172 | 0.178 | 0.848 | 0.067 | 0.035 | 0.153 | 0.360 | 0.615 | 0.013 | | |
| J3 | 1/26/2015 | 0.092 | 0.327 | 0.152 | 0.259 | 0.830 | 0.067 | 0.036 | 0.152 | 0.383 | 0.638 | 0.013 | | |
| J3 | 4/16/2015 | 0.132 | 0.356 | 0.119 | 0.247 | 0.855 | 0.018 | 0.039 | 0.152 | 0.408 | 0.618 | 0.002 | | |
| J3 | 6/20/2015 | 0.182 | 0.335 | 0.243 | 0.371 | 1.131 | 0.094 | 0.041 | 0.180 | 0.452 | 0.768 | 0.009 | | |
| J3 | 7/24/2015 | 0.136 | 0.431 | 0.155 | 0.172 | 0.894 | 0.074 | 0.039 | 0.169 | 0.411 | 0.692 | 0.015 | | |
| 5. Nashville 1 | | | | | | | | | | | | | | |
| N1 | 6/8/2014 | 0.410 | 46.106 | 0.548 | 0.934 | 47.998 | 1.262 | 0.019 | 12.708 | 23.140 | 37.130 | 0.361 | | |
| N1 | 7/1/2014 | 0.454 | 46.829 | 0.527 | 0.762 | 48.573 | 1.161 | 0.018 | 12.654 | 23.338 | 37.171 | 0.378 | | |
| N1 | 7/22/2014 | 0.462 | 47.538 | 0.536 | 0.861 | 49.397 | 1.093 | 0.017 | 12.641 | 23.084 | 36.835 | 0.369 | | |
| N1 | 8/26/2014 | 0.336 | 36.811 | 0.440 | 0.312 | 37.898 | 0.954 | 0.018 | 12.472 | 22.632 | 36.076 | 0.278 | | |
| N1 | 10/11/2014 | 0.415 | 36.757 | 0.291 | 0.285 | 37.749 | 1.191 | 0.018 | 13.589 | 23.947 | 38.745 | 0.282 | | |
| N1 | 1/9/2015 | 0.326 | 36.338 | 0.292 | 0.285 | 37.241 | 1.223 | 0.018 | 12.111 | 23.429 | 36.782 | 0.269 | | |
| N1 | 1/30/2015 | 0.326 | 36.331 | 0.397 | 0.280 | 37.335 | 1.092 | 0.018 | 13.144 | 23.226 | 37.480 | 0.262 | | |
| N1 | 3/23/2015 | 0.378 | 49.611 | 0.488 | 0.280 | 50.756 | 0.044 | 0.015 | 14.747 | 30.977 | 45.784 | 0.004 | | |
| N1 | 6/1/2015 | 0.326 | 46.044 | 0.464 | 0.301 | 47.134 | 0.770 | 0.015 | 16.365 | 31.386 | 48.535 | 0.045 | | |
| N1 | 7/24/2015 | 0.502 | 47.627 | 0.590 | 0.710 | 49.429 | 1.236 | 0.018 | 13.278 | 22.436 | 36.968 | 0.356 | | |
| N1 | 8/21/2015 | 0.531 | 47.793 | 0.541 | 0.849 | 49.714 | 1.108 | 0.019 | 12.404 | 23.269 | 36.800 | 0.347 | | |
| 6. Nashville 2 | | | | | | | | | | | | | | |
| N2 | 7/1/2014 | 0.469 | 17.949 | 0.649 | 1.076 | 20.143 | 1.341 | 0.053 | 12.075 | 25.671 | 39.140 | 5.406 | | |
| N2 | 8/26/2014 | 0.517 | 17.419 | 0.701 | 1.051 | 19.687 | 1.273 | 0.049 | 11.898 | 26.873 | 40.094 | 4.614 | | |
| N2 | 10/11/2014 | 0.399 | 18.158 | 0.758 | 0.602 | 19.917 | 1.511 | 0.048 | 10.744 | 26.530 | 38.832 | 4.517 | | |
| N2 | 1/9/2015 | 0.358 | 15.292 | 0.330 | 1.343 | 17.323 | 11.931 | 0.164 | 3.520 | 24.788 | 40.403 | 4.949 | | |
| N2 | 3/23/2015 | 0.465 | 15.987 | 0.556 | 0.740 | 17.748 | 0.000 | 0.045 | 10.514 | 31.141 | 41.700 | 0.002 | | |
| N2 | 6/1/2015 | 0.414 | 17.497 | 0.557 | 0.614 | 19.082 | 2.057 | 0.049 | 12.630 | 29.135 | 43.871 | 4.487 | | |
| N2 | 7/24/2015 | 0.327 | 17.911 | 0.236 | 0.980 | 19.455 | 1.248 | 0.049 | 11.419 | 25.984 | 38.700 | 5.563 | | |
| N2 | 8/21/2015 | 0.422 | 9.505 | 0.552 | 0.929 | 11.409 | 1.365 | 0.050 | 11.338 | 26.391 | 39.144 | 5.159 | | |

| Site | Date Collected | Acid Anions | | | | | | Base Cations | | | | | | Dissolved Metals | | |
|-------------------|----------------|-------------|-------------------|---------|-------------------|---------|---------|--------------|------------------|------------------|---------|------------------|---------|------------------|---------|--|
| | | Cl- | SO4 ²⁻ | NO3- | PO4 ³⁻ | Sum AA | Na+ | K+ | Mg ²⁺ | Ca ²⁺ | Sum BC | Al ³⁺ | (meq/L) | (meq/L) | (meq/L) | |
| | | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) | | | | |
| 7. Ocoee 1 | | | | | | | | | | | | | | | | |
| O1 | 6/12/2014 | 0.126 | 1.037 | 0.370 | 0.304 | 1.836 | 0.530 | 0.008 | 1.230 | 2.361 | 4.129 | 0.036 | | | | |
| O1 | 6/19/2014 | 0.208 | 0.619 | 0.307 | 0.188 | 1.322 | 0.654 | 0.010 | 1.528 | 2.038 | 4.229 | 0.040 | | | | |
| O1 | 6/26/2014 | 0.162 | 1.512 | 0.536 | 0.345 | 2.555 | 0.564 | 0.009 | 1.454 | 2.507 | 4.534 | 0.034 | | | | |
| O1 | 7/1/2014 | 0.144 | 1.064 | 0.319 | 0.294 | 1.822 | 0.512 | 0.009 | 1.478 | 2.219 | 4.218 | 0.034 | | | | |
| O1 | 7/17/2014 | 0.168 | 0.601 | 0.202 | 0.340 | 1.312 | 0.502 | 0.008 | 1.362 | 2.391 | 4.263 | 0.037 | | | | |
| O1 | 7/20/2014 | 0.118 | 0.417 | 0.524 | 0.343 | 1.403 | 0.525 | 0.009 | 1.324 | 2.508 | 4.365 | 0.038 | | | | |
| O1 | 9/28/2014 | 0.095 | 0.510 | 0.256 | 0.227 | 1.088 | 0.484 | 0.007 | 1.270 | 2.110 | 3.871 | 0.031 | | | | |
| O1 | 11/2/2014 | 0.115 | 1.019 | 0.234 | 0.178 | 1.546 | 0.521 | 0.008 | 1.327 | 2.065 | 3.921 | 0.032 | | | | |
| O1 | 1/3/2015 | 0.108 | 0.800 | 0.192 | 0.175 | 1.275 | 0.497 | 0.008 | 1.315 | 2.045 | 3.865 | 0.031 | | | | |
| O1 | 1/25/2015 | 0.159 | 0.441 | 0.231 | 0.219 | 1.049 | 0.495 | 0.008 | 1.290 | 2.096 | 3.888 | 0.031 | | | | |
| O1 | 2/10/2015 | 0.081 | 0.426 | 0.268 | 0.209 | 0.984 | 0.499 | 0.008 | 1.201 | 1.992 | 3.700 | 0.032 | | | | |
| O1 | 4/8/2015 | 0.176 | 0.932 | 0.345 | 0.299 | 1.753 | 0.094 | 0.006 | 1.053 | 1.759 | 2.912 | 0.001 | | | | |
| O1 | 5/18/2015 | 0.153 | 0.985 | 0.358 | 0.236 | 1.733 | 0.075 | 0.006 | 1.082 | 1.747 | 2.910 | 0.001 | | | | |
| O1 | 6/9/2015 | 0.177 | 0.728 | 0.230 | 0.392 | 1.528 | 0.606 | 0.010 | 1.365 | 2.242 | 4.223 | 0.027 | | | | |
| O1 | 7/6/2015 | 0.159 | 0.952 | 0.477 | 0.339 | 1.926 | 0.547 | 0.008 | 1.354 | 2.251 | 4.161 | 0.034 | | | | |
| 8. Ocoee 2 | | | | | | | | | | | | | | | | |
| O2 | 6/26/2014 | 0.180 | 1.216 | 0.209 | 0.323 | 1.928 | 0.245 | 0.032 | 0.864 | 1.660 | 2.802 | 0.051 | | | | |
| O2 | 7/1/2014 | 0.183 | 1.051 | 0.220 | 0.351 | 1.806 | 0.234 | 0.031 | 0.787 | 1.704 | 2.756 | 0.033 | | | | |
| O2 | 7/17/2014 | 0.179 | 1.157 | 0.201 | 0.331 | 1.869 | 0.076 | 0.025 | 0.655 | 1.501 | 2.256 | 0.000 | | | | |
| O2 | 7/20/2014 | 0.185 | 0.987 | 0.204 | 0.334 | 1.710 | 0.222 | 0.030 | 0.756 | 1.669 | 2.677 | 0.009 | | | | |
| O2 | 9/28/2014 | 0.159 | 0.981 | 0.193 | 0.309 | 1.642 | 0.205 | 0.024 | 0.639 | 1.431 | 2.300 | 0.028 | | | | |
| O2 | 11/2/2014 | 0.161 | 0.891 | 0.186 | 0.295 | 1.533 | 0.162 | 0.024 | 0.619 | 1.359 | 2.164 | 0.041 | | | | |
| O2 | 1/3/2015 | 0.156 | 1.029 | 0.177 | 0.303 | 1.664 | 0.310 | 0.024 | 0.683 | 1.370 | 2.388 | 0.013 | | | | |
| O2 | 1/25/2015 | 0.160 | 0.964 | 0.189 | 0.304 | 1.617 | 0.242 | 0.024 | 0.673 | 1.348 | 2.287 | 0.019 | | | | |
| O2 | 2/10/2015 | 0.161 | 0.909 | 0.189 | 0.309 | 1.569 | 0.225 | 0.023 | 0.647 | 1.422 | 2.316 | 0.012 | | | | |
| O2 | 4/8/2015 | 0.180 | 1.117 | 0.195 | 0.320 | 1.812 | 0.247 | 0.035 | 0.973 | 1.639 | 2.894 | 0.034 | | | | |
| O2 | 5/18/2015 | 0.177 | 1.083 | 0.232 | 0.351 | 1.843 | 0.233 | 0.028 | 0.784 | 1.558 | 2.604 | 0.036 | | | | |
| O2 | 6/9/2015 | 0.191 | 1.152 | 0.208 | 0.279 | 1.829 | 0.275 | 0.026 | 1.027 | 1.477 | 2.806 | 0.002 | | | | |
| O2 | 7/6/2015 | 0.175 | 0.986 | 0.212 | 0.327 | 1.699 | 0.403 | 0.034 | 0.857 | 1.783 | 3.076 | 0.068 | | | | |

| Site | Date Collected | Acid Anions | | | | Base Cations | | | | | | Dissolved Metals | | |
|------------------------|----------------|-------------|---------------------------|--------------|---------------------------|----------------|-------------|------------|--------------------------|--------------------------|----------------|--------------------------|--|--|
| | | Cl- (meq/L) | SO4 ²⁻ (meq/L) | NO3- (meq/L) | PO4 ³⁻ (meq/L) | Sum AA (meq/L) | Na+ (meq/L) | K+ (meq/L) | Mg ²⁺ (meq/L) | Ca ²⁺ (meq/L) | Sum BC (meq/L) | Al ³⁺ (meq/L) | | |
| 9. Ocoee 3 | | | | | | | | | | | | | | |
| O3 | 7/1/2014 | 0.162 | 1.682 | 0.220 | 0.331 | 2.395 | 1.458 | 0.058 | 0.224 | 1.423 | 3.163 | 0.571 | | |
| O3 | 7/17/2014 | 0.180 | 1.964 | 0.228 | 0.369 | 2.740 | 1.396 | 0.064 | 0.222 | 1.165 | 2.847 | 0.695 | | |
| O3 | 7/20/2014 | 0.187 | 1.572 | 0.221 | 0.377 | 2.356 | 1.031 | 0.048 | 0.251 | 1.106 | 2.436 | 0.557 | | |
| O3 | 9/28/2014 | 0.158 | 1.603 | 0.201 | 0.311 | 2.273 | 1.006 | 0.038 | 0.212 | 1.036 | 2.292 | 0.502 | | |
| O3 | 11/2/2014 | 0.156 | 1.753 | 0.201 | 0.309 | 2.419 | 1.011 | 0.052 | 0.225 | 1.023 | 2.311 | 0.597 | | |
| O3 | 1/3/2015 | 0.257 | 2.804 | 0.392 | 0.470 | 3.924 | 1.862 | 0.074 | 0.407 | 1.757 | 4.100 | 0.990 | | |
| O3 | 1/22/2015 | 0.254 | 2.889 | 0.321 | 0.436 | 3.901 | 1.471 | 0.088 | 0.363 | 1.935 | 3.858 | 0.772 | | |
| O3 | 1/25/2015 | 0.221 | 2.901 | 0.306 | 0.531 | 3.959 | 2.656 | 0.093 | 0.358 | 1.578 | 4.685 | 0.727 | | |
| O3 | 2/10/2015 | 0.250 | 2.510 | 0.306 | 0.460 | 3.525 | 1.856 | 0.073 | 0.467 | 1.507 | 3.903 | 1.003 | | |
| O3 | 4/8/2015 | 0.170 | 1.876 | 0.232 | 0.316 | 2.593 | 0.000 | 0.024 | 0.119 | 0.989 | 1.132 | 0.320 | | |
| O3 | 5/18/2015 | 0.178 | 2.004 | 0.221 | 0.358 | 2.760 | 0.000 | 0.028 | 0.156 | 1.348 | 1.531 | 0.377 | | |
| O3 | 6/9/2015 | 0.187 | 1.356 | 0.215 | 0.393 | 2.151 | 1.773 | 0.027 | 0.153 | 1.390 | 3.344 | 0.513 | | |
| O3 | 7/6/2015 | 0.181 | 2.011 | 0.235 | 0.344 | 2.771 | 1.535 | 0.075 | 0.273 | 1.123 | 3.006 | 0.389 | | |
| O3 | 8/8/2015 | 0.172 | 1.787 | 0.224 | 0.345 | 2.528 | 0.901 | 0.028 | 0.214 | 1.364 | 2.506 | 0.801 | | |
| 10. Seiverville | | | | | | | | | | | | | | |
| S1 | 7/17/2014 | 0.172 | 2.548 | 0.208 | 0.311 | 3.237 | 0.750 | 0.074 | 2.083 | 2.736 | 5.644 | 0.255 | | |
| S1 | 7/21/2014 | 0.181 | 2.442 | 0.224 | 0.371 | 3.217 | 0.590 | 0.074 | 2.137 | 2.663 | 5.464 | 0.257 | | |
| S1 | 7/29/2014 | 0.169 | 2.428 | 0.176 | 0.349 | 3.122 | 0.748 | 0.073 | 2.087 | 2.676 | 5.584 | 0.255 | | |
| S1 | 8/10/2014 | 0.186 | 2.380 | 0.167 | 0.347 | 3.081 | 0.608 | 0.074 | 2.117 | 2.557 | 5.356 | 0.243 | | |
| S1 | 8/30/2014 | 0.162 | 2.151 | 0.161 | 0.299 | 2.773 | 0.556 | 0.067 | 1.924 | 2.422 | 4.968 | 0.219 | | |
| S1 | 11/16/2014 | 0.161 | 2.176 | 0.181 | 0.304 | 2.822 | 0.580 | 0.067 | 1.927 | 2.431 | 5.004 | 0.219 | | |
| S1 | 1/22/2015 | 0.293 | 4.618 | 0.227 | 0.444 | 5.582 | 1.054 | 0.094 | 2.861 | 3.806 | 7.815 | 0.253 | | |
| S1 | 1/29/2015 | 0.298 | 4.457 | 0.297 | 0.419 | 5.471 | 0.727 | 0.111 | 3.227 | 4.079 | 8.144 | 0.330 | | |
| S1 | 4/16/2015 | 0.183 | 2.578 | 0.178 | 0.344 | 3.283 | 0.000 | 0.073 | 2.141 | 2.801 | 5.015 | 0.000 | | |
| S1 | 5/18/2015 | 0.181 | 2.638 | 0.233 | 0.330 | 3.383 | 0.000 | 0.073 | 2.134 | 2.751 | 4.958 | 0.000 | | |
| S1 | 6/9/2015 | 0.191 | 2.251 | 0.233 | 0.387 | 3.062 | 0.969 | 0.071 | 2.204 | 2.897 | 6.140 | 0.093 | | |
| S1 | 7/6/2015 | 0.179 | 2.555 | 0.189 | 0.336 | 3.259 | 0.637 | 0.074 | 2.129 | 2.681 | 5.521 | 0.244 | | |

| Site | Date Collected | Fe ²⁺ (meq/L) | Mn ²⁺ (meq/L) | Si ⁴⁺ (meq/L) | Zn ²⁺ (meq/L) | Ba ²⁺ (meq/L) | Cu ²⁺ (meq/L) | Cd ²⁺ (meq/L) | Ni ²⁺ (meq/L) | ANC (meq/L) |
|-----------------------|----------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|-------------|
| 1. Grainger | | | | | | | | | | |
| G1 | 7/21/2014 | 0.009 | 0.027 | 0.787 | 0.014 | 0.002 | 0.002 | 0.000 | 0.004 | -0.240 |
| G1 | 10/10/2014 | 0.006 | 0.029 | 0.838 | 0.014 | 0.002 | 0.002 | 0.000 | 0.003 | 0.017 |
| G1 | 10/23/2014 | 0.007 | 0.029 | 0.811 | 0.015 | 0.002 | 0.002 | 0.000 | 0.003 | -0.038 |
| G1 | 1/29/2015 | 0.008 | 0.026 | 0.840 | 0.015 | 0.002 | 0.002 | 0.000 | 0.003 | -0.406 |
| G1 | 4/16/2015 | 0.002 | 0.015 | 0.780 | 0.006 | 0.001 | 0.002 | 0.000 | 0.003 | -0.638 |
| G1 | 5/18/2015 | 0.002 | 0.014 | 0.783 | 0.006 | 0.001 | 0.001 | 0.000 | 0.002 | -0.730 |
| G1 | 6/9/2015 | 0.005 | 0.023 | 1.017 | 0.009 | 0.002 | 0.002 | 0.000 | 0.003 | 0.181 |
| G1 | 7/6/2015 | 0.007 | 0.027 | 0.785 | 0.014 | 0.002 | 0.002 | 0.000 | 0.004 | -0.326 |
| 2. Jamestown 1 | | | | | | | | | | |
| J1 | 7/22/2014 | 0.007 | 0.012 | 0.082 | 0.008 | 0.001 | 0.002 | 0.000 | 0.001 | -1.139 |
| J1 | 7/31/2014 | 0.008 | 0.012 | 0.080 | 0.008 | 0.001 | 0.002 | 0.000 | 0.001 | -1.358 |
| J1 | 9/15/2014 | 0.003 | 0.012 | 0.083 | 0.008 | 0.001 | 0.002 | 0.000 | 0.001 | -1.049 |
| J1 | 12/10/2014 | 0.004 | 0.012 | 0.078 | 0.009 | 0.001 | 0.002 | 0.000 | 0.001 | -0.841 |
| J1 | 1/2/2015 | 0.004 | 0.012 | 0.076 | 0.009 | 0.001 | 0.002 | 0.000 | 0.001 | -0.865 |
| J1 | 1/26/2015 | 0.003 | 0.012 | 0.087 | 0.008 | 0.001 | 0.002 | 0.000 | 0.000 | 0.074 |
| J1 | 4/16/2015 | 0.002 | 0.000 | 0.118 | 0.000 | 0.001 | 0.001 | 0.000 | 0.000 | -1.234 |
| J1 | 6/20/2015 | 0.002 | 0.001 | 0.153 | 0.002 | 0.001 | 0.001 | 0.000 | 0.001 | 0.227 |
| J1 | 7/24/2015 | 0.008 | 0.012 | 0.079 | 0.008 | 0.001 | 0.002 | 0.000 | 0.001 | -1.162 |
| 3. Jamestown 2 | | | | | | | | | | |
| J2 | 7/15/2014 | 0.066 | 0.082 | 1.069 | 0.004 | 0.004 | 0.005 | 0.001 | 0.006 | -13.877 |
| J2 | 9/15/2014 | 0.004 | 0.055 | 0.767 | 0.003 | 0.003 | 0.006 | 0.001 | 0.004 | -7.483 |
| J2 | 12/10/2014 | 0.003 | 0.086 | 0.669 | 0.002 | 0.003 | 0.005 | 0.002 | 0.003 | -8.227 |
| J2 | 1/2/2015 | 0.002 | 0.079 | 0.544 | 0.003 | 0.003 | 0.005 | 0.001 | 0.004 | -2.060 |
| J2 | 1/26/2015 | 0.003 | 0.074 | 0.927 | 0.004 | 0.003 | 0.006 | 0.001 | 0.004 | -0.259 |
| J2 | 4/16/2015 | 0.004 | 0.084 | 1.093 | 0.004 | 0.004 | 0.005 | 0.001 | 0.002 | -11.749 |
| J2 | 6/20/2015 | 0.095 | 0.076 | 0.956 | 0.004 | 0.004 | 0.002 | 0.001 | 0.007 | -10.848 |
| 4. Jamestown 3 | | | | | | | | | | |
| J3 | 7/4/2014 | 0.002 | 0.003 | 0.347 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 0.168 |
| J3 | 7/15/2014 | 0.002 | 0.003 | 0.381 | 0.003 | 0.001 | 0.001 | 0.000 | 0.000 | 0.033 |

| Site | Date Collected | Fe ²⁺ (meq/L) | Mn ²⁺ (meq/L) | Si ⁴⁺ (meq/L) | Zn ²⁺ (meq/L) | Ba ²⁺ (meq/L) | Cu ²⁺ (meq/L) | Cd ²⁺ (meq/L) | Ni ²⁺ (meq/L) | ANC (meq/L) |
|-----------------------|----------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|-------------|
| 4. Jamestown 3 | | | | | | | | | | |
| J3 | 7/22/2014 | 0.002 | 0.003 | 0.358 | 0.003 | 0.001 | 0.001 | 0.000 | 0.000 | 0.154 |
| J3 | 7/31/2014 | 0.002 | 0.003 | 0.393 | 0.003 | 0.001 | 0.001 | 0.000 | 0.000 | 0.244 |
| J3 | 9/15/2014 | 0.001 | 0.003 | 0.330 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 0.214 |
| J3 | 12/10/2014 | 0.002 | 0.003 | 0.333 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 0.156 |
| J3 | 1/2/2015 | 0.002 | 0.003 | 0.327 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 0.085 |
| J3 | 1/22/2015 | 0.002 | 0.003 | 0.327 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 0.116 |
| J3 | 1/26/2015 | 0.002 | 0.003 | 0.335 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 0.165 |
| J3 | 4/16/2015 | 0.001 | 0.000 | 0.365 | 0.001 | 0.001 | 0.000 | 0.000 | 0.000 | 0.133 |
| J3 | 6/20/2015 | 0.002 | 0.000 | 0.409 | 0.001 | 0.001 | 0.000 | 0.000 | 0.000 | 0.059 |
| J3 | 7/24/2015 | 0.002 | 0.003 | 0.362 | 0.003 | 0.001 | 0.001 | 0.000 | 0.000 | 0.184 |
| 5. Nashville 1 | | | | | | | | | | |
| N1 | 6/8/2014 | 0.021 | 0.095 | 1.039 | 0.073 | 0.007 | 0.017 | 0.002 | 0.011 | -9.242 |
| N1 | 7/1/2014 | 0.022 | 0.096 | 1.027 | 0.068 | 0.008 | 0.016 | 0.003 | 0.010 | -9.774 |
| N1 | 7/22/2014 | 0.021 | 0.095 | 1.048 | 0.070 | 0.008 | 0.016 | 0.002 | 0.011 | -10.924 |
| N1 | 8/26/2014 | 0.018 | 0.095 | 1.058 | 0.071 | 0.008 | 0.018 | 0.003 | 0.006 | -0.265 |
| N1 | 10/11/2014 | 0.016 | 0.092 | 1.026 | 0.067 | 0.007 | 0.017 | 0.000 | 0.006 | 2.509 |
| N1 | 1/9/2015 | 0.019 | 0.095 | 1.050 | 0.070 | 0.008 | 0.020 | 0.003 | 0.008 | 1.081 |
| N1 | 1/30/2015 | 0.017 | 0.097 | 1.029 | 0.071 | 0.008 | 0.017 | 0.002 | 0.011 | 1.659 |
| N1 | 3/23/2015 | 0.004 | 0.039 | 1.667 | 0.022 | 0.005 | 0.008 | 0.002 | 0.020 | -3.203 |
| N1 | 6/1/2015 | 0.003 | 0.035 | 1.356 | 0.008 | 0.005 | 0.007 | 0.003 | 0.014 | 2.877 |
| N1 | 7/24/2015 | 0.022 | 0.095 | 1.056 | 0.068 | 0.008 | 0.018 | 0.002 | 0.008 | -10.828 |
| N1 | 8/21/2015 | 0.020 | 0.095 | 0.978 | 0.066 | 0.007 | 0.019 | 0.003 | 0.008 | -11.373 |
| 6. Nashville 2 | | | | | | | | | | |
| N2 | 7/1/2014 | 0.090 | 0.207 | 2.937 | 0.276 | 0.007 | 0.034 | 0.002 | 0.072 | 28.028 |
| N2 | 8/26/2014 | 0.078 | 0.222 | 3.130 | 0.273 | 0.007 | 0.034 | 0.002 | 0.067 | 28.832 |
| N2 | 10/11/2014 | 0.124 | 0.221 | 3.015 | 0.269 | 0.008 | 0.030 | 0.003 | 0.075 | 27.177 |
| N2 | 1/9/2015 | 0.055 | 0.210 | 3.157 | 0.275 | 0.008 | 0.034 | 0.004 | 0.072 | 31.844 |
| N2 | 3/23/2015 | 0.004 | 0.045 | 0.925 | 0.005 | 0.005 | 0.008 | 0.002 | 0.005 | 24.951 |
| N2 | 6/1/2015 | 0.006 | 0.133 | 3.235 | 0.227 | 0.005 | 0.030 | 0.001 | 0.082 | 32.996 |
| N2 | 7/24/2015 | 0.027 | 0.209 | 2.927 | 0.260 | 0.007 | 0.034 | 0.002 | 0.066 | 28.340 |
| N2 | 8/21/2015 | 0.083 | 0.212 | 3.074 | 0.276 | 0.007 | 0.032 | 0.001 | 0.073 | 36.652 |

| Site | Date Collected | Fe ²⁺ (meq/L) | Mn ²⁺ (meq/L) | Si ⁴⁺ (meq/L) | Zn ²⁺ (meq/L) | Ba ²⁺ (meq/L) | Cu ²⁺ (meq/L) | Cd ²⁺ (meq/L) | Ni ²⁺ (meq/L) | ANC (meq/L) |
|-------------------|----------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|-------------|
| 7. Ocoee 1 | | | | | | | | | | |
| O1 | 6/12/2014 | 0.002 | 0.001 | 1.218 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 3.554 |
| O1 | 6/19/2014 | 0.001 | 0.001 | 1.068 | 0.002 | 0.001 | 0.002 | 0.000 | 0.000 | 4.022 |
| O1 | 6/26/2014 | 0.002 | 0.001 | 1.160 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 3.181 |
| O1 | 7/1/2014 | 0.002 | 0.001 | 1.188 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 3.625 |
| O1 | 7/17/2014 | 0.002 | 0.001 | 1.126 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 4.121 |
| O1 | 7/20/2014 | 0.001 | 0.001 | 1.139 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 4.146 |
| O1 | 9/28/2014 | 0.002 | 0.001 | 1.053 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 3.874 |
| O1 | 11/2/2014 | 0.002 | 0.001 | 1.037 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 3.451 |
| O1 | 1/3/2015 | 0.001 | 0.001 | 1.083 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 3.710 |
| O1 | 1/25/2015 | 0.002 | 0.001 | 0.998 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 3.874 |
| O1 | 2/10/2015 | 0.001 | 0.001 | 1.011 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 3.764 |
| O1 | 4/8/2015 | 0.001 | 0.000 | 0.675 | 0.001 | 0.001 | 0.002 | 0.000 | 0.000 | 1.840 |
| O1 | 5/18/2015 | 0.001 | 0.000 | 0.745 | 0.001 | 0.001 | 0.002 | 0.000 | 0.000 | 1.928 |
| O1 | 6/9/2015 | 0.001 | 0.000 | 1.045 | 0.002 | 0.001 | 0.001 | 0.000 | 0.000 | 3.771 |
| O1 | 7/6/2015 | 0.002 | 0.001 | 1.120 | 0.002 | 0.001 | 0.002 | 0.000 | 0.000 | 3.396 |
| 8. Ocoee 2 | | | | | | | | | | |
| O2 | 6/26/2014 | 0.004 | 0.006 | 0.351 | 0.008 | 0.001 | 0.003 | 0.000 | 0.000 | 1.299 |
| O2 | 7/1/2014 | 0.004 | 0.007 | 0.411 | 0.005 | 0.001 | 0.003 | 0.001 | 0.000 | 1.416 |
| O2 | 7/17/2014 | 0.001 | 0.002 | 0.347 | 0.001 | 0.001 | 0.001 | 0.000 | 0.001 | 0.741 |
| O2 | 7/20/2014 | 0.001 | 0.000 | 0.391 | 0.001 | 0.001 | 0.001 | 0.000 | 0.000 | 1.373 |
| O2 | 9/28/2014 | 0.002 | 0.009 | 0.315 | 0.004 | 0.001 | 0.002 | 0.000 | 0.000 | 1.020 |
| O2 | 11/2/2014 | 0.002 | 0.009 | 0.310 | 0.005 | 0.001 | 0.002 | 0.000 | 0.000 | 1.001 |
| O2 | 1/3/2015 | 0.002 | 0.001 | 0.319 | 0.005 | 0.001 | 0.002 | 0.000 | 0.000 | 1.068 |
| O2 | 1/25/2015 | 0.002 | 0.011 | 0.312 | 0.006 | 0.001 | 0.002 | 0.000 | 0.000 | 1.024 |
| O2 | 2/10/2015 | 0.002 | 0.006 | 0.312 | 0.007 | 0.001 | 0.001 | 0.000 | 0.000 | 1.090 |
| O2 | 4/8/2015 | 0.004 | 0.012 | 0.382 | 0.007 | 0.001 | 0.002 | 0.000 | 0.000 | 1.525 |
| O2 | 5/18/2015 | 0.005 | 0.006 | 0.356 | 0.009 | 0.001 | 0.002 | 0.001 | 0.000 | 1.176 |
| O2 | 6/9/2015 | 0.005 | 0.014 | 0.498 | 0.002 | 0.001 | 0.002 | 0.000 | 0.001 | 1.502 |
| O2 | 7/6/2015 | 0.005 | 0.018 | 0.398 | 0.013 | 0.002 | 0.004 | 0.001 | 0.001 | 1.886 |

| Site | Date Collected | Fe ²⁺ (meq/L) | Mn ²⁺ (meq/L) | Si ⁴⁺ (meq/L) | Zn ²⁺ (meq/L) | Ba ²⁺ (meq/L) | Cu ²⁺ (meq/L) | Cd ²⁺ (meq/L) | Ni ²⁺ (meq/L) | ANC (meq/L) |
|------------------------|----------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|-------------|
| 9. Ocoee 3 | | | | | | | | | | |
| O3 | 7/1/2014 | 0.025 | 0.129 | 0.747 | 0.084 | 0.006 | 0.029 | 0.003 | 0.004 | 2.364 |
| O3 | 7/17/2014 | 0.023 | 0.116 | 0.804 | 0.085 | 0.006 | 0.020 | 0.001 | 0.000 | 1.859 |
| O3 | 7/20/2014 | 0.027 | 0.121 | 0.582 | 0.079 | 0.007 | 0.020 | 0.002 | 0.006 | 1.479 |
| O3 | 9/28/2014 | 0.022 | 0.105 | 0.724 | 0.073 | 0.005 | 0.019 | 0.001 | 0.002 | 1.473 |
| O3 | 11/2/2014 | 0.021 | 0.103 | 0.669 | 0.072 | 0.005 | 0.015 | 0.001 | 0.003 | 1.379 |
| O3 | 1/3/2015 | 0.040 | 0.116 | 1.039 | 0.100 | 0.008 | 0.026 | 0.005 | 0.001 | 2.501 |
| O3 | 1/22/2015 | 0.036 | 0.144 | 0.701 | 0.124 | 0.010 | 0.030 | 0.006 | 0.005 | 1.785 |
| O3 | 1/25/2015 | 0.036 | 0.170 | 1.528 | 0.117 | 0.008 | 0.036 | 0.005 | 0.007 | 3.358 |
| O3 | 2/10/2015 | 0.037 | 0.174 | 1.108 | 0.111 | 0.012 | 0.030 | 0.005 | 0.010 | 2.867 |
| O3 | 4/8/2015 | 0.027 | 0.025 | 0.841 | 0.010 | 0.005 | 0.008 | 0.003 | 0.002 | -0.220 |
| O3 | 5/18/2015 | 0.007 | 0.027 | 1.216 | 0.014 | 0.006 | 0.008 | 0.002 | 0.005 | 0.431 |
| O3 | 6/9/2015 | 0.005 | 0.027 | 1.112 | 0.015 | 0.005 | 0.011 | 0.000 | 0.006 | 2.886 |
| O3 | 7/6/2015 | 0.023 | 0.112 | 0.425 | 0.081 | 0.006 | 0.022 | 0.003 | 0.001 | 1.296 |
| O3 | 8/8/2015 | 0.027 | 0.119 | 1.184 | 0.078 | 0.007 | 0.018 | 0.002 | 0.005 | 2.219 |
| 10. Seiverville | | | | | | | | | | |
| S1 | 7/17/2014 | 0.014 | 0.067 | 0.384 | 0.046 | 0.005 | 0.009 | 0.002 | 0.001 | 3.190 |
| S1 | 7/21/2014 | 0.015 | 0.066 | 0.359 | 0.047 | 0.005 | 0.010 | 0.003 | 0.002 | 3.010 |
| S1 | 7/29/2014 | 0.014 | 0.068 | 0.373 | 0.047 | 0.005 | 0.009 | 0.003 | 0.001 | 3.238 |
| S1 | 8/10/2014 | 0.014 | 0.068 | 0.344 | 0.045 | 0.005 | 0.010 | 0.002 | 0.001 | 3.007 |
| S1 | 8/30/2014 | 0.012 | 0.062 | 0.335 | 0.041 | 0.004 | 0.009 | 0.002 | 0.002 | 2.882 |
| S1 | 11/16/2014 | 0.012 | 0.062 | 0.336 | 0.041 | 0.004 | 0.009 | 0.001 | 0.002 | 2.869 |
| S1 | 1/22/2015 | 0.019 | 0.074 | 0.596 | 0.061 | 0.008 | 0.018 | 0.004 | 0.003 | 3.269 |
| S1 | 1/29/2015 | 0.022 | 0.108 | 0.455 | 0.064 | 0.008 | 0.017 | 0.005 | 0.001 | 3.683 |
| S1 | 4/16/2015 | 0.003 | 0.013 | 0.522 | 0.008 | 0.004 | 0.005 | 0.001 | 0.002 | 2.289 |
| S1 | 5/18/2015 | 0.002 | 0.013 | 0.534 | 0.007 | 0.004 | 0.006 | 0.001 | 0.002 | 2.144 |
| S1 | 6/9/2015 | 0.003 | 0.013 | 0.412 | 0.008 | 0.003 | 0.005 | 0.000 | 0.000 | 3.616 |
| S1 | 7/6/2015 | 0.013 | 0.069 | 0.369 | 0.045 | 0.005 | 0.010 | 0.002 | 0.002 | 3.021 |

Appendix E: Literature Review of Toxicity Thresholds to Acidification for Aquatic Biota

Fish and macroinvertebrate survival, growth and productivity are dependent on both biological and environmental factors. Long-term acid deposition caused chronic and episodic aquatic acidification, leading to the depression of pH and increase of aluminum and metals. Most studies about the toxicity of stream acidification to salmonids and macroinvertebrate are emphasized on the effects of reduced pH, elevated aluminum and metals concentration to fish abundance and mortality, and macroinvertebrate biodiversity. Few studies also researched the toxicity threshold of nitrate and nitrite to aquatic biota (Westin 1974; Lewis and Morris 1986). As the stream concentrations of nitrate are generally about 100 times lower than the toxicity threshold, current report will not discuss the toxicity of nitrate and nitrite but just pH, aluminum and metals.

E.1. pH

Protons (H^+) could be lethal to fish by causing loss of Na^+ and Cl^- across the gill (Spry and Wiener 1991). The mechanism of acid toxicity to fish and macroinvertebrate is to disrupt ion regulation, leading to a severe deficiency of extracellular ions (Courtney and Clements 1998; Felten and Guérolde 2006). Generally, the reduction of pH will lead to the increase of aluminum concentrations in stream to increase the toxicity. However, the co-existing calcium can prolong survival time of fishes at acidic solution. The concentration above 1.4 mg L^{-1} calcium can bring a marked improvement in fishery status even in the most acid lakes with pH 4.3-4.6 (Howells *et al.* 1983). Aqueous calcium reduces the toxicity of both H^+ and aluminum at the gills, presumably by reducing gill membrane permeability and subsequent loss of ions (Spry and Wiener 1991).

Albaster and Lloyd (1980) reported that there is likely harmful to the eggs and fry of salmonids when pH at the range of 4.5 to 5.0. When pH is reduced to 3.5 - 4.0, it is lethal to salmonids. For macroinvertebrate, acidity may affect the biodiversity. The abundance of some species sensitive to acid will be significantly reduced. It was reported that pelecypoda cannot usually tolerate a pH below 4.7-5.5 (Johnson *et al.* 1993). One study about the macroinvertebrate communities in 200 streams of the western Adirondack Mountains found that macroinvertebrate assemblages were usually unaffected above pH 6.4, were slightly impacted at pHs of 5.7-6.4, moderately impacted from pHs of 5.1-5.7, and severely impacted at pHs less than 5.1 (Baldigo *et al.* 2009). Table E.1 summarizes the effects of different pH to different biota and life stage by different experiments.

E.2. Metals

Acute metal toxicity to salmonids is often characterized by gill damage and the hypersecretion of mucus (Handy and Eddy 1990). Mortalities are related to physiological disturbances to respiration resulting in hypoxia and also ionoregulatory disturbances resulting in body ion depletion.

E.2.1 Dissolved Aluminum

Dissolved aluminum is often regarded as the most toxic metal for fish and invertebrates in acidified waters (Hermann *et al.* 1993). The mechanism of Al toxicity to fish are attributed to the inability of fish to maintain their osmoregulatory balance and respiratory problems associated with the coagulation of mucous on the gills (Driscoll 1985; Exley *et al.* 1991; Hermann *et al.* 1993). Aluminum tends to accumulate on the gill rather than in other organs (Spry and Wiener 1991). The accumulates of aluminum on the gills is presumed to displace Ca^{2+} and cause increased ion and electrolyte loss, loss of ability to adsorb ions, hemo-concentration, and impair oxygen delivery to the fish tissues (Dussault *et al.* 2004). Other than the precipitation of solid $Al(OH)_3$ or cellular

Table E.1. Toxicity threshold values of pH for trout (salmonids) and benthic macroinvertebrates.

| Biota name | Methods | pH | Co-existing chemicals | Effects | References |
|---|---|-----------|--|--|--------------------------------|
| Fish | | 4.5 | $\text{Ca}^{2+} < 0.8 \text{ mg L}^{-1}$ | Lakes will be fishless | Howells <i>et al.</i> 1983 |
| Brook trout | Laboratory exposures for 5 months | 5.5 | | Reduced growth | Menendez 1976 |
| Brook trout | field experiments in acid brook water | ~5 | | Reduced growth | Muniz and Leivestad 1979 |
| Brook trout fry | Field exposure to episodic acidification for 20 days in Adirondack Lake | 4.8 | | 100% mortality | Van Offelen <i>et al.</i> 1994 |
| Brook trout, eggs, larvae and young | Lab exposure for 30 days | 4.5 | | Adverse effects on mortality, growth, behavior and biochemical responses | Cleveland <i>et al.</i> 1986 |
| | | 5.5 | $\text{Al} = 300 \mu\text{g L}^{-1}$ | | |
| Brook trout | Field exposure to episodic acidification | < 5.0-5.2 | Inorganic Al > 100 -200 $\mu\text{g L}^{-1}$ | Trout abundance was reduced | Baker <i>et al.</i> 1996 |
| Introduced brook trout, sac fry | In-situ experiment within the North Branch of the Moose River | 4.32-4.4 | $\text{Al}_{\text{im}} = 0.19\text{-}0.21 \text{ mg L}^{-1}$ $\text{Ca}^{2+} = 1.13\text{-}1.40 \text{ mg L}^{-1}$ $\text{DOC} = 6.0\text{-}6.9 \text{ mg L}^{-1}$ | 0% survival after 240 hours | Johnson <i>et al.</i> 1987 |
| Introduced brook trout, feeding fry | | 4.53~4.87 | $\text{Al}_{\text{im}} = 0.18\text{-}0.25 \text{ mg L}^{-1}$ $\text{Ca}^{2+} = 1.08\text{-}1.68 \text{ mg L}^{-1}$ $\text{DOC} = 3.8\text{-}6.4 \text{ mg L}^{-1}$ | 0% survival after 336 hours | |
| Introduced brook trout, young of the year | | 4.37~4.68 | $\text{Al}_{\text{im}} = 0.11\text{-}0.34 \text{ mg L}^{-1}$ $\text{Ca}^{2+} = 0.41\text{-}1.30 \text{ mg L}^{-1}$ $\text{DOC} = 7.3\text{-}9.0 \text{ mg L}^{-1}$ | 0% survival after 1920 hours | |
| Introduced brook trout, yearling | | 4.44~4.68 | $\text{Al}_{\text{im}} = 0.11\text{-}0.18 \text{ mg L}^{-1}$ $\text{Ca}^{2+} = 0.41\text{-}1.03 \text{ mg L}^{-1}$ $\text{DOC} = 8.0\text{-}9.0 \text{ mg L}^{-1}$ | 0% survival after 672 hours | |
| Rainbow trout | Lab exposure up to 8 weeks | 5.2 | $\text{Ca}^{2+} = 12 \pm 7 \mu\text{mol L}^{-1}$ | decreased swimming capacity by 5% | Dussault <i>et al.</i> 2004 |
| Juvenile rainbow trout | Lab exposure to synthesized solution for 36 days | 5.2 | $\text{Ca}^{2+} = 28 \mu\text{eq L}^{-1}$ | 9-16% reduction of swimming capacity | Wilson <i>et al.</i> 1994 |

| Biota name | Methods | pH | Co-existing chemicals | Effects | Reference |
|--|--|-------------|---|--|----------------------------|
| Rainbow trout | Laboratory exposures to sublethal acid conditions over 3.5 months | 5.5 | | Reduced growth | Edwards and Hjeldnes 1977 |
| <i>G. fossarum</i> (Amphipoda), <i>H. pellucidula</i> (Trichoptera), <i>D. cephalotes</i> (Plecoptera) | Exposure for 24, 72 and 120 h in a stream in France | 4.73 ± 0.08 | Al _{tot} = 28.4 ± 1 μmol L ⁻¹ Ca ²⁺ = 39.1 ± 0.6 μmol L ⁻¹ | Decrease in survival rate and Na ⁺ , Cl ⁻ . <i>G. fossarum</i> most sensitive than <i>H. pellucidula</i> and <i>D. cephalotes</i> | Felten and Guérolde 2006 |
| Benthic macroinvertebra | Exposure to a stream with artificially added HNO ₃ to control pH 4.0, 5.5, 6.5 and 7.4 for 7 days | 4.0 | | Significant fewer individuals and taxa. Reduced abundance resulted primarily from reduced abundance of mayflies (Ephemeroptera) | Courtney and Clements 1998 |
| Macroinvertebrate | Native macroinvertebrate in streams affected by episodic acidification in Swiss | < 5.0 | Al _{tot} up to 140 μg L ⁻¹ | Lower taxonomic richness; scarce empididae, <i>Isoperla rivulorum</i> , <i>Rhithrogena</i> spp. and <i>Baetis</i> spp. | Lepori <i>et al.</i> 2003 |
| Macroinvertebrate | Native macroinvertebrate affected by episodic acidification in British streams | < 5.7-6 | | <i>Baetis muticus</i> , <i>Heptagenia lateralis</i> and <i>R. semicolorata</i> absent | Kowalik <i>et al.</i> 2007 |
| <i>Baetis alpinus</i> | Native <i>B. alpinus</i> affected by episodic acidification | 4.5-5.6 | | Decline to 10-20% during acid episodes | Lepori and Ormerod 2005 |

internalization of Al^{3+} , Poléo (1995) proposed that the process of aluminum polymerization is a mechanism of acute toxicity of aluminum to fish at pH 5.0-6.0. Also, Poléo (1995) suggested that positively charged Al-hydroxides bind to negatively charged sites of the gill surface to produce Al polymer, leading to severe clogging of the inter-lamellar space. This physical surface effect leads to acute hypoxia. As a result, the toxicity of aluminum applies primarily to fish at the gill-breathing stages. Therefore, mortality of fish by aluminum is primarily due to asphyxia at pH 6.1 and to electrolyte loss at pH 4.5 (Neville and Campbell 1988). In contrast to the adverse effect of high levels of aluminum, low levels of aluminum may protect fish from the effects of high hydrogen ion concentration by blocking the membrane permeability of hydrogen ions (Evans et al. 1988; Herrmann et al. 1993). The toxicity level was determined by the forms of dissolved aluminum (Table E.2). In general, inorganic monomeric aluminum is most toxic to fish, and the aluminum complexed to organic matter has least toxicity (Driscoll *et al.* 1980; Driscoll 1985; Baker *et al.* 1996).

The toxicity of aluminum is affected by other factors, including pH, calcium and DOC, and also by fish stage (Table D.2). It was reported that aluminum at less than $500 \mu\text{g L}^{-1}$ at pH 4.8-5.2 demonstrated a toxic effect to brook trout but had no effect at higher or lower pH (Schofield and Trojnar 1980). Calcium can moderate the toxicity of aluminum. Calcium was also shown to reduce plasma ion loss (Muniz and Leivestad 1980). At the conditions with low pH, low calcium and high aluminum concentrations, survival may be reduced, growth may be affected and consequently productivity will be low. Aluminum could complex with DOC to be less toxic (Spry and Wiener 1991; Serrano *et al.* 2008).

The most sensitive stage to acid is the newly hatched fry, but the later swim-up fry is more sensitive to aluminum (Baker and Schofield 1980). The embryo is the life stage least sensitive to aluminum. After hatching, the sensitivity of fish to both acid and aluminum decreases with increasing age- a pattern reported for brook trout (Spry and Wiener 1991). It was suggested that salmonid eggs and yolk sac fry are less vulnerable to the combination of low pH and aluminum than other early life stages (Serrano *et al.* 2008).

Driscoll *et al.* (2001) suggested that the appropriate thresholds for chemical and biological recovery in streams and lakes of the Northeast U S are pH of 6.0 and Al_{IM} of $2.0 \mu\text{mol L}^{-1}$. The mortality of fish is also determined to the length of time the fish are exposure. Some studies show that 2 day of exposure to acutely toxic Al_{IM} concentrations the approximate minimum exposure before brook trout begin to die (Gagen *et al.* 1993; Simonin *et al.* 1993; Van Sickle *et al.* 1996; Baldigo and Murdoch 1997). Some other aluminum thresholds were reported in the northeastern US. Significant mortality of brook trout was found when Al_{IM} levels exceeded 0.2 mg L^{-1} for 2 or more days (Baldigo and Murdoch 1997); or when Al_{tot} concentration reached $0.2 - 0.3 \text{ mg L}^{-1}$ for 1.5 or more days (Gagen and Sharpe 1987a, 1987b); or when Al_{M} concentration of 0.1 mg L^{-1} during acid episodes (Simonin *et al.* 1993); or when Al_{IM} and (or) Al_{tot} exceed either 0.2 or 0.3 mg L^{-1} under low Ca ($< 2.0 \text{ mg L}^{-1}$), DOC ($< 2.0 \text{ mg L}^{-1}$) and pH (4.4 - 5.2) conditions (Van Sickle *et al.* 1996).

E.2.2 Other dissolved metals

Many dissolved metals, especially select heavy metals, were studied about the toxicity to aquatic biota. However, in eastern Tennessee (Blue Ridge region) only five metals in streams were monitored (Al, Cu, Fe, Mn, and Zn). Therefore, review of toxicity threshold values will focus on these five metals. Table E.3 summarized toxicity threshold values of metals to aquatic biota.

Table E.2. Toxicity threshold values of dissolved aluminum concentrations for trout (salmonids).

| Fish name | Methods | Al concentrations | | Co-existing chemicals | Effects | References |
|------------------------------------|--|------------------------------|--|--|--|----------------------------|
| | | Total Al | Inorganic monomeric Al | | | |
| Brown trout | | 7 $\mu\text{mol L}^{-1}$ | | pH = 5.0 | Loss of Na and Cl from the blood | Muniz and Leivestad 1979 |
| Brook trout fry | Lab exposure to synthesized solutions for 14 days | 18-36 $\mu\text{mol L}^{-1}$ | | pH = 4.8 | Gill damage | Schofield and Trojnar 1980 |
| Brook trout | Lab exposure for 193 days | | 47 $\mu\text{g L}^{-1}$ | pH = 5.0 Ca ²⁺ = 0.5 mg L ⁻¹ | 44% mortality | Mount <i>et al.</i> 1988 |
| Brook trout, young-of-the-year | Exposure to stream waters for 30 days during each spring from 1995 to 2000 | | Median: 4.48 $\mu\text{mol L}^{-1}$; Range: 2.02-13.89 | Median: pH = 5.03; NO ₃ ⁻ = 263 $\mu\text{mol L}^{-1}$ | 100% mortality | Baldigo <i>et al.</i> 2005 |
| Brook trout | Field exposure to episodic acidification for 10 days | > 200 $\mu\text{g L}^{-1}$ | | pH < 5.1 | 10-19% loss of whole-body sodium | Neff <i>et al.</i> 2009 |
| Brook trout, young-of-the-year | Exposure to spring episodic acidification for 30 days in the SW Adirondack Mountains (2001- 2003). | | > 4 $\mu\text{mol L}^{-1}$ | | 50-100% mortality during two to four days of exposure | Baldigo and Lawrence 2007 |
| Brook trout, larvae and postlarvae | Lab exposure to softened and dechlorinated water for 13 to 14 days | 0.2 mg L ⁻¹ | | pH range from 4.2 to 5.6 | Measurable reductions in survival and growth | Baker and Schofield 1982 |
| Adult rainbow trout | Lab exposure to synthesized solutions for 10 days | >10 $\mu\text{mol L}^{-1}$ | | Pathological changes were more severe with aluminum at pH 5.4 than at pH 4.7 | Cause chloride cell necrosis and a decline in cell numbers | Evans <i>et al.</i> 1988 |

| Fish name | Methods | Al concentration | | Co-existing chemicals | Effects | References |
|----------------------------|---|----------------------------|------------------------|--|--|---------------------------|
| | | Total Al | Inorganic monomeric Al | | | |
| Juvenile rainbow trout | 26h of exposure to dechlorinated tap water with added aluminum | > 200 $\mu\text{g/L}$ | | pH = 6.0 | Affect cough rate, which is defined as disruptions in the ventilation pattern. | Ogilvie and Stechey 1983 |
| Juvenile rainbow trout | 26h of exposure to dechlorinated tap water with added aluminum | > 500 $\mu\text{g L}^{-1}$ | | pH = 6.0 | Affect ventilation rate, which is the number of opercular cycles per unit of time. | Ogilvie and Stechey 1983 |
| Rainbow trout | Lab exposure up to 8 weeks | 89 $\mu\text{g L}^{-1}$ | | pH = 5.1-5.2 $\text{Ca}^{2+} = 12 \pm 7 \mu\text{mol L}^{-1}$ | 25% survival, decreased swimming capacity by 21% | Dussault et al. 2004 |
| Juvenile rainbow trout | Lab exposure to synthesized solution for 36 days | 38 $\mu\text{g L}^{-1}$ | | pH = 5.2 $\text{Ca}^{2+} = 28 \mu\text{eq L}^{-1}$ | 15-21% reduction of swimming capacity | Wilson et al. 1994 |
| Juvenile rainbow trout | Lab exposure to different pH levels with same Al concentration solution for 11 days | 2.8 $\mu\text{mol L}^{-1}$ | | pH = 6.1 | Uptake of O_2 across the gill epithelium was reduced | Neville and Campbell 1988 |
| | | | | pH = 4.5 | Increase gill membrane permeability to H^+ , Na^+ and Cl^- ions | |
| Rainbow trout, fingerlings | Exposure to synthesized solution with varied Al and pH (7.0-9.0) for 45 days | 5.2 mg L^{-1} | | pH range from 7.0 to 9.0 | Seriously disturb any natural population of young trout with longer than 6 weeks exposure | Freeman and Everhart 1971 |

Table E.3. Toxicity threshold values of metals other than aluminum to fish and benthic macroinvertebrates.

| Fish name | Methods | Metals | Co-existing chemicals and conditions | Effects | References |
|------------------------------|--|--|---|--|---------------------------|
| Rainbow trout, swim-up stage | Laboratory exposures to synthesized well water with designed metal concentration | Cd = 1.9 $\mu\text{g L}^{-1}$ Cu = 40 $\mu\text{g L}^{-1}$ Zn = 219 $\mu\text{g L}^{-1}$ | pH = 8.24 Alk = 92 mg L^{-1} Hardness = 103 mg L^{-1} as CaCO_3 Ca^{2+} = 25 mg L^{-1} Mg^{2+} = 8.0 mg L^{-1} Na^+ = 8.3 mg L^{-1} SO_4^{2-} = 18 mg L^{-1} Cl^- = 9 mg L^{-1} DOC < 1 mg L^{-1} | Start to affect survival | Besser et al. 2007 |
| Rainbow trout | | Zn = 47 $\mu\text{g L}^{-1}$ | 112 mg L^{-1} CaCO_3 pH = 7.6 | Fish avoidance | Black and Birge 1980 |
| Rainbow trout | | Zn = 144 $\mu\text{g L}^{-1}$ | 25 mg L^{-1} CaCO_3 pH = 7.6 | Effect on fish ventilation | Cairns et al. 1982 |
| Benthic macroinvertebrate | Exposure to a mixture of Cd, Cu and Zn for 7 days in a stream microcosm | Cd = 1.1 $\mu\text{g L}^{-1}$, Cu = 12 $\mu\text{g L}^{-1}$, Zn = 110 $\mu\text{g L}^{-1}$ | | Abundance of three mayfly species was reduced by more than 50% | Clements and Kiffney 1994 |

E.3 References: Toxicity Literature Review

- Alabaster, J.S. and Lloyd, R. (1980). *Water quality criteria for freshwater fish*. London, Butterworths, 21-45.
- Baker, J.P. and Schofield, C.L. (1980). Aluminium toxicity to fish as related to acid precipitation and Adirondack surface water quality. In *Proc. Int. Conf. Ecological Impact of Acid Precipitation* Sandefjor, Norway.
- Baker, J.P. and Schofield, C.L. (1982). Aluminum toxicity to fish in acidic waters. *Water, Air, and Soil Pollution*, 18, 289-309.
- Baker, J.P., Van Sickle, J., Gagen, C.J., DeWalle, D.R., Shappe, W.E., Carline, R.F., Baldigo, B.P., Murdoch, P.S., Bath, D.W., Kretser, W.A., Simonin, H.A. and Wigington Jr., P.J. (1996). Episodic acidification of small streams in the northeastern United States: effects on fish populations. *Ecological Applications*, 6(2), 422-437.
- Baldigo, B.P. and Lawrence, G. (2007). Persistent mortality of brook trout in episodically acidified streams of the southwestern Adirondack Mountains, New York. *Transactions of the American Fisheries Society*, 136, 121-134.

- Baldigo, B.P., Lawrence, G.B., Bode, R.W., Simonin, H.A., Roy, K.M. and Smith, A.J. (2009). Impacts of acidification on macroinvertebrate communities in streams of the western Adirondack Mountains, New York, USA. *Ecological Indicators*, 9, 226-239.
- Baldigo, B.P. and Murdoch, P.S. (1997). Effect of stream acidification and inorganic aluminum on mortality of brook trout (*Salvelinus fontinalis*) in the Catskill Mountains, New York. *Canadian Journal of Fisheries and Aquatic Sciences*, 54(3), 603–615.
- Baldigo, B.P., Murdoch, P.S. and Burns, D.A. (2005). Stream acidification and mortality of brook trout (*Salvelinus fontinalis*) in response to timber harvest in Catskill Mountain watershed, New York, USA. *Canadian Journal of Fisheries and Aquatic Sciences*, 62, 1168-1183.
- Besser, J.M., Mebane, C.A., Mount, D.R., Ivey, C.D., Kunz, J.L., Greer, I.E., May, T.W. and Ingersoll, C.G. (2007). Sensitivity of mottled sculpins (*Cottus bairdi*) and rainbow trout (*Onchorhynchus mykiss*) to acute and chronic toxicity of cadmium, copper, and zinc. *Environmental Toxicology and Chemistry*, 26(8), 1657-1665.
- Black, J.A. and Birge, W.J. (1980). An avoidance response bioassay for aquatic pollutants. University of Kentucky, *Water Resources Research Institute Research Report*, 123, 1-34.
- Cairns, M.A., Garton, R.R. and Tubb, R.A. (1982). Use of fish ventilation frequency to estimate chronically safe toxicant concentrations. *Transactions of the American Fisheries Society*, 111, 70-77.
- Clements, W.H. and Kiffney, P.M. (1994). Integrated laboratory and field approach for assessing impacts of heavy metals at the Arkansas River, Colorado. *Environmental Toxicology and Chemistry*, 13, 397-404.
- Cleveland, L., Little, E.E., Hamilton, S.J., Buckler, D.R. and Hunn, J.B. (1986). Interactive toxicity of aluminum and acidity to early life stages of brook trout. *Transactions of the American Fisheries Society*, 115, 610–620.
- Courtney, L.A. and Clements, W.H. (1998). Effects of acidic pH on benthic macroinvertebrate communities in stream microcosms. *Hydrobiologia*, 379, 135-145.
- Driscoll, C.T. (1985). Aluminum in acidic surface waters: chemistry, transport, and effects. *Environmental Health Perspectives*, 63, 93-104.
- Driscoll, C.T., Baker, J.P., Bisogni, J.J. and Schofield, C.L. (1980). Effect of aluminum speciation on fish in dilute acidified waters. *Nature*, 284, 161-164.
- Driscoll, C.T., Lawrence, G.B., Bulger, A.J., Butler, T.J., Cronan, C.S., Eagar, C., Lambert, K.F., Likens, G.E., Stoddard, J.L. and Weathers, K.C. (2001). Acidic deposition in the northeastern United States: sources and inputs, ecosystem effects, and management strategies. *BioScience*, 51(3), 180-198.
- Dussault È.B., Playle, R.C., Dixon, D.G. and McKinley, R.S. (2004). Effects of chronic aluminum exposure on swimming and cardiac performance in rainbow trout, *Oncorhynchus mykiss*. *Fish Physiology and Biochemistry*, 30, 137-148.
- Edwards, D.J. and Hjeldnes, S. (1977). Growth and survival of salmonids in water of different pH. SNSF, FR 10/77.
- Evans, R.E., Brown, S.B. and Hara, T.J. (1988). The effects of aluminum and acid on the gill morphology in rainbow trout, *Salmo gairdneri*. *Environmental Biology of Fishes*, 22(4), 299-311.
- Exley, C., Chappell, J.S. and Birchall, J.D. (1991). A mechanism for acute aluminium toxicity in fish. *Journal of Theoretical Biology*, 151, 417-428.
- Felten, V. and Guérol, F. (2006). Short-term physiological responses to a severe acid stress in three macroinvertebrate species: A comparative study. *Chemosphere*, 63, 1427-1435.

- Freeman, R.A. and Everhart, W.H. (1971). Toxicity of aluminum hydroxide complexes in neutral and basic media to rainbow trout. *Transactions of the American Fisheries Society*, 4, 644–658.
- Gagen, C.J. and Sharpe, W.E. (1987a). Net sodium loss and mortality of three salmonid species exposed to a stream acidified by atmospheric deposition. *Bulletin of Environmental Contamination and Toxicology*, 39, 7-14.
- Gagen, C.J. and Sharpe, W.E. (1987b). Influence of acid runoff episodes on survival and net sodium balance of brook trout (*Salvelinus fontinalis*) confined in a mountain stream. In: *Ecophysiology of Acid Stress in Aquatic Organisms*. (Eds Witters, H. and Vanderborght, O.) Societe Royale Zoologique de Belgique, Antwerp, Belgium. 219-230.
- Gagen, C.J., Sharpe, W.E. and Carline, R.F. (1993). Mortality of brook trout, mottled sculpins, and slimy sculpins during acidic episodes. *Transactions of the American Fisheries Society*, 122(4), 616–628.
- Handy, R.D. and Eddy, F.B. (1990). The interactions between the surface of rainbow trout, *Oncorhynchus mykiss*, and waterborne metal toxicants. *Functional Ecology*, 4, 385-392.
- Hermann, J., Degerman, E., Gerhardt, A., Johansson, C., Lingdell, P. and Muniz, I.P. (1993). Acid-stress effects on stream biology. *Ambio*, 22(5), 298-306.
- Howells, G.D., Brown, D.J.A. and Sadler, K. (1983). Effects of acidity, calcium, and aluminum on fish survival and productivity – A review. *Journal of the Science of Food and Agriculture*, 34, 559-570.
- Johnson, D.W., Simonin, H.A., Colquhoun, J.R. and Flack, F.M. (1987). In situ toxicity tests of fishes in acid waters. *Biogeochemistry*, 3, 181-208.
- Johnson, R.K., Wiederholm, T. and Rosenberg, D.M. (1993). Freshwater biomonitoring using individual organisms, populations, and species assemblages of benthic macroinvertebrates. In: *Freshwater Biomonitoring and Benthic Macroinvertebrates* (Eds Rosenerg, D.M. and Resh, V.H.), Chapman and Hall, New York, 40-158.
- Kowalik, R.A., Cooper, D.M., Evans, C.D. and Ormerod, S.J. (2007). Acidic episodes retard the biological recovery of upland British streams from chronic acidification. *Global Change Biology*, 13, 2439-2452.
- Lepori, F., Barbieri, A. and Ormerod, J. (2003). Effects of episodic acidification on macroinvertebrate assemblages in Swiss Alpine streams. *Freshwater Biology*, 48, 1873-1885.
- Lepori, F. and Ormerod, S.J. (2005). Effects of spring acid episodes on macroinvertebrates revealed by population data and *in situ* toxicity tests. *Freshwater Biology*, 50, 1568-1577.
- Lewis Jr., W.M. and Morris, D.P. (1986). Toxicity of nitrite to fish: A review. *Transactions of the American Fisheries Society*, 115, 183-195.
- Menendez, R. (1976). Chronic effects of reduced pH on brook trout (*Salvelinus fontinalis*). *Journal of the Fisheries Research Board of Canada*, 33, 118-123.
- Mount, D.R., Hockett, J.R. and Gern, W.A. (1988). Effect of long-term exposure to acid, aluminum, and low calcium on adult brook trout *Salvelinus fontinalis*. 2. Vitellogenesis and osmoregulation. *Canadian Journal of Fisheries and Aquatic Sciences*, 45, 1633-42.
- Muniz, I.P. and Leivestad, H. (1979). Long term exposure of fish to acid water (*Salvelinus fontinalis*) SNSF, IR 44/79 (in Norwegian, English summary).
- Muniz, I.P. and Leivestad, H. (1979). Toxic effects of aluminum on the brown trout, *Salmo trutta* L. In: *Ecological Impact of Acid Precipitation* (D. Drablos and A. Tollan, Ed.), SNSF Project, Oslo, Norway, 1979, 268-269.
- Muniz, I.P. and Leivestad, H. (1980). Acidification-effects on freshwater fish. In *Proc. Int. Conf. Ecological Impact of Acid Precipitation Sandefjor, Norway*, 84-92.

- Neff K.J., Schwartz, J.S., Henry T.B., Robinson, R.B., Moore, S.E. and Kulp M.A. (2009). Physiological stress in native southern brook trout during episodic stream acidification in the Great Smoky Mountains National Park. *Archives of Environmental Contamination and Toxicology*, 57(2), 366-376.
- Neville, C.M. and Campbell, P.G. (1988). Possible mechanisms of aluminum toxicity in a dilute, acidic environment to fingerlings and older life stages of salmonids. *Water, Air, and Soil Pollution*, 42, 311-327.
- Ogilvie, D.M. and Stechey, D.M. (1983). Effects of aluminum on respiratory responses and spontaneous activity of rainbow trout, *Salmo gairdneri*. *Environmental Toxicology and Chemistry*, 2, 43-48.
- Poléo, A.B.S. (1995). Aluminium polymerization – a mechanism of acute toxicity of aqueous aluminium to fish. *Aquatic Toxicology*, 31, 347-356.
- Schofield, C.L. and Trojnar, J.R. (1980). Aluminum toxicity to brook trout (*Salvelinus fontinalis*) in acidified waters. In: *Polluted Rain* (T.Y. Toribara, M.W. Miler and P.E. Morrow, Eds.), Plenum Press, New York, 341-366.
- Simonin, H.A., Kretser, W.A., Bath, D.W., Olson, M. and Gallagher, J. (1993). In situ bioassays of brook trout (*Salvelinus fontinalis*) and blacknose dace (*Rhinichthys atratulus*) in Adirondack streams affected by episodic acidification. *Canadian Journal of Fisheries and Aquatic Sciences*, 50(5), 902–912.
- Spry, D.J. and Wiener, J.G. (1991). Metal bioavailability and toxicity to fish in low-alkalinity lakes: A critical review. *Environmental Pollution*, 71, 243-304.
- Serrano, I., Buffam, I., Palm, D., Brannas, E. and Laudon, H. (2008). Thresholds for survival of brown trout during the spring flood acid pulse in streams high in dissolved organic carbon. *Transactions of the American Fisheries Society*, 137, 1363–1377.
- Van Offelen, H.K., Krueger, C.C., Schofield, C.L. and Keleher, C. (1994). Survival, distribution, and ion composition in two strains of brook trout (*Salvelinus fontinalis*) fry after exposure to episodic pH depressions in an Adirondack Lake. *Canadian Journal of Fisheries and Aquatic Sciences*, 51, 792-799.
- Van Sickle, J., Baker, J.P., Simonin, H.A., Baldigo, B.P., Kretser, W.A. and Sharpe, W.E. (1996). Episodic acidification of small streams in the northeastern United States: fish mortality in field bioassays. *Ecological Applications*, 6(2), 408–421.
- Westin, D.T. (1974). Nitrate and nitrite toxicity to salmonid fishes. *The Progressive Fish-Culturist*, 36(2), 86-89.
- Wilson, R.W., Bergman, H.L. and Wood, C.M. (1994). Metabolic costs and physiological consequences of acclimation to aluminum in juvenile rainbow trout (*Oncorhynchus mykiss*). 2: Gill morphology, swimming performance, and aerobic scope. *Canadian Journal of Fisheries and Aquatic Sciences*, 51, 536–544.

Text Contribution Acknowledgement:

Dr. Meijun Cai, University of Minnesota, Duluth, authored parts of Appendix E.

**Appendix F: Pollutant Export Mass Loadings from Monitoring Station
Storm Events**

APPENDIX F: ROAD CUT RUNOFF MASS LOADINGS SUMMARY SHEET

| Site | Date Collected | Rainfall (RF) | Runoff | Acid Anions | | | | Base Cations | | | |
|-----------------------|----------------|----------------------|--------------------------|-------------|---------------------------------------|---------------------------|---------------------------------------|--------------|------------|--------------------------|--------------------------|
| | | Rainfall (RF) (days) | Runoff (m ³) | Cl- (g/day) | SO ₄ ²⁻ (g/day) | NO ₃ - (g/day) | PO ₄ ³⁻ (g/day) | Na+ (g/day) | K+ (g/day) | Mg ²⁺ (g/day) | Ca ²⁺ (g/day) |
| 1. Grainger | | | | | | | | | | | |
| G1 | 7/21/2014 | 3 | 1.86 | 3.70 | 57.59 | 8.62 | 7.97 | 3.40 | 1.02 | 6.32 | 5.31 |
| G1 | 10/10/2014 | 1 | 1.22 | 7.52 | 102.28 | 17.77 | 14.48 | 7.21 | 2.27 | 11.46 | 11.86 |
| G1 | 10/23/2014 | 6 | 3.15 | 3.11 | 43.07 | 8.06 | 6.56 | 3.15 | 1.02 | 4.75 | 4.70 |
| G1 | 1/29/2015 | 3 | 0.57 | 1.17 | 18.53 | 2.74 | 2.35 | 1.25 | 0.39 | 1.55 | 1.56 |
| G1 | 4/16/2015 | 3 | 1.35 | 2.57 | 39.44 | 6.53 | 5.55 | 0.43 | 0.71 | 3.61 | 3.47 |
| G1 | 5/18/2015 | 3 | 0.08 | 0.16 | 2.31 | 0.37 | 0.34 | 0.02 | 0.04 | 0.20 | 0.19 |
| G1 | 6/9/2015 | 2 | 1.17 | 3.33 | 55.92 | 9.43 | 8.15 | 4.03 | 1.18 | 6.91 | 6.68 |
| G1 | 7/6/2015 | 4 | 1.92 | 2.84 | 44.48 | 7.00 | 6.01 | 2.71 | 0.87 | 4.34 | 4.12 |
| 2. Jamestown 1 | | | | | | | | | | | |
| J1 | 7/22/2014 | 2 | 1.04 | 3.14 | 41.79 | 7.44 | 6.28 | 1.82 | 3.31 | 1.54 | 6.18 |
| J1 | 7/31/2014 | 1 | 6.01 | 34.80 | 504.95 | 92.42 | 73.58 | 23.30 | 38.64 | 15.12 | 59.05 |
| J1 | 9/15/2014 | 1 | 1.11 | 6.61 | 79.99 | 16.05 | 14.05 | 4.20 | 7.24 | 3.14 | 12.15 |
| J1 | 12/10/2014 | 1 | 1.44 | 9.47 | 93.64 | 21.11 | 18.14 | 5.23 | 8.78 | 4.50 | 17.98 |
| J1 | 1/2/2015 | 3 | 0.84 | 1.68 | 19.62 | 4.09 | 3.54 | 1.02 | 1.78 | 0.90 | 3.83 |
| J1 | 1/26/2015 | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| J1 | 4/16/2015 | 2 | 0.98 | 2.95 | 33.61 | 7.01 | 6.08 | 0.00 | 2.83 | 1.14 | 5.09 |
| J1 | 6/20/2015 | 1 | 0.67 | 4.32 | 28.06 | 8.68 | 7.30 | 3.42 | 4.30 | 2.85 | 11.73 |
| J1 | 7/24/2015 | 1 | 1.11 | 6.11 | 84.13 | 16.29 | 13.68 | 3.63 | 7.84 | 2.61 | 11.65 |
| 3. Jamestown 2 | | | | | | | | | | | |
| J2 | 7/15/2014 | 1 | 1.78 | 31.57 | 1399.61 | 70.33 | 52.17 | 24.24 | 5.78 | 26.73 | 48.74 |
| J2 | 9/15/2014 | 1 | 1.59 | 34.52 | 611.34 | 40.12 | 56.83 | 0.00 | 4.69 | 17.72 | 24.95 |
| J2 | 12/10/2014 | 1 | 2.07 | 48.51 | 877.80 | 58.00 | 67.05 | 0.00 | 5.33 | 29.87 | 28.97 |
| J2 | 1/2/2015 | 3 | 1.20 | 3.03 | 54.58 | 16.42 | 11.61 | 0.00 | 1.10 | 5.61 | 5.25 |
| J2 | 1/26/2015 | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| J2 | 4/16/2015 | 2 | 1.41 | 10.01 | 457.23 | 23.35 | 27.44 | 0.00 | 2.34 | 10.32 | 18.85 |
| J2 | 6/20/2015 | 1 | 0.97 | 20.39 | 608.02 | 56.57 | 37.34 | 22.76 | 2.95 | 14.76 | 26.58 |

| Site | Date Collected | Rainfall (RF) | Runoff | Acid Anions | | | | Base Cations | | | |
|-----------------------|----------------|----------------------|--------------------------|-------------|---------------------------|--------------|---------------------------|--------------|------------|--------------------------|--------------------------|
| | | Rainfall (RF) (days) | Runoff (m ³) | Cl- (g/day) | SO4 ²⁻ (g/day) | NO3- (g/day) | PO4 ³⁻ (g/day) | Na+ (g/day) | K+ (g/day) | Mg ²⁺ (g/day) | Ca ²⁺ (g/day) |
| 4. Jamestown 3 | | | | | | | | | | | |
| J3 | 7/4/2014 | 3 | 0.02 | 0.03 | 0.11 | 0.10 | 0.05 | 0.01 | 0.01 | 0.02 | 0.06 |
| J3 | 7/15/2014 | 1 | 0.18 | 0.77 | 4.21 | 1.94 | 1.36 | 0.30 | 0.28 | 0.35 | 1.37 |
| J3 | 7/22/2014 | 2 | 0.49 | 1.09 | 4.99 | 1.60 | 1.86 | 0.43 | 0.38 | 0.53 | 1.81 |
| J3 | 7/31/2014 | 1 | 1.80 | 9.39 | 29.62 | 16.45 | 10.17 | 3.35 | 2.66 | 3.57 | 13.01 |
| J3 | 9/15/2014 | 1 | 0.52 | 2.31 | 6.90 | 4.88 | 3.30 | 0.80 | 0.72 | 0.96 | 3.76 |
| J3 | 12/10/2014 | 1 | 0.56 | 2.95 | 7.70 | 6.30 | 3.98 | 0.87 | 0.78 | 1.04 | 4.31 |
| J3 | 1/2/2015 | 3 | 0.39 | 0.63 | 2.57 | 1.05 | 0.89 | 0.20 | 0.18 | 0.24 | 0.98 |
| J3 | 1/22/2015 | 1 | 0.12 | 0.43 | 2.28 | 1.27 | 0.67 | 0.18 | 0.17 | 0.22 | 0.86 |
| J3 | 1/26/2015 | 1 | 0.03 | 0.10 | 0.50 | 0.30 | 0.26 | 0.05 | 0.04 | 0.06 | 0.24 |
| J3 | 4/16/2015 | 2 | 0.34 | 0.80 | 2.91 | 1.25 | 1.33 | 0.07 | 0.26 | 0.31 | 1.39 |
| J3 | 6/20/2015 | 1 | 0.34 | 2.19 | 5.48 | 5.13 | 4.01 | 0.74 | 0.54 | 0.75 | 3.08 |
| J3 | 7/24/2015 | 1 | 0.45 | 2.16 | 9.29 | 4.32 | 2.45 | 0.77 | 0.68 | 0.92 | 3.69 |
| 5. Nashville 1 | | | | | | | | | | | |
| N1 | 6/8/2014 | 1 | 0.61 | 8.89 | 1355.25 | 20.80 | 18.11 | 17.78 | 0.44 | 94.55 | 283.41 |
| N1 | 7/1/2014 | 2 | 0.40 | 3.24 | 452.63 | 6.57 | 4.86 | 5.38 | 0.14 | 30.96 | 93.99 |
| N1 | 7/22/2014 | 1 | 2.57 | 41.99 | 5855.11 | 85.24 | 69.95 | 64.51 | 1.69 | 394.09 | 1184.65 |
| N1 | 8/26/2014 | 1 | 0.09 | 1.08 | 159.98 | 2.47 | 0.89 | 1.99 | 0.06 | 13.72 | 40.98 |
| N1 | 10/11/2014 | 2 | 1.76 | 12.94 | 1553.25 | 15.89 | 7.95 | 24.12 | 0.62 | 145.35 | 421.63 |
| N1 | 1/9/2015 | 2 | 0.99 | 5.71 | 861.80 | 8.93 | 4.47 | 13.90 | 0.35 | 72.71 | 231.52 |
| N1 | 1/30/2015 | 2 | 0.26 | 1.50 | 226.23 | 3.19 | 1.15 | 3.26 | 0.09 | 20.72 | 60.26 |
| N1 | 3/23/2015 | 2 | 0.13 | 0.84 | 150.11 | 1.91 | 0.56 | 0.06 | 0.04 | 11.29 | 39.05 |
| N1 | 6/1/2015 | 2 | 1.86 | 10.73 | 2056.69 | 26.76 | 8.87 | 16.47 | 0.54 | 185.03 | 584.14 |
| N1 | 7/24/2015 | 1 | 0.39 | 7.01 | 901.03 | 14.42 | 8.86 | 11.21 | 0.27 | 63.58 | 176.86 |
| N1 | 8/21/2015 | 2 | 1.07 | 10.06 | 1226.70 | 17.95 | 14.38 | 13.63 | 0.39 | 80.59 | 248.86 |
| 6. Nashville 2 | | | | | | | | | | | |
| N2 | 7/1/2014 | 2 | 0.31 | 2.57 | 133.45 | 6.23 | 5.28 | 4.78 | 0.32 | 22.72 | 79.53 |
| N2 | 8/26/2014 | 1 | 0.07 | 1.27 | 58.23 | 3.03 | 2.32 | 2.04 | 0.13 | 10.07 | 37.43 |
| N2 | 10/11/2014 | 2 | 1.35 | 9.56 | 590.22 | 31.82 | 12.91 | 23.53 | 1.26 | 88.40 | 359.31 |
| N2 | 1/9/2015 | 2 | 0.76 | 4.81 | 278.97 | 7.78 | 16.17 | 104.29 | 2.44 | 16.26 | 188.42 |
| N2 | 3/23/2015 | 2 | 0.10 | 0.80 | 37.21 | 1.67 | 1.14 | 0.00 | 0.08 | 6.19 | 30.20 |
| N2 | 6/1/2015 | 2 | 1.43 | 10.50 | 601.20 | 24.73 | 13.92 | 33.86 | 1.38 | 109.84 | 417.12 |
| N2 | 7/24/2015 | 1 | 0.30 | 3.51 | 260.66 | 4.45 | 9.41 | 8.70 | 0.58 | 42.06 | 157.56 |

| | | | | | | | | | | | |
|----|-----------|---|------|------|--------|-------|-------|-------|------|-------|--------|
| N2 | 8/21/2015 | 2 | 0.82 | 6.14 | 187.68 | 14.08 | 12.11 | 12.91 | 0.81 | 56.66 | 217.11 |
|----|-----------|---|------|------|--------|-------|-------|-------|------|-------|--------|

3

| Site | Date Collected | Rainfall (RF) | Runoff | Acid Anions | | | | Base Cations | | | |
|-------------------|----------------|---------------|-------------------|-------------|---------------------------|--------------|---------------------------|--------------|------------|--------------------------|--------------------------|
| | | (days) | (m ³) | Cl- (g/day) | SO4 ²⁻ (g/day) | NO3- (g/day) | PO4 ³⁻ (g/day) | Na+ (g/day) | K+ (g/day) | Mg ²⁺ (g/day) | Ca ²⁺ (g/day) |
| 7. Ocoee 1 | | | | | | | | | | | |
| O1 | 6/12/2014 | 3 | 0.55 | 0.81 | 9.05 | 4.17 | 1.75 | 2.22 | 0.06 | 2.72 | 8.59 |
| O1 | 6/19/2014 | 2 | 0.06 | 0.21 | 0.83 | 0.53 | 0.17 | 0.42 | 0.01 | 0.52 | 1.14 |
| O1 | 6/26/2014 | 4 | 1.04 | 1.49 | 18.88 | 8.64 | 2.84 | 3.37 | 0.09 | 4.60 | 13.05 |
| O1 | 7/1/2014 | 3 | 1.97 | 3.36 | 33.60 | 13.02 | 6.13 | 7.75 | 0.23 | 11.81 | 29.18 |
| O1 | 7/17/2014 | 6 | 0.33 | 0.33 | 1.60 | 0.70 | 0.60 | 0.64 | 0.02 | 0.92 | 2.65 |
| O1 | 7/20/2014 | 2 | 0.48 | 1.00 | 4.75 | 7.72 | 2.58 | 2.87 | 0.08 | 3.82 | 11.92 |
| O1 | 9/28/2014 | 1 | 0.09 | 0.29 | 2.10 | 1.36 | 0.62 | 0.95 | 0.02 | 1.32 | 3.62 |
| O1 | 11/2/2014 | 4 | 0.03 | 0.03 | 0.39 | 0.11 | 0.04 | 0.09 | 0.00 | 0.13 | 0.33 |
| O1 | 1/3/2015 | 1 | 0.00 | 0.01 | 0.09 | 0.03 | 0.01 | 0.03 | 0.00 | 0.04 | 0.09 |
| O1 | 1/25/2015 | 2 | 0.46 | 1.29 | 4.84 | 3.28 | 1.59 | 2.60 | 0.07 | 3.59 | 9.59 |
| O1 | 2/10/2015 | 1 | 0.12 | 0.34 | 2.44 | 1.98 | 0.79 | 1.37 | 0.04 | 1.74 | 4.76 |
| O1 | 4/8/2015 | 1 | 0.23 | 1.41 | 10.10 | 4.83 | 2.14 | 0.49 | 0.05 | 2.89 | 7.94 |
| O1 | 5/18/2015 | 4 | 1.01 | 1.37 | 11.95 | 5.61 | 1.89 | 0.44 | 0.06 | 3.32 | 8.83 |
| O1 | 6/9/2015 | 1 | 0.68 | 4.29 | 23.87 | 9.75 | 8.48 | 9.52 | 0.26 | 11.32 | 30.62 |
| O1 | 7/6/2015 | 4 | 5.07 | 7.11 | 57.86 | 37.45 | 13.61 | 15.94 | 0.42 | 20.84 | 57.02 |
| 8. Ocoee 2 | | | | | | | | | | | |
| O2 | 6/26/2014 | 4 | 0.87 | 1.38 | 12.66 | 2.81 | 2.22 | 1.22 | 0.27 | 2.28 | 7.20 |
| O2 | 7/1/2014 | 3 | 1.64 | 3.56 | 27.66 | 7.47 | 6.09 | 2.94 | 0.66 | 5.24 | 18.68 |
| O2 | 7/17/2014 | 6 | 0.28 | 0.29 | 2.57 | 0.58 | 0.48 | 0.08 | 0.05 | 0.37 | 1.39 |
| O2 | 7/20/2014 | 2 | 0.40 | 1.30 | 9.37 | 2.50 | 2.10 | 1.01 | 0.23 | 1.82 | 6.61 |
| O2 | 9/28/2014 | 1 | 0.07 | 0.40 | 3.36 | 0.85 | 0.70 | 0.34 | 0.07 | 0.55 | 2.04 |
| O2 | 11/2/2014 | 4 | 0.03 | 0.04 | 0.28 | 0.08 | 0.06 | 0.02 | 0.01 | 0.05 | 0.18 |
| O2 | 1/3/2015 | 1 | 0.00 | 0.01 | 0.09 | 0.02 | 0.02 | 0.01 | 0.00 | 0.02 | 0.05 |
| O2 | 1/25/2015 | 2 | 0.38 | 1.08 | 8.83 | 2.23 | 1.83 | 1.06 | 0.18 | 1.56 | 5.14 |
| O2 | 2/10/2015 | 1 | 0.10 | 0.57 | 4.34 | 1.17 | 0.98 | 0.52 | 0.09 | 0.78 | 2.83 |
| O2 | 4/8/2015 | 1 | 0.19 | 1.20 | 10.08 | 2.27 | 1.91 | 1.07 | 0.26 | 2.22 | 6.16 |
| O2 | 5/18/2015 | 4 | 0.84 | 1.32 | 10.95 | 3.02 | 2.34 | 1.13 | 0.23 | 2.01 | 6.56 |
| O2 | 6/9/2015 | 1 | 0.57 | 3.84 | 31.46 | 7.33 | 5.03 | 3.60 | 0.58 | 7.10 | 16.81 |
| O2 | 7/6/2015 | 4 | 4.22 | 6.53 | 49.95 | 13.84 | 10.92 | 9.78 | 1.39 | 10.99 | 37.64 |

| Site | Date Collected | Rainfall (RF) | Runoff | Acid Anions | | | | Base Cations | | | |
|------------------------|----------------|----------------------|--------------------------|-------------|---------------------------------------|---------------------------|---------------------------------------|--------------|------------|--------------------------|--------------------------|
| | | Rainfall (RF) (days) | Runoff (m ³) | Cl- (g/day) | SO ₄ ²⁻ (g/day) | NO ₃ - (g/day) | PO ₄ ³⁻ (g/day) | Na+ (g/day) | K+ (g/day) | Mg ²⁺ (g/day) | Ca ²⁺ (g/day) |
| 9. Ocoee 3 | | | | | | | | | | | |
| O3 | 7/1/2014 | 3 | 3.23 | 6.17 | 86.79 | 14.69 | 11.27 | 36.05 | 2.43 | 2.93 | 30.58 |
| O3 | 7/17/2014 | 6 | 0.81 | 0.86 | 12.74 | 1.91 | 1.58 | 4.34 | 0.34 | 0.36 | 3.15 |
| O3 | 7/20/2014 | 2 | 0.66 | 2.17 | 24.85 | 4.50 | 3.94 | 7.81 | 0.61 | 1.00 | 7.28 |
| O3 | 9/28/2014 | 1 | 0.16 | 0.87 | 12.03 | 1.95 | 1.54 | 3.62 | 0.23 | 0.40 | 3.24 |
| O3 | 11/2/2014 | 4 | 0.09 | 0.12 | 1.88 | 0.28 | 0.22 | 0.52 | 0.05 | 0.06 | 0.46 |
| O3 | 1/3/2015 | 1 | 0.02 | 0.20 | 2.96 | 0.53 | 0.33 | 0.94 | 0.06 | 0.11 | 0.77 |
| O3 | 1/22/2015 | 1 | 0.13 | 1.15 | 17.76 | 2.55 | 1.77 | 4.33 | 0.44 | 0.56 | 4.96 |
| O3 | 1/25/2015 | 2 | 0.53 | 2.08 | 37.13 | 5.06 | 4.49 | 16.29 | 0.97 | 1.16 | 8.42 |
| O3 | 2/10/2015 | 1 | 0.00 | 0.01 | 0.09 | 0.01 | 0.01 | 0.03 | 0.00 | 0.00 | 0.02 |
| O3 | 4/8/2015 | 1 | 0.23 | 1.36 | 20.35 | 3.25 | 2.26 | 0.00 | 0.22 | 0.33 | 4.47 |
| O3 | 5/18/2015 | 4 | 1.79 | 2.83 | 43.11 | 6.13 | 5.08 | 0.00 | 0.48 | 0.85 | 12.08 |
| O3 | 6/9/2015 | 1 | 1.24 | 8.17 | 80.44 | 16.51 | 15.39 | 50.40 | 1.32 | 2.30 | 34.36 |
| O3 | 7/6/2015 | 4 | 7.41 | 11.88 | 178.76 | 27.03 | 20.19 | 65.38 | 5.43 | 6.14 | 41.61 |
| O3 | 8/8/2015 | 2 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 10. Seiverville | | | | | | | | | | | |
| S1 | 7/17/2014 | 3 | 0.38 | 0.77 | 15.41 | 1.62 | 1.24 | 2.17 | 0.36 | 3.19 | 6.90 |
| S1 | 7/21/2014 | 2 | 0.76 | 2.42 | 44.29 | 5.24 | 4.44 | 5.13 | 1.09 | 9.81 | 20.13 |
| S1 | 7/29/2014 | 1 | 2.12 | 12.71 | 247.08 | 23.20 | 23.42 | 36.48 | 6.05 | 53.76 | 113.49 |
| S1 | 8/10/2014 | 3 | 0.72 | 1.59 | 27.50 | 2.49 | 2.65 | 3.37 | 0.70 | 6.19 | 12.31 |
| S1 | 8/30/2014 | 6 | 0.24 | 0.23 | 4.07 | 0.39 | 0.37 | 0.50 | 0.10 | 0.92 | 1.91 |
| S1 | 11/16/2014 | 1 | 0.22 | 1.27 | 23.31 | 2.51 | 2.15 | 2.98 | 0.58 | 5.23 | 10.85 |
| S1 | 1/22/2015 | 2 | 0.18 | 0.96 | 20.49 | 1.30 | 1.30 | 2.24 | 0.34 | 3.21 | 7.04 |
| S1 | 1/29/2015 | 4 | 0.45 | 1.18 | 23.88 | 2.06 | 1.48 | 1.86 | 0.49 | 4.38 | 9.10 |
| S1 | 4/16/2015 | 2 | 0.52 | 1.70 | 32.48 | 2.89 | 2.86 | 0.00 | 0.75 | 6.83 | 14.70 |
| S1 | 5/18/2015 | 3 | 0.12 | 0.25 | 4.86 | 0.55 | 0.40 | 0.00 | 0.11 | 1.00 | 2.11 |
| S1 | 6/9/2015 | 1 | 0.34 | 2.31 | 36.86 | 4.94 | 4.18 | 7.60 | 0.94 | 9.13 | 19.77 |
| S1 | 7/6/2015 | 4 | 0.81 | 1.28 | 24.74 | 2.36 | 2.15 | 2.96 | 0.58 | 5.22 | 10.82 |

APPENDIX F: ROAD CUT RUNOFF MASS LOADINGS SUMMARY SHEET

| Dissolved Metals | | | | | | | | | | |
|-------------------------|-----------------------|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------|
| Site | Date Collected | Al (g/day) | Cu (g/day) | Fe (g/day) | Mn (g/day) | Si (g/day) | Zn (g/day) | Ba (g/day) | Cd (g/day) | Ni (g/day) |
| 1. Grainger | | | | | | | | | | |
| G1 | 7/21/2014 | 0.57 | 0.04 | 0.15 | 0.46 | 3.42 | 0.29 | 0.07 | 0.01 | 0.06 |
| G1 | 10/10/2014 | 1.00 | 0.09 | 0.20 | 0.97 | 7.15 | 0.57 | 0.16 | 0.03 | 0.12 |
| G1 | 10/23/2014 | 0.51 | 0.03 | 0.11 | 0.42 | 2.99 | 0.25 | 0.07 | 0.01 | 0.05 |
| G1 | 1/29/2015 | 0.18 | 0.01 | 0.04 | 0.14 | 1.12 | 0.10 | 0.02 | 0.00 | 0.02 |
| G1 | 4/16/2015 | 0.12 | 0.03 | 0.02 | 0.19 | 2.47 | 0.09 | 0.04 | 0.00 | 0.04 |
| G1 | 5/18/2015 | 0.01 | 0.00 | 0.00 | 0.01 | 0.14 | 0.00 | 0.00 | 0.00 | 0.00 |
| G1 | 6/9/2015 | 0.41 | 0.04 | 0.07 | 0.37 | 4.18 | 0.18 | 0.08 | 0.02 | 0.06 |
| G1 | 7/6/2015 | 0.42 | 0.04 | 0.10 | 0.36 | 2.64 | 0.22 | 0.06 | 0.01 | 0.05 |
| 2. Jamestown 1 | | | | | | | | | | |
| J1 | 7/22/2014 | 0.22 | 0.04 | 0.10 | 0.17 | 0.30 | 0.14 | 0.04 | 0.01 | 0.01 |
| J1 | 7/31/2014 | 2.49 | 0.40 | 1.29 | 1.98 | 3.37 | 1.62 | 0.38 | 0.02 | 0.10 |
| J1 | 9/15/2014 | 0.34 | 0.06 | 0.11 | 0.37 | 0.64 | 0.30 | 0.07 | 0.01 | 0.02 |
| J1 | 12/10/2014 | 0.42 | 0.09 | 0.17 | 0.49 | 0.78 | 0.41 | 0.10 | 0.01 | 0.03 |
| J1 | 1/2/2015 | 0.08 | 0.01 | 0.03 | 0.09 | 0.15 | 0.08 | 0.02 | 0.00 | 0.01 |
| J1 | 1/26/2015 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| J1 | 4/16/2015 | 0.00 | 0.02 | 0.02 | 0.00 | 0.41 | 0.01 | 0.03 | 0.01 | 0.00 |
| J1 | 6/20/2015 | 0.37 | 0.03 | 0.04 | 0.02 | 0.72 | 0.05 | 0.04 | 0.01 | 0.01 |
| J1 | 7/24/2015 | 0.49 | 0.07 | 0.24 | 0.37 | 0.62 | 0.29 | 0.07 | 0.00 | 0.03 |
| 3. Jamestown 2 | | | | | | | | | | |
| J2 | 7/15/2014 | 1.33 | 0.29 | 3.30 | 4.01 | 13.33 | 0.23 | 0.51 | 0.11 | 0.30 |
| J2 | 9/15/2014 | 0.72 | 0.31 | 0.16 | 2.39 | 8.57 | 0.17 | 0.29 | 0.10 | 0.19 |
| J2 | 12/10/2014 | 0.67 | 0.33 | 0.18 | 4.86 | 9.70 | 0.15 | 0.44 | 0.18 | 0.19 |
| J2 | 1/2/2015 | 0.18 | 0.06 | 0.02 | 0.87 | 1.53 | 0.05 | 0.09 | 0.03 | 0.04 |
| J2 | 1/26/2015 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| J2 | 4/16/2015 | 0.44 | 0.12 | 0.08 | 1.64 | 5.43 | 0.08 | 0.18 | 0.03 | 0.05 |
| J2 | 6/20/2015 | 1.33 | 0.07 | 2.56 | 2.01 | 6.48 | 0.11 | 0.24 | 0.07 | 0.20 |

| Dissolved Metals | | | | | | | | | | |
|-------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | Date | Al | Cu | Fe | Mn | Si | Zn | Ba | Cd | Ni |
| Site | Collected | (g/day) | (g/day) | (g/day) | (g/day) | (g/day) | (g/day) | (g/day) | (g/day) | (g/day) |
| 4. Jamestown 3 | | | | | | | | | | |
| J3 | 7/4/2014 | 0.00 | 0.00 | 0.00 | 0.00 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 |
| J3 | 7/15/2014 | 0.02 | 0.00 | 0.01 | 0.02 | 0.47 | 0.01 | 0.01 | 0.00 | 0.00 |
| J3 | 7/22/2014 | 0.03 | 0.00 | 0.01 | 0.02 | 0.62 | 0.02 | 0.01 | 0.00 | 0.00 |
| J3 | 7/31/2014 | 0.22 | 0.04 | 0.10 | 0.16 | 4.98 | 0.16 | 0.09 | 0.02 | 0.00 |
| J3 | 9/15/2014 | 0.06 | 0.01 | 0.02 | 0.04 | 1.20 | 0.04 | 0.03 | 0.01 | 0.00 |
| J3 | 12/10/2014 | 0.07 | 0.01 | 0.03 | 0.05 | 1.32 | 0.05 | 0.03 | 0.01 | 0.00 |
| J3 | 1/2/2015 | 0.02 | 0.00 | 0.01 | 0.01 | 0.30 | 0.01 | 0.01 | 0.00 | 0.00 |
| J3 | 1/22/2015 | 0.01 | 0.00 | 0.01 | 0.01 | 0.27 | 0.01 | 0.01 | 0.00 | 0.00 |
| J3 | 1/26/2015 | 0.00 | 0.00 | 0.00 | 0.00 | 0.07 | 0.00 | 0.00 | 0.00 | 0.00 |
| J3 | 4/16/2015 | 0.00 | 0.00 | 0.00 | 0.00 | 0.44 | 0.01 | 0.01 | 0.00 | 0.00 |
| J3 | 6/20/2015 | 0.03 | 0.00 | 0.02 | 0.00 | 0.98 | 0.01 | 0.02 | 0.00 | 0.00 |
| J3 | 7/24/2015 | 0.06 | 0.01 | 0.02 | 0.04 | 1.14 | 0.04 | 0.02 | 0.00 | 0.00 |
| 5. Nashville 1 | | | | | | | | | | |
| N1 | 6/8/2014 | 1.99 | 0.34 | 0.36 | 1.60 | 4.47 | 1.46 | 0.30 | 0.07 | 0.20 |
| N1 | 7/1/2014 | 0.68 | 0.11 | 0.12 | 0.53 | 1.45 | 0.45 | 0.11 | 0.03 | 0.06 |
| N1 | 7/22/2014 | 8.51 | 1.33 | 1.52 | 6.67 | 18.88 | 5.87 | 1.43 | 0.25 | 0.81 |
| N1 | 8/26/2014 | 0.23 | 0.05 | 0.05 | 0.24 | 0.67 | 0.21 | 0.05 | 0.02 | 0.02 |
| N1 | 10/11/2014 | 2.23 | 0.48 | 0.40 | 2.23 | 6.34 | 1.94 | 0.42 | 0.00 | 0.15 |
| N1 | 1/9/2015 | 1.19 | 0.31 | 0.26 | 1.29 | 3.64 | 1.14 | 0.26 | 0.07 | 0.12 |
| N1 | 1/30/2015 | 0.31 | 0.07 | 0.06 | 0.34 | 0.94 | 0.30 | 0.07 | 0.02 | 0.04 |
| N1 | 3/23/2015 | 0.00 | 0.02 | 0.01 | 0.07 | 0.74 | 0.05 | 0.02 | 0.01 | 0.04 |
| N1 | 6/1/2015 | 0.38 | 0.20 | 0.08 | 0.89 | 8.86 | 0.24 | 0.34 | 0.13 | 0.39 |
| N1 | 7/24/2015 | 1.26 | 0.22 | 0.24 | 1.03 | 2.92 | 0.88 | 0.21 | 0.04 | 0.09 |
| N1 | 8/21/2015 | 1.67 | 0.32 | 0.30 | 1.39 | 3.67 | 1.15 | 0.27 | 0.08 | 0.13 |
| 6. Nashville 2 | | | | | | | | | | |
| N2 | 7/1/2014 | 7.53 | 0.17 | 0.39 | 0.88 | 3.19 | 1.40 | 0.07 | 0.02 | 0.33 |
| N2 | 8/26/2014 | 2.89 | 0.07 | 0.15 | 0.42 | 1.53 | 0.62 | 0.04 | 0.01 | 0.14 |
| N2 | 10/11/2014 | 27.51 | 0.65 | 2.34 | 4.11 | 14.34 | 5.96 | 0.35 | 0.11 | 1.50 |
| N2 | 1/9/2015 | 16.91 | 0.42 | 0.58 | 2.19 | 8.43 | 3.42 | 0.22 | 0.09 | 0.80 |
| N2 | 3/23/2015 | 0.00 | 0.01 | 0.01 | 0.06 | 0.31 | 0.01 | 0.02 | 0.00 | 0.01 |
| N2 | 6/1/2015 | 28.89 | 0.69 | 0.13 | 2.62 | 16.26 | 5.31 | 0.25 | 0.05 | 1.71 |
| N2 | 7/24/2015 | 15.17 | 0.32 | 0.23 | 1.74 | 6.23 | 2.58 | 0.14 | 0.04 | 0.58 |

| | | | | | | | | | | |
|----|-----------|-------|------|------|------|------|------|------|------|------|
| N2 | 8/21/2015 | 19.08 | 0.42 | 0.95 | 2.39 | 8.88 | 3.71 | 0.20 | 0.03 | 0.88 |
|----|-----------|-------|------|------|------|------|------|------|------|------|

7

Dissolved Metals

| Site | Date Collected | Al (g/day) | Cu (g/day) | Fe (g/day) | Mn (g/day) | Si (g/day) | Zn (g/day) | Ba (g/day) | Cd (g/day) | Ni (g/day) |
|-------------------|----------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|
| 7. Ocoee 1 | | | | | | | | | | |
| O1 | 6/12/2014 | 0.06 | 0.01 | 0.01 | 0.00 | 1.56 | 0.01 | 0.02 | 0.00 | 0.00 |
| O1 | 6/19/2014 | 0.01 | 0.00 | 0.00 | 0.00 | 0.21 | 0.00 | 0.00 | 0.00 | 0.00 |
| O1 | 6/26/2014 | 0.08 | 0.01 | 0.01 | 0.01 | 2.12 | 0.01 | 0.02 | 0.00 | 0.00 |
| O1 | 7/1/2014 | 0.20 | 0.03 | 0.03 | 0.01 | 5.49 | 0.03 | 0.06 | 0.01 | 0.00 |
| O1 | 7/17/2014 | 0.02 | 0.00 | 0.00 | 0.00 | 0.44 | 0.00 | 0.00 | 0.00 | 0.00 |
| O1 | 7/20/2014 | 0.08 | 0.01 | 0.01 | 0.00 | 1.90 | 0.01 | 0.02 | 0.00 | 0.00 |
| O1 | 9/28/2014 | 0.02 | 0.00 | 0.00 | 0.00 | 0.63 | 0.00 | 0.01 | 0.00 | 0.00 |
| O1 | 11/2/2014 | 0.00 | 0.00 | 0.00 | 0.00 | 0.06 | 0.00 | 0.00 | 0.00 | 0.00 |
| O1 | 1/3/2015 | 0.00 | 0.00 | 0.00 | 0.00 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 |
| O1 | 1/25/2015 | 0.06 | 0.01 | 0.01 | 0.00 | 1.60 | 0.01 | 0.02 | 0.00 | 0.00 |
| O1 | 2/10/2015 | 0.03 | 0.00 | 0.00 | 0.00 | 0.85 | 0.01 | 0.01 | 0.00 | 0.00 |
| O1 | 4/8/2015 | 0.00 | 0.01 | 0.01 | 0.00 | 1.07 | 0.00 | 0.02 | 0.00 | 0.00 |
| O1 | 5/18/2015 | 0.00 | 0.01 | 0.00 | 0.00 | 1.32 | 0.01 | 0.02 | 0.00 | 0.00 |
| O1 | 6/9/2015 | 0.16 | 0.02 | 0.01 | 0.00 | 5.01 | 0.04 | 0.04 | 0.01 | 0.00 |
| O1 | 7/6/2015 | 0.39 | 0.06 | 0.06 | 0.02 | 9.96 | 0.07 | 0.11 | 0.02 | 0.01 |
| 8. Ocoee 2 | | | | | | | | | | |
| O2 | 6/26/2014 | 0.10 | 0.02 | 0.02 | 0.03 | 0.53 | 0.05 | 0.02 | 0.00 | 0.00 |
| O2 | 7/1/2014 | 0.16 | 0.05 | 0.06 | 0.11 | 1.58 | 0.10 | 0.05 | 0.02 | 0.00 |
| O2 | 7/17/2014 | 0.00 | 0.00 | 0.00 | 0.00 | 0.11 | 0.00 | 0.00 | 0.00 | 0.00 |
| O2 | 7/20/2014 | 0.02 | 0.01 | 0.00 | 0.00 | 0.54 | 0.01 | 0.02 | 0.00 | 0.00 |
| O2 | 9/28/2014 | 0.02 | 0.00 | 0.00 | 0.02 | 0.16 | 0.01 | 0.01 | 0.00 | 0.00 |
| O2 | 11/2/2014 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 |
| O2 | 1/3/2015 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| O2 | 1/25/2015 | 0.03 | 0.01 | 0.01 | 0.06 | 0.42 | 0.04 | 0.01 | 0.00 | 0.00 |
| O2 | 2/10/2015 | 0.01 | 0.00 | 0.00 | 0.02 | 0.22 | 0.02 | 0.01 | 0.00 | 0.00 |
| O2 | 4/8/2015 | 0.06 | 0.01 | 0.02 | 0.06 | 0.50 | 0.04 | 0.02 | 0.00 | 0.00 |
| O2 | 5/18/2015 | 0.07 | 0.01 | 0.03 | 0.03 | 0.53 | 0.06 | 0.02 | 0.01 | 0.00 |
| O2 | 6/9/2015 | 0.01 | 0.04 | 0.07 | 0.22 | 1.99 | 0.03 | 0.06 | 0.01 | 0.01 |
| O2 | 7/6/2015 | 0.65 | 0.13 | 0.14 | 0.51 | 2.95 | 0.46 | 0.12 | 0.05 | 0.02 |

| Dissolved Metals | | | | | | | | | | |
|-------------------------|-----------------------|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------|
| Site | Date Collected | Al (g/day) | Cu (g/day) | Fe (g/day) | Mn (g/day) | Si (g/day) | Zn (g/day) | Ba (g/day) | Cd (g/day) | Ni (g/day) |
| 9. Ocoee 3 | | | | | | | | | | |
| O3 | 7/1/2014 | 5.52 | 0.98 | 0.76 | 3.81 | 5.64 | 2.94 | 0.46 | 0.16 | 0.11 |
| O3 | 7/17/2014 | 0.84 | 0.08 | 0.09 | 0.43 | 0.76 | 0.38 | 0.06 | 0.01 | 0.00 |
| O3 | 7/20/2014 | 1.65 | 0.21 | 0.24 | 1.09 | 1.35 | 0.85 | 0.15 | 0.03 | 0.06 |
| O3 | 9/28/2014 | 0.71 | 0.09 | 0.10 | 0.45 | 0.79 | 0.37 | 0.06 | 0.01 | 0.01 |
| O3 | 11/2/2014 | 0.12 | 0.01 | 0.01 | 0.06 | 0.11 | 0.05 | 0.01 | 0.00 | 0.00 |
| O3 | 1/3/2015 | 0.20 | 0.02 | 0.02 | 0.07 | 0.16 | 0.07 | 0.01 | 0.01 | 0.00 |
| O3 | 1/22/2015 | 0.89 | 0.12 | 0.13 | 0.51 | 0.63 | 0.52 | 0.09 | 0.04 | 0.02 |
| O3 | 1/25/2015 | 1.74 | 0.30 | 0.27 | 1.24 | 2.86 | 1.02 | 0.14 | 0.07 | 0.05 |
| O3 | 2/10/2015 | 0.01 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 |
| O3 | 4/8/2015 | 0.65 | 0.06 | 0.17 | 0.16 | 1.33 | 0.08 | 0.08 | 0.04 | 0.02 |
| O3 | 5/18/2015 | 1.52 | 0.11 | 0.09 | 0.33 | 3.83 | 0.21 | 0.17 | 0.04 | 0.06 |
| O3 | 6/9/2015 | 5.70 | 0.44 | 0.18 | 0.90 | 9.65 | 0.62 | 0.42 | 0.00 | 0.23 |
| O3 | 7/6/2015 | 6.48 | 1.27 | 1.17 | 5.68 | 5.53 | 4.90 | 0.80 | 0.34 | 0.05 |
| O3 | 8/8/2015 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 10. Seiverville | | | | | | | | | | |
| S1 | 7/17/2014 | 0.29 | 0.04 | 0.05 | 0.23 | 0.34 | 0.19 | 0.04 | 0.02 | 0.00 |
| S1 | 7/21/2014 | 0.87 | 0.12 | 0.15 | 0.68 | 0.95 | 0.58 | 0.14 | 0.06 | 0.02 |
| S1 | 7/29/2014 | 4.86 | 0.62 | 0.85 | 3.99 | 5.55 | 3.28 | 0.72 | 0.33 | 0.05 |
| S1 | 8/10/2014 | 0.53 | 0.07 | 0.09 | 0.45 | 0.58 | 0.35 | 0.08 | 0.02 | 0.01 |
| S1 | 8/30/2014 | 0.08 | 0.01 | 0.01 | 0.07 | 0.09 | 0.05 | 0.01 | 0.00 | 0.00 |
| S1 | 11/16/2014 | 0.44 | 0.06 | 0.07 | 0.38 | 0.53 | 0.30 | 0.07 | 0.02 | 0.01 |
| S1 | 1/22/2015 | 0.21 | 0.05 | 0.05 | 0.19 | 0.39 | 0.18 | 0.05 | 0.02 | 0.01 |
| S1 | 1/29/2015 | 0.33 | 0.06 | 0.07 | 0.33 | 0.36 | 0.24 | 0.06 | 0.03 | 0.00 |
| S1 | 4/16/2015 | 0.00 | 0.04 | 0.02 | 0.09 | 0.96 | 0.06 | 0.07 | 0.01 | 0.02 |
| S1 | 5/18/2015 | 0.00 | 0.01 | 0.00 | 0.01 | 0.14 | 0.01 | 0.01 | 0.00 | 0.00 |
| S1 | 6/9/2015 | 0.29 | 0.06 | 0.03 | 0.13 | 0.99 | 0.09 | 0.07 | 0.01 | 0.00 |
| S1 | 7/6/2015 | 0.44 | 0.07 | 0.07 | 0.38 | 0.52 | 0.30 | 0.07 | 0.03 | 0.01 |

ROAD CUT RUNOFF WATER CHEMISTRY: SUMMARY SHEET
MASS LOADINGS

Units: g/day

| Chemical Parameter | SITE | | | | | | | | | | | |
|-----------------------|------------|-------|------|--------|-------------|--------|------|--------|-------------|--------|------|---------|
| | Grainger 1 | | | | Jamestown 1 | | | | Jamestown 2 | | | |
| | Mean | StDev | Min | Max | Mean | StDev | Min | Max | Mean | StDev | Min | Max |
| Cl- | 3.05 | 2.16 | 0.16 | 7.52 | 7.67 | 10.56 | 0.00 | 34.80 | 22.85 | 17.32 | 0.00 | 48.51 |
| SO4 ²⁻ | 45.45 | 29.56 | 2.31 | 102.28 | 98.42 | 155.76 | 0.00 | 504.95 | 564.19 | 446.16 | 0.00 | 1399.61 |
| NO3- | 7.56 | 5.14 | 0.37 | 17.77 | 19.23 | 28.23 | 0.00 | 92.42 | 44.65 | 30.53 | 0.00 | 92.42 |
| PO4 ³⁻ | 6.42 | 4.22 | 0.34 | 14.48 | 15.85 | 22.39 | 0.00 | 73.58 | 40.75 | 26.34 | 0.00 | 73.58 |
| Na+ | 2.78 | 2.31 | 0.02 | 7.21 | 4.74 | 7.21 | 0.00 | 23.30 | 8.79 | 12.14 | 0.00 | 24.24 |
| K+ | 0.94 | 0.66 | 0.04 | 2.27 | 8.30 | 11.75 | 0.00 | 38.64 | 7.60 | 12.70 | 0.00 | 38.64 |
| Mg ²⁺ | 4.89 | 3.47 | 0.20 | 11.46 | 3.53 | 4.55 | 0.00 | 15.12 | 15.01 | 10.01 | 0.00 | 29.87 |
| Ca ²⁺ | 4.74 | 3.54 | 0.19 | 11.86 | 14.18 | 17.68 | 0.00 | 59.05 | 26.55 | 19.91 | 0.00 | 59.05 |
| Al | 0.40 | 0.31 | 0.01 | 1.00 | 0.49 | 0.77 | 0.00 | 2.49 | 0.89 | 0.80 | 0.00 | 2.49 |
| Cu | 0.03 | 0.03 | 0.00 | 0.09 | 0.08 | 0.12 | 0.00 | 0.40 | 0.20 | 0.15 | 0.00 | 0.40 |
| Fe | 0.09 | 0.07 | 0.00 | 0.20 | 0.22 | 0.41 | 0.00 | 1.29 | 0.95 | 1.31 | 0.00 | 3.30 |
| Mn | 0.36 | 0.29 | 0.01 | 0.97 | 0.39 | 0.62 | 0.00 | 1.98 | 2.22 | 1.58 | 0.00 | 4.86 |
| Si | 3.02 | 2.10 | 0.14 | 7.15 | 0.78 | 1.01 | 0.00 | 3.37 | 6.05 | 4.43 | 0.00 | 13.33 |
| Zn | 0.21 | 0.17 | 0.00 | 0.57 | 0.32 | 0.51 | 0.00 | 1.62 | 0.30 | 0.54 | 0.00 | 1.62 |
| Ba | 0.06 | 0.05 | 0.00 | 0.16 | 0.08 | 0.12 | 0.00 | 0.38 | 0.27 | 0.18 | 0.00 | 0.51 |
| Cd | 0.01 | 0.01 | 0.00 | 0.03 | 0.01 | 0.01 | 0.00 | 0.02 | 0.07 | 0.06 | 0.00 | 0.18 |
| Ni | 0.05 | 0.04 | 0.00 | 0.12 | 0.02 | 0.03 | 0.00 | 0.10 | 0.13 | 0.10 | 0.00 | 0.30 |

ROAD CUT RUNOFF WATER CHEMISTRY: SUMMARY SHEET
MASS LOADINGS

Units: g/day

| Chemical Parameter | SITE | | | | | | | | | | | |
|-----------------------|-------------|-------|------|-------|-------------|---------|--------|---------|-------------|--------|-------|--------|
| | Jamestown 3 | | | | Nashville 1 | | | | Nashville 2 | | | |
| | Mean | StDev | Min | Max | Mean | StDev | Min | Max | Mean | StDev | Min | Max |
| Cl- | 1.90 | 2.54 | 0.03 | 9.39 | 9.45 | 11.57 | 0.84 | 41.99 | 4.90 | 3.62 | 0.80 | 10.50 |
| SO4 ²⁻ | 6.38 | 7.83 | 0.11 | 29.62 | 1345.34 | 1619.13 | 150.11 | 5855.11 | 268.45 | 219.36 | 37.21 | 601.20 |
| NO3- | 3.72 | 4.51 | 0.10 | 16.45 | 18.56 | 23.56 | 1.91 | 85.24 | 11.72 | 11.04 | 1.67 | 31.82 |
| PO4 ³⁻ | 2.53 | 2.76 | 0.05 | 10.17 | 12.73 | 19.78 | 0.56 | 69.95 | 9.16 | 5.62 | 1.14 | 16.17 |
| Na+ | 0.65 | 0.91 | 0.01 | 3.35 | 15.66 | 17.83 | 0.06 | 64.51 | 23.76 | 34.49 | 0.00 | 104.29 |
| K + | 0.56 | 0.71 | 0.01 | 2.66 | 0.42 | 0.46 | 0.04 | 1.69 | 0.88 | 0.79 | 0.08 | 2.44 |
| Mg ²⁺ | 0.75 | 0.96 | 0.02 | 3.57 | 101.14 | 111.62 | 11.29 | 394.09 | 44.03 | 38.31 | 6.19 | 109.84 |
| Ca ²⁺ | 2.88 | 3.50 | 0.06 | 13.01 | 305.94 | 336.99 | 39.05 | 1184.65 | 185.84 | 142.89 | 30.20 | 417.12 |
| Al | 0.04 | 0.06 | 0.00 | 0.22 | 1.68 | 2.39 | 0.00 | 8.51 | 14.75 | 10.67 | 0.00 | 28.89 |
| Cu | 0.01 | 0.01 | 0.00 | 0.04 | 0.31 | 0.37 | 0.02 | 1.33 | 0.34 | 0.25 | 0.01 | 0.69 |
| Fe | 0.02 | 0.03 | 0.00 | 0.10 | 0.31 | 0.42 | 0.01 | 1.52 | 0.60 | 0.77 | 0.01 | 2.34 |
| Mn | 0.03 | 0.04 | 0.00 | 0.16 | 1.48 | 1.84 | 0.07 | 6.67 | 1.80 | 1.32 | 0.06 | 4.11 |
| Si | 0.98 | 1.33 | 0.02 | 4.98 | 4.78 | 5.32 | 0.67 | 18.88 | 7.40 | 5.78 | 0.31 | 16.26 |
| Zn | 0.03 | 0.04 | 0.00 | 0.16 | 1.24 | 1.64 | 0.05 | 5.87 | 2.88 | 2.14 | 0.01 | 5.96 |
| Ba | 0.02 | 0.02 | 0.00 | 0.09 | 0.32 | 0.39 | 0.02 | 1.43 | 0.16 | 0.12 | 0.02 | 0.35 |
| Cd | 0.00 | 0.00 | 0.00 | 0.02 | 0.06 | 0.07 | 0.00 | 0.25 | 0.04 | 0.04 | 0.00 | 0.11 |
| Ni | 0.00 | 0.00 | 0.00 | 0.00 | 0.19 | 0.23 | 0.02 | 0.81 | 0.74 | 0.61 | 0.01 | 1.71 |

ROAD CUT RUNOFF WATER CHEMISTRY: SUMMARY SHEET
MASS LOADINGS

Units: g/day

| Chemical Parameter | SITE | | | | | | | | | | | |
|-----------------------|---------|-------|------|-------|---------|-------|------|-------|---------|-------|------|--------|
| | Ocoee 1 | | | | Ocoee 2 | | | | Ocoee 3 | | | |
| | Mean | StDev | Min | Max | Mean | StDev | Min | Max | Mean | StDev | Min | Max |
| Cl- | 1.56 | 1.96 | 0.01 | 7.11 | 1.66 | 1.89 | 0.01 | 6.53 | 2.71 | 3.56 | 0.00 | 11.88 |
| SO4 ²⁻ | 12.16 | 15.98 | 0.09 | 57.86 | 13.20 | 14.63 | 0.09 | 49.95 | 37.06 | 49.33 | 0.01 | 178.76 |
| NO3- | 6.61 | 9.39 | 0.03 | 37.45 | 3.40 | 3.95 | 0.02 | 13.84 | 6.03 | 7.93 | 0.00 | 27.03 |
| PO4 ³⁻ | 2.88 | 3.78 | 0.01 | 13.61 | 2.67 | 3.06 | 0.02 | 10.92 | 4.86 | 6.30 | 0.00 | 20.19 |
| Na+ | 3.25 | 4.49 | 0.03 | 15.94 | 1.75 | 2.64 | 0.01 | 9.78 | 13.55 | 21.34 | 0.00 | 65.38 |
| K + | 0.09 | 0.12 | 0.00 | 0.42 | 0.31 | 0.38 | 0.00 | 1.39 | 0.90 | 1.46 | 0.00 | 5.43 |
| Mg ²⁺ | 4.64 | 5.74 | 0.04 | 20.84 | 2.69 | 3.23 | 0.02 | 10.99 | 1.16 | 1.68 | 0.00 | 6.14 |
| Ca ²⁺ | 12.62 | 15.39 | 0.09 | 57.02 | 8.56 | 10.46 | 0.05 | 37.64 | 10.81 | 13.99 | 0.00 | 41.61 |
| Al | 0.08 | 0.11 | 0.00 | 0.39 | 0.09 | 0.17 | 0.00 | 0.65 | 1.86 | 2.27 | 0.00 | 6.48 |
| Cu | 0.01 | 0.02 | 0.00 | 0.06 | 0.02 | 0.04 | 0.00 | 0.13 | 0.26 | 0.39 | 0.00 | 1.27 |
| Fe | 0.01 | 0.02 | 0.00 | 0.06 | 0.03 | 0.04 | 0.00 | 0.14 | 0.23 | 0.33 | 0.00 | 1.17 |
| Mn | 0.00 | 0.01 | 0.00 | 0.02 | 0.08 | 0.14 | 0.00 | 0.51 | 1.05 | 1.65 | 0.00 | 5.68 |
| Si | 2.15 | 2.71 | 0.02 | 9.96 | 0.74 | 0.89 | 0.00 | 2.95 | 2.33 | 2.88 | 0.00 | 9.65 |
| Zn | 0.01 | 0.02 | 0.00 | 0.07 | 0.06 | 0.12 | 0.00 | 0.46 | 0.86 | 1.39 | 0.00 | 4.90 |
| Ba | 0.02 | 0.03 | 0.00 | 0.11 | 0.03 | 0.03 | 0.00 | 0.12 | 0.18 | 0.23 | 0.00 | 0.80 |
| Cd | 0.00 | 0.01 | 0.00 | 0.02 | 0.01 | 0.01 | 0.00 | 0.05 | 0.05 | 0.09 | 0.00 | 0.34 |
| Ni | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.01 | 0.00 | 0.02 | 0.04 | 0.06 | 0.00 | 0.23 |

ROAD CUT RUNOFF WATER CHEMISTRY: SUMMARY SHEET
MASS LOADINGS

Units: g/day

| Chemical Parameter | SITE | | | |
|-----------------------|---------------|-------|------|--------|
| | Seiverville 1 | | | |
| | Mean | StDev | Min | Max |
| Cl- | 2.96 | 4.20 | 0.23 | 12.71 |
| SO4 ²⁻ | 52.59 | 73.37 | 4.07 | 247.08 |
| NO3- | 5.89 | 8.68 | 0.39 | 27.03 |
| PO4 ³⁻ | 5.14 | 7.53 | 0.37 | 23.42 |
| Na+ | 10.05 | 19.19 | 0.00 | 65.38 |
| K + | 1.35 | 1.97 | 0.10 | 6.05 |
| Mg ²⁺ | 8.85 | 13.76 | 0.92 | 53.76 |
| Ca ²⁺ | 20.82 | 29.66 | 1.91 | 113.49 |
| Al | 1.14 | 2.05 | 0.00 | 6.48 |
| Cu | 0.19 | 0.36 | 0.01 | 1.27 |
| Fe | 0.20 | 0.37 | 0.00 | 1.17 |
| Mn | 0.97 | 1.76 | 0.01 | 5.68 |
| Si | 1.62 | 2.80 | 0.09 | 9.65 |
| Zn | 0.81 | 1.50 | 0.01 | 4.90 |
| Ba | 0.17 | 0.27 | 0.01 | 0.80 |
| Cd | 0.07 | 0.12 | 0.00 | 0.34 |
| Ni | 0.03 | 0.06 | 0.00 | 0.23 |

SUMMARY STATS FOR MASS LOADINGS

| | | Chemical Parameter | overall median |
|------|---------|-----------------------|-------------------|
| min | max | | |
| 0.00 | 48.51 | Cl- | 1.68 |
| 0.00 | 5855.11 | SO4 ²⁻ | 24.85 |
| 0.00 | 92.42 | NO3- | 4.50 |
| 0.00 | 73.58 | PO4 ³⁻ | 2.84 |
| 0.00 | 104.29 | Na+ | 2.04 |
| 0.00 | 38.64 | K + | 0.39 |
| 0.00 | 394.09 | Mg ²⁺ | 3.19 |
| 0.00 | 1184.65 | Ca ²⁺ | 8.42 |
| 0.00 | 28.89 | Al | 0.21 |
| 0.00 | 1.33 | Cu | 0.04 |
| 0.00 | 3.30 | Fe | 0.05 |
| 0.00 | 6.67 | Mn | 0.16 |
| 0.00 | 18.88 | Si | 0.98 |
| 0.00 | 5.96 | Zn | 0.08 |
| 0.00 | 1.43 | Ba | 0.05 |
| 0.00 | 0.34 | Cd | 0.01 |
| 0.00 | 1.71 | Ni | 0.01 |

Appendix G: EPT XTRAM® HPL Specifications



| Element | Rating* | | Element | | Rating* | | Element | | Rating* | |
|-----------------------------------|---------|-------|-----------------------------------|-----|---------|-----|---|-----|---------|-----|
| | EIA/KEE | PVC | EIA/KEE | PVC | EIA/KEE | PVC | EIA/KEE | PVC | EIA/KEE | PVC |
| Acetic Acid (5%) | A | C | Diesel for Locomotive (ULC) | | | | Nitric Acid (50%) | | | |
| Acetic Acid (50%) | C | C | Diethylene Sebacate | | X | X | Nitric Acid (100%) | | | C |
| Acetone | C | C | Dioctyl Phthalate (DOP) | | | C | Oxygenated Unleaded Gasoline containing Ehtanol (ULC) | | | T |
| Asphalt | T | B | Ethyl Acetate | | | C | Nitrobenzene | | | X |
| ASTM #1 Oil (ULC) | A | A | Ethyl Alcohol | | | A | Perchloroethylene | | | C |
| ASTM #2 (ULC) | A | A | Ethylene Dichloride | | X | X | Peroxide - 200 ppm | | | T |
| ASTM #3 Oil (ULC) | A | A | Ethylene Glycol (Anti-Freeze) | | | X | Phenol (50%) | | | C |
| ASTM Reference Fuel A (ULC) | A | C | Formaldehyde | | X | A | Phosphoric Acid (50%) | | | C |
| ASTM Reference Fuel B (ULC) | A | C | Fuel Ethanol (15%) (ULC) | | T | X | Phosphoric Acid (100%) | | | C |
| ASTM Reference Fuel C | B | X | Fuel Methanol (15%) (ULC) | | T | X | Phthalate Plasticizer | | | C |
| Ammonium Phosphate | T | X | Furfural | | | X | Potassium Chloride | | | T |
| Ammonium Sulfate | T | X | Gasoline (ULC) | | B | X | Potassium Sulphate | | | T |
| Aqua Regia | X | X | Gas Turbine Fuel Oils (ULC) | | T | X | Pydraul 312C | | | X |
| Automatic Trans. Fluid | T | T | Gear Oil | | T | T | Regular Sulphur Diesel Fuel (ULC) | | | A |
| Aviation Gasoline (ULC) | B | X | Glycerine | | B | A | SAE-30 Oil | | | A |
| Benzaldehyde | X | C | Home Heating Oil (ULC) | | A | X | Salt Water (25%) | | | B |
| Benzene | X | C | Hydraulic Fluid - Petroleum | | A | B | Sea Water | | | A |
| Bromine, Anhydrous Liquid | X | C | Hydraulic Fluid - Phosphate Ester | | C | C | Soap Solution (1%) | | | T |
| Butyl Acetate | X | B | Hydrocarbon Type II (40%Aromatic) | | C | C | Sodium Acetate Solution | | | T |
| Butyl Alcohol | T | C | Hydrochloric Acid (20%) | | A | A | Sodium Bisulfite Solution | | | T |
| Calcium Chloride (30%) | T | C | Hydrochloric Solution (12.5%) | | A | A | Sodium Hydroxide (40%) | | | A |
| Calcium Hydroxide Solutions | T | A | Isocetane | | A | A | Sodium Phosphate | | | T |
| Calcium Bisulfide | X | A | Isopropyl Alcohol | | T | T | Sulfuric Acid (50%) | | | A |
| Carbon Tetrachloride | X | C | Jet Fuel, JP-4 | | T | C | Sulfuric Acid (97%) | | | C |
| Caucstic Soda Liquid 50% | T | A | Lactic Acid | | B | B | Tannic Acid (50%) | | | T |
| Clorobenzene | X | C | Linseed Oil - Raw | | A | A | Tetrahydrofuran | | | X |
| Chloroform | X | C | Magnesium Chloride | | T | X | Transformer Oil | | | A |
| Chlorosulfonic Acid | X | C | Magnesium Hydroxide | | T | X | Tributyl Phosphate | | | X |
| Clorox/Bleach/Sodium Hypochlorite | A | A/15% | Methyl Alcohol | | A | A | Toluene | | | C |
| Coagulant | T | A | Methylene Chloride | | A | C | UAN | | | X |
| Crude Oil | A | B | Methyl Ethyl Ketone | | T | C | Water (70° F) | | | A |
| Cyclohexane | B | C | Mineral Oil | | T | T | Water (200° F) | | | A |
| Diesel Fuel (ULC) | A | X | Naphtha | | A | C | Xylene | | | C |
| Diesel Fuel Low-Sulphur (ULC) | T | X | Nitric Acid (10%) | | B | A | Zinc Chloride | | | T |

Rating Key: A - Fluid has little or no effect. B - Fluid has minor to moderate effect. C - Fluid has severe effect. T - No data - likely to be acceptable. X - No data - not likely to be acceptable. ULC - Meets the requirements of ULC S668.

* Ratings are based on visual and physical examinations of samples after removal from the test chemical after the samples of EPT Xtrm Ply HP 36 we immersed for 28 days at room temperature. Results are intended to represent ability of the material to retain its performance properties when in contact with the listed chemical. The data above was obtained on samples of the material under laboratory conditions. To the best of EPT's knowledge, this data is within the accuracy and precision of the respective tests. Because of testing and sampling variability, we cannot guarantee that other laboratories will obtain the same results and NO WARRANTY IS EXPRESSED OR IMPLIED.

September 2015

Appendix H: Water Chemistry for Middlebrook Pike Experiment: Simulated Rainfall

PYRITE GEOLOGY WATER CHEMISTRY: SUMMARY SHEET MIDDLEBROOK PIKE EXPERIMENT

Simulated Rainfall

| Sample/Tray Type | Date | Time (hrs) | uS/cm Conctvty | pH | measured | | | | | | | |
|--------------------------------|-----------|------------|-------------------|------|--------------|------------|-------------|-------------|------------|-----------|------------|--|
| | | | | | ueq/L ANC | mg/L Cl | mg/L NO3 | mg/L SO4 | mg/L Na | mg/L K | mg/L Mg | |
| GRASS = soil/vegetation | | | | | | | | | | | | |
| PY001-G01C 160427 GRASS | 23-May-16 | 0.000 | 1273.56 | 6.67 | 3031.3 | 250.13 | 34.68 | 144.19 | 10.06 | 215.91 | 19.28 | |
| PY002-G02C 160427 | 23-May-16 | 0.833 | 1354.24 | 6.52 | 2407.5 | 183.23 | 14.59 | 248.93 | 8.58 | 207.50 | 22.95 | |
| PY003-G03C 160427 | 23-May-16 | 1.667 | 1274.53 | 6.51 | 2020.7 | 155.12 | 14.68 | 295.08 | 7.59 | 206.47 | 19.93 | |
| PY004-G04C 160427 | 23-May-16 | 2.383 | 1124.84 | 6.63 | 2175.8 | 118.19 | 16.50 | 212.42 | 6.87 | 189.50 | 15.86 | |
| PY005-G05C 160427 | 23-May-16 | 3.117 | 1031.52 | 6.68 | 1867.1 | 113.64 | 14.18 | 186.22 | 7.22 | 175.76 | 14.18 | |
| PY006-G06C 160427 | 23-May-16 | 3.833 | 1005.27 | 6.71 | 2020.4 | 110.63 | 12.20 | 168.86 | 7.36 | 179.34 | 13.26 | |
| PY007-G07C 160427 | 23-May-16 | 4.133 | 1001.39 | 6.76 | 1724.3 | 109.48 | 26.66 | 167.00 | 6.59 | 177.91 | 12.71 | |
| PY008-G08C 160427 | 23-May-16 | 4.317 | 951.81 | 6.84 | 1873.4 | 108.55 | 15.95 | 156.63 | 7.10 | 172.48 | 12.52 | |
| PY009-G09C 160429 | 24-May-16 | 0.000 | 1668.21 | 4.70 | -30.8 | 100.87 | 10.99 | 727.41 | 8.18 | 182.23 | 35.97 | |
| PY015-G15C 160519 | 24-May-16 | 0.833 | 1125.81 | 6.63 | 953.8 | 97.69 | 74.80 | 311.97 | 9.82 | 128.97 | 25.82 | |
| PY016-G16C 160519 | 24-May-16 | 1.667 | 821.56 | 6.59 | 724.2 | 76.62 | 52.76 | 246.32 | 8.52 | 100.00 | 19.28 | |
| PY017-G17C 160519 | 24-May-16 | 2.383 | 660.19 | 6.79 | 977.1 | 57.07 | 30.42 | 181.97 | 7.99 | 81.37 | 14.48 | |
| ROCK = pyrite rock | | | | | | | | | | | | |
| PY001-U01C 160427 ROCK | 23-May-16 | 0.000 | 2273.80 | 3.47 | -250.7 | 32.88 | 11.73 | 867.40 | 8.56 | 4.21 | 33.67 | |
| PY002-U02C 160427 | 23-May-16 | 0.833 | 849.74 | 3.91 | -51.4 | 19.21 | 1.96 | 903.11 | 8.99 | 2.47 | 22.92 | |
| PY003-U03C 160427 | 23-May-16 | 1.667 | 745.74 | 4.01 | -20.9 | 20.50 | 1.49 | 418.13 | 9.41 | 2.20 | 21.51 | |
| PY004-U04C 160427 | 23-May-16 | 2.383 | 717.55 | 4.06 | -60.0 | 20.58 | 1.49 | 370.31 | 9.05 | 2.12 | 20.15 | |
| PY005-U05C 160427 | 23-May-16 | 3.117 | 665.05 | 4.15 | 19.1 | 21.17 | 1.45 | 331.58 | 10.55 | 2.60 | 19.76 | |
| PY006-U06C 160427 | 23-May-16 | 3.833 | 552.30 | 4.35 | -7.7 | 23.27 | 1.47 | 280.55 | 10.05 | 2.37 | 17.37 | |
| PY007-U07C 160427 | 23-May-16 | 4.133 | 516.33 | 4.58 | -7.5 | 23.51 | 1.57 | 248.08 | 11.65 | 3.76 | 17.02 | |
| PY008-U08C 160427 | 23-May-16 | 4.317 | 96.50 | 5.95 | 26.5 | 22.72 | 1.51 | 203.93 | 9.69 | 2.44 | 13.50 | |
| PY009-U09C 160429 | 24-May-16 | 0.000 | 1169.55 | 3.66 | -125.1 | 21.25 | 1.71 | 426.02 | 10.11 | 2.47 | 21.85 | |
| PY010-U10C 160504 | 24-May-16 | 0.833 | 607.70 | 3.49 | -123.1 | 18.61 | 2.28 | 433.07 | 7.57 | 1.62 | 13.77 | |
| PY011-U11C 160504 | 24-May-16 | 1.667 | 306.37 | 4.22 | -69.8 | 15.37 | 2.47 | 182.65 | 7.33 | 1.68 | 8.68 | |
| PY012-U12C 160504 | 24-May-16 | 2.383 | 123.33 | 6.40 | 85.0 | 17.09 | 1.22 | 90.85 | 7.21 | 1.70 | 6.70 | |
| PY013-U13C 160504 | 24-May-16 | 3.117 | 223.74 | 6.64 | 238.7 | 14.86 | 1.13 | 68.11 | 7.02 | 1.73 | 6.56 | |
| PY014-U14C 160504 | 24-May-16 | 3.833 | 210.13 | 6.77 | 289.3 | 14.19 | 2.56 | 50.09 | 6.58 | 1.50 | 5.91 | |
| PY015-U15C 160519 | 24-May-16 | 4.133 | 385.10 | 4.52 | 10.4 | 15.79 | 1.90 | 122.69 | 7.82 | 1.73 | 10.84 | |
| PY016-U16C 160519 | 24-May-16 | 4.317 | 257.76 | 6.58 | 215.1 | 16.51 | 1.52 | 119.52 | 7.60 | 1.71 | 8.10 | |
| PY017-U17C 160519 | 24-May-16 | 6.000 | 258.74 | 6.80 | 273.7 | 16.32 | 0.37 | 70.31 | 7.55 | 1.71 | 7.67 | |

Simulated Rainfall

| Sample/Tray Type | Date | Time (hrs) | uS/cm Conctvty | pH | measured | | | | | | |
|---------------------------------|-----------|------------|-------------------|------|--------------|------------|-------------|-------------|------------|-----------|------------|
| | | | | | ueq/L ANC | mg/L Cl | mg/L NO3 | mg/L SO4 | mg/L Na | mg/L K | mg/L Mg |
| CONC = shotcrete | | | | | | | | | | | |
| PY018-C18C 160526 CONC | 7-Jun-16 | 0.000 | 916.82 | 9.07 | | 22.25 | 0.97 | 179.32 | 51.08 | 202.05 | 4.38 |
| PY019-C19C 160526 | 7-Jun-16 | 0.833 | 411.35 | 9.10 | | 19.95 | 0.67 | 118.45 | 10.19 | 7.87 | 7.29 |
| PY020-C20C 160526 | 7-Jun-16 | 1.667 | 332.61 | 9.16 | | 17.81 | 0.25 | 17.07 | 12.70 | 24.70 | 5.62 |
| PY021-C21C 160526 | 7-Jun-16 | 2.383 | 280.12 | 8.88 | | 17.76 | 0.08 | 18.11 | 10.60 | 23.82 | 6.01 |
| PY022-C22C 160527 | 7-Jun-16 | 3.117 | 323.86 | 8.97 | 1967.8 | 17.31 | 0.05 | 16.90 | 12.26 | 33.48 | 5.86 |
| PY023-C23C 160527 | 7-Jun-16 | 3.833 | 266.51 | 8.45 | | 16.54 | 0.06 | 14.21 | 9.81 | 17.60 | 6.21 |
| PY024-C24C 160527 | 7-Jun-16 | 4.133 | 274.29 | 8.63 | | 17.55 | 0.05 | 13.08 | 10.72 | 19.22 | 6.45 |
| LINER = geoliner | | | | | | | | | | | |
| PY018-L18C 160526 LINER | 7-Jun-16 | 0.000 | 225.69 | 7.58 | 1276.5 | 15.95 | 2.33 | 12.47 | 8.25 | 2.07 | 6.43 |
| PY019-L19C 160526 | 7-Jun-16 | 0.833 | 246.10 | 7.82 | 1397.4 | 17.98 | 1.47 | 23.97 | 17.15 | 35.74 | 5.32 |
| PY020-L20C 160526 | 7-Jun-16 | 1.667 | 231.52 | 7.96 | | 18.34 | 0.41 | 25.60 | 8.58 | 2.97 | 6.60 |
| PY021-L21C 160526 | 7-Jun-16 | 2.383 | 230.55 | 7.81 | 1310.2 | 16.71 | 0.96 | 12.60 | 8.25 | 1.83 | 6.61 |
| PY022-L22C 160527 | 7-Jun-16 | 3.117 | 228.60 | 7.66 | 1275.1 | 16.89 | 0.88 | 13.20 | 8.62 | 1.89 | 6.77 |
| PY023-L23C 160527 | 7-Jun-16 | 3.833 | 232.49 | 7.77 | 1311.4 | 17.36 | 0.03 | 12.90 | 8.46 | 1.79 | 6.75 |
| PY024-L24C 160527 | 7-Jun-16 | 4.133 | 233.46 | 7.94 | 1325.0 | 17.17 | 0.03 | 12.76 | 8.53 | 1.86 | 6.76 |
| POOL = water supply tank | | | | | | | | | | | |
| PY009-P09C 160429 POOL | 23-May-16 | 0.000 | 239.30 | 7.89 | 1429.1 | 14.43 | 0.42 | 12.34 | 7.00 | 1.75 | 7.08 |
| PY009-P09D 160429 | 23-May-16 | 4.317 | | | | 14.38 | 0.02 | 12.54 | 0.71 | 0.20 | 0.76 |
| PY010-P10C 160504 | 24-May-16 | 0.000 | | | | 13.32 | 1.08 | 11.34 | 5.96 | 1.45 | 4.58 |
| PY010-P10D 160504 | 24-May-16 | 4.317 | | | | 18.83 | 0.16 | 13.81 | 0.83 | 0.24 | 0.59 |
| PY015-P15C 160519 | 7-Jun-16 | 0.000 | | | | 13.15 | 1.04 | 10.78 | 6.84 | 1.56 | 5.68 |
| PY015-P15D 160519 | 7-Jun-16 | 0.833 | | | | 13.85 | 0.05 | 11.22 | 6.93 | 1.58 | 5.77 |
| PY018-P18A 160526 | 7-Jun-16 | 1.667 | | | | 14.14 | 1.16 | 11.61 | 7.78 | 1.64 | 6.25 |
| PY018-P18B 160526 | 7-Jun-16 | 2.383 | | | | 14.60 | 0.77 | 12.63 | 8.32 | 1.89 | 6.32 |
| PY022-P22C 160527 | 7-Jun-16 | 3.117 | 215.97 | 7.74 | 1211.0 | 14.57 | 1.05 | 12.50 | 7.83 | 1.62 | 6.27 |
| PY022-P22D 160527 | 7-Jun-16 | 3.833 | | | | 14.48 | 1.17 | 11.97 | 7.88 | 1.69 | 6.30 |

PYRITE GEOLOGY WATER CHEMISTRY: SUMMARY SHEET MIDDLEBROOK PIKE EXPERIMENT

Simulated Rainfall

| Date | Time (hrs) | mg/L | mg/L | mg/L | mg/L | mg/L | mg/L | mg/L | mg/L | mg/L | Acid Anions | |
|--------------------------------|------------|--------|-------|------|--------|-------|-------|------|------|------|--------------------|------------------------------|
| | | Ca | Al | Cu | Fe | Mn | Si | Zn | Cd | Ni | Cl- (meq/L) | SO4 ²⁻ (meq/L) |
| GRASS = soil/vegetation | | | | | | | | | | | | |
| 23-May-16 | 0.000 | 59.39 | 0.29 | 0.04 | 13.04 | 4.02 | 3.31 | 0.22 | 0.00 | 0.06 | 7.066 | 0.722 |
| 23-May-16 | 0.833 | 83.08 | 0.61 | 0.06 | 12.15 | 4.53 | 8.86 | 0.25 | 0.00 | 0.13 | 5.176 | 0.304 |
| 23-May-16 | 1.667 | 90.20 | 0.46 | 0.04 | 9.76 | 0.48 | 10.36 | 0.15 | 0.00 | 0.08 | 4.382 | 0.306 |
| 23-May-16 | 2.383 | 81.08 | 0.26 | 0.03 | 3.83 | 0.44 | 9.98 | 0.09 | 0.00 | 0.06 | 3.339 | 0.344 |
| 23-May-16 | 3.117 | 73.79 | 0.24 | 0.03 | 3.42 | 1.04 | 9.42 | 0.07 | 0.00 | 0.06 | 3.210 | 0.295 |
| 23-May-16 | 3.833 | 68.05 | 0.27 | 0.03 | 3.06 | 0.89 | 8.07 | 0.08 | 0.00 | 0.04 | 3.125 | 0.254 |
| 23-May-16 | 4.133 | 66.66 | 0.36 | 0.03 | 2.59 | 0.79 | 9.93 | 0.08 | 0.00 | 0.05 | 3.093 | 0.556 |
| 23-May-16 | 4.317 | 62.44 | 0.31 | 0.03 | 3.74 | 0.75 | 10.35 | 0.10 | 0.00 | 0.06 | 3.066 | 0.332 |
| 24-May-16 | 0.000 | 155.34 | 0.64 | 0.03 | 2.91 | 10.07 | 15.54 | 1.25 | 0.00 | 0.66 | 2.849 | 0.229 |
| 24-May-16 | 0.833 | 95.12 | 0.47 | 0.02 | 3.25 | 3.69 | 6.22 | 0.41 | 0.00 | 0.18 | 2.760 | 1.558 |
| 24-May-16 | 1.667 | 79.55 | 0.39 | 0.02 | 2.80 | 1.99 | 5.46 | 0.28 | 0.00 | 0.14 | 2.164 | 1.099 |
| 24-May-16 | 2.383 | 65.46 | 0.33 | 0.02 | 2.28 | 1.04 | 4.81 | 0.19 | 0.00 | 0.10 | 1.612 | 0.634 |
| ROCK = pyrite rock | | | | | | | | | | | | |
| 23-May-16 | 0.000 | 72.64 | 15.17 | 0.63 | 134.67 | 25.16 | 3.15 | 6.15 | 0.00 | 3.08 | 0.929 | 0.244 |
| 23-May-16 | 0.833 | 61.81 | 1.66 | 0.19 | 143.50 | 15.37 | 2.49 | 2.00 | 0.00 | 1.04 | 0.543 | 0.041 |
| 23-May-16 | 1.667 | 43.53 | 1.39 | 0.19 | 56.17 | 14.14 | 2.59 | 1.78 | 0.00 | 0.94 | 0.579 | 0.031 |
| 23-May-16 | 2.383 | 40.95 | 0.57 | 0.15 | 48.25 | 12.64 | 2.55 | 1.54 | 0.00 | 0.83 | 0.581 | 0.031 |
| 23-May-16 | 3.117 | 40.76 | 0.50 | 0.11 | 39.17 | 11.48 | 2.55 | 1.44 | 0.00 | 0.79 | 0.598 | 0.030 |
| 23-May-16 | 3.833 | 39.97 | 0.15 | 0.04 | 21.90 | 9.02 | 2.28 | 1.07 | 0.00 | 0.62 | 0.657 | 0.031 |
| 23-May-16 | 4.133 | 41.38 | 0.18 | 0.03 | 14.05 | 8.01 | 2.23 | 0.92 | 0.00 | 0.55 | 0.664 | 0.033 |
| 23-May-16 | 4.317 | 38.84 | 0.03 | 0.01 | 4.13 | 5.08 | 1.97 | 0.46 | 0.00 | 0.34 | 0.642 | 0.031 |
| 24-May-16 | 0.000 | 44.20 | 15.75 | 0.61 | 51.89 | 23.63 | 4.65 | 3.58 | 0.00 | 1.68 | 0.600 | 0.036 |
| 24-May-16 | 0.833 | 28.33 | 4.04 | 0.25 | 60.64 | 8.39 | 3.09 | 1.34 | 0.00 | 0.68 | 0.526 | 0.048 |
| 24-May-16 | 1.667 | 23.51 | 0.24 | 0.04 | 6.06 | 3.60 | 2.48 | 0.50 | 0.00 | 0.27 | 0.434 | 0.051 |
| 24-May-16 | 2.383 | 22.87 | 0.04 | 0.01 | 2.68 | 1.37 | 2.33 | 0.19 | 0.00 | 0.11 | 0.483 | 0.025 |
| 24-May-16 | 3.117 | 22.46 | 0.01 | 0.01 | 2.49 | 1.39 | 2.30 | 0.15 | 0.00 | 0.10 | 0.420 | 0.024 |
| 24-May-16 | 3.833 | 21.63 | 0.00 | 0.00 | 0.15 | 0.85 | 2.11 | 0.03 | 0.00 | 0.06 | 0.401 | 0.053 |
| 24-May-16 | 4.133 | 28.37 | 0.11 | 0.04 | 10.92 | 5.20 | 2.62 | 0.84 | 0.00 | 0.39 | 0.446 | 0.039 |
| 24-May-16 | 4.317 | 26.04 | 0.01 | 0.01 | 1.76 | 2.16 | 2.32 | 0.17 | 0.00 | 0.13 | 0.466 | 0.032 |
| 24-May-16 | 6.000 | 25.65 | 0.01 | 0.01 | 2.17 | 1.23 | 2.42 | 0.18 | 0.00 | 0.12 | 0.461 | 0.008 |

Simulated Rainfall

| Date | Time (hrs) | mg/L | mg/L | mg/L | mg/L | mg/L | mg/L | mg/L | mg/L | mg/L | Acid Anions | |
|---------------------------------|------------|-------|------|------|------|------|-------|------|------|-------|--------------------|------------------------------|
| | | Ca | Al | Cu | Fe | Mn | Si | Zn | Cd | Ni | Cl- (meq/L) | SO4 ²⁻ (meq/L) |
| CONC = shotcrete | | | | | | | | | | | | |
| 7-Jun-16 | 0.000 | 26.73 | 0.17 | 0.02 | 0.22 | 0.19 | 12.52 | 0.01 | 0.00 | -0.01 | 0.628 | 0.020 |
| 7-Jun-16 | 0.833 | 19.13 | 0.03 | 0.02 | 0.02 | 0.01 | 3.29 | 0.07 | 0.00 | 0.00 | 0.564 | 0.014 |
| 7-Jun-16 | 1.667 | 23.77 | 0.05 | 0.01 | 0.02 | 0.01 | 4.99 | 0.02 | 0.00 | -0.01 | 0.503 | 0.005 |
| 7-Jun-16 | 2.383 | 24.62 | 0.04 | 0.01 | 0.02 | 0.01 | 3.93 | 0.02 | 0.00 | -0.01 | 0.502 | 0.002 |
| 7-Jun-16 | 3.117 | 23.78 | 0.05 | 0.01 | 0.02 | 0.02 | 4.50 | 0.02 | 0.00 | -0.01 | 0.489 | 0.001 |
| 7-Jun-16 | 3.833 | 24.99 | 0.04 | 0.01 | 0.01 | 0.01 | 3.48 | 0.02 | 0.00 | 0.00 | 0.467 | 0.001 |
| 7-Jun-16 | 4.133 | 26.32 | 0.05 | 0.01 | 0.01 | 0.01 | 3.77 | 0.03 | 0.00 | 0.00 | 0.496 | 0.001 |
| LINER = geoliner | | | | | | | | | | | | |
| 7-Jun-16 | 0.000 | 26.25 | 0.02 | 0.02 | 0.02 | 0.01 | 2.71 | 0.08 | 0.00 | 0.00 | 0.451 | 0.049 |
| 7-Jun-16 | 0.833 | 24.68 | 0.07 | 0.02 | 0.04 | 0.03 | 6.58 | 0.02 | 0.00 | -0.01 | 0.508 | 0.031 |
| 7-Jun-16 | 1.667 | 26.57 | 0.02 | 0.01 | 0.01 | 0.00 | 2.80 | 0.07 | 0.00 | 0.00 | 0.518 | 0.008 |
| 7-Jun-16 | 2.383 | 26.96 | 0.02 | 0.01 | 0.01 | 0.00 | 2.71 | 0.07 | 0.00 | 0.00 | 0.472 | 0.020 |
| 7-Jun-16 | 3.117 | 27.21 | 0.02 | 0.01 | 0.01 | 0.00 | 2.79 | 0.07 | 0.00 | 0.00 | 0.477 | 0.018 |
| 7-Jun-16 | 3.833 | 27.23 | 0.02 | 0.01 | 0.00 | 0.00 | 2.78 | 0.07 | 0.00 | 0.00 | 0.490 | 0.001 |
| 7-Jun-16 | 4.133 | 27.29 | 0.02 | 0.01 | 0.00 | 0.00 | 2.83 | 0.08 | 0.00 | 0.00 | 0.485 | 0.001 |
| POOL = water supply tank | | | | | | | | | | | | |
| 23-May-16 | 0.000 | 0.00 | 0.02 | 0.01 | 0.01 | 0.00 | 1.50 | 0.07 | 0.00 | 0.00 | 0.408 | 0.009 |
| 23-May-16 | 4.317 | 3.30 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.406 | 0.000 |
| 24-May-16 | 0.000 | 19.53 | 0.01 | 0.01 | 0.00 | 0.00 | 2.13 | 0.06 | 0.00 | 0.00 | 0.376 | 0.023 |
| 24-May-16 | 4.317 | 2.35 | 0.00 | 0.00 | 0.00 | 0.00 | 0.11 | 0.01 | 0.00 | 0.00 | 0.532 | 0.003 |
| 7-Jun-16 | 0.000 | 23.46 | 0.01 | 0.01 | 0.01 | 0.00 | 2.38 | 0.07 | 0.00 | 0.00 | 0.371 | 0.022 |
| 7-Jun-16 | 0.833 | 23.82 | 0.01 | 0.01 | 0.01 | 0.00 | 2.42 | 0.07 | 0.00 | 0.00 | 0.391 | 0.001 |
| 7-Jun-16 | 1.667 | 25.16 | 0.02 | 0.01 | 0.01 | 0.00 | 2.57 | 0.08 | 0.00 | 0.00 | 0.400 | 0.024 |
| 7-Jun-16 | 2.383 | 25.38 | 0.02 | 0.01 | 0.01 | 0.00 | 2.57 | 0.07 | 0.00 | 0.00 | 0.412 | 0.016 |
| 7-Jun-16 | 3.117 | 25.29 | 0.02 | 0.01 | 0.01 | 0.00 | 2.56 | 0.07 | 0.00 | 0.00 | 0.412 | 0.022 |
| 7-Jun-16 | 3.833 | 25.50 | 0.02 | 0.01 | 0.00 | 0.00 | 2.60 | 0.07 | 0.00 | 0.00 | 0.409 | 0.024 |

PYRITE GEOLOGY WATER CHEMISTRY: SUMMARY SHEET MIDDLEBROOK PIKE EXPERIMENT

Simulated Rainfall

Base Cations

Dissolved Metals

| Date | Time (hrs) | NO3- | PO4 ³⁻ | Sum AA | Na+ | K + | Mg ²⁺ | Ca ²⁺ | Sum BC | Al ³⁺ | Cu ²⁺ | Fe ²⁺ | Mn ²⁺ |
|--------------------------------|------------|---------|-------------------|---------|---------|---------|------------------|------------------|---------|------------------|------------------|------------------|------------------|
| | | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) | (meq/L) |
| GRASS = soil/vegetation | | | | | | | | | | | | | |
| 23-May-16 | 0.000 | 2.326 | | 10.114 | 0.437 | 5.536 | 1.587 | 2.969 | 10.530 | 0.032 | 0.001 | 0.467 | 0.146 |
| 23-May-16 | 0.833 | 4.015 | | 9.495 | 0.373 | 5.321 | 1.889 | 4.154 | 11.736 | 0.068 | 0.002 | 0.435 | 0.165 |
| 23-May-16 | 1.667 | 4.759 | | 9.447 | 0.330 | 5.294 | 1.641 | 4.510 | 11.775 | 0.052 | 0.001 | 0.350 | 0.017 |
| 23-May-16 | 2.383 | 3.426 | | 7.109 | 0.299 | 4.859 | 1.306 | 4.054 | 10.517 | 0.029 | 0.001 | 0.137 | 0.016 |
| 23-May-16 | 3.117 | 3.004 | | 6.509 | 0.314 | 4.507 | 1.167 | 3.689 | 9.677 | 0.026 | 0.001 | 0.123 | 0.038 |
| 23-May-16 | 3.833 | 2.724 | | 6.103 | 0.320 | 4.598 | 1.092 | 3.403 | 9.413 | 0.030 | 0.001 | 0.110 | 0.032 |
| 23-May-16 | 4.133 | 2.694 | | 6.342 | 0.286 | 4.562 | 1.046 | 3.333 | 9.228 | 0.040 | 0.001 | 0.093 | 0.029 |
| 23-May-16 | 4.317 | 2.526 | | 5.925 | 0.309 | 4.423 | 1.030 | 3.122 | 8.883 | 0.034 | 0.001 | 0.134 | 0.027 |
| 24-May-16 | 0.000 | 11.732 | | 14.811 | 0.356 | 4.673 | 2.960 | 7.767 | 15.756 | 0.071 | 0.001 | 0.104 | 0.367 |
| 24-May-16 | 0.833 | 5.032 | | 9.350 | 0.427 | 3.307 | 2.125 | 4.756 | 10.615 | 0.053 | 0.001 | 0.116 | 0.134 |
| 24-May-16 | 1.667 | 3.973 | | 7.236 | 0.370 | 2.564 | 1.587 | 3.977 | 8.499 | 0.044 | 0.001 | 0.100 | 0.072 |
| 24-May-16 | 2.383 | 2.935 | | 5.181 | 0.347 | 2.086 | 1.191 | 3.273 | 6.898 | 0.037 | 0.001 | 0.082 | 0.038 |
| ROCK = pyrite rock | | | | | | | | | | | | | |
| 23-May-16 | 0.000 | 13.990 | | 15.163 | 0.372 | 0.108 | 2.772 | 3.632 | 6.884 | 1.687 | 0.020 | 4.823 | 0.916 |
| 23-May-16 | 0.833 | 14.566 | | 15.150 | 0.391 | 0.063 | 1.886 | 3.090 | 5.431 | 0.184 | 0.006 | 5.139 | 0.560 |
| 23-May-16 | 1.667 | 6.744 | | 7.354 | 0.409 | 0.056 | 1.770 | 2.177 | 4.412 | 0.154 | 0.006 | 2.011 | 0.515 |
| 23-May-16 | 2.383 | 5.973 | | 6.585 | 0.394 | 0.054 | 1.658 | 2.047 | 4.154 | 0.064 | 0.005 | 1.728 | 0.460 |
| 23-May-16 | 3.117 | 5.348 | | 5.976 | 0.459 | 0.067 | 1.627 | 2.038 | 4.190 | 0.056 | 0.003 | 1.403 | 0.418 |
| 23-May-16 | 3.833 | 4.525 | | 5.213 | 0.437 | 0.061 | 1.430 | 1.999 | 3.926 | 0.017 | 0.001 | 0.784 | 0.328 |
| 23-May-16 | 4.133 | 4.001 | | 4.698 | 0.507 | 0.096 | 1.401 | 2.069 | 4.073 | 0.020 | 0.001 | 0.503 | 0.292 |
| 23-May-16 | 4.317 | 3.289 | | 3.962 | 0.421 | 0.063 | 1.111 | 1.942 | 3.537 | 0.003 | 0.000 | 0.148 | 0.185 |
| 24-May-16 | 0.000 | 6.871 | | 7.507 | 0.439 | 0.063 | 1.798 | 2.210 | 4.511 | 1.752 | 0.019 | 1.858 | 0.860 |
| 24-May-16 | 0.833 | 6.985 | | 7.558 | 0.329 | 0.042 | 1.134 | 1.417 | 2.921 | 0.449 | 0.008 | 2.172 | 0.305 |
| 24-May-16 | 1.667 | 2.946 | | 3.431 | 0.319 | 0.043 | 0.715 | 1.176 | 2.252 | 0.027 | 0.001 | 0.217 | 0.131 |
| 24-May-16 | 2.383 | 1.465 | | 1.973 | 0.314 | 0.044 | 0.552 | 1.144 | 2.052 | 0.005 | 0.000 | 0.096 | 0.050 |
| 24-May-16 | 3.117 | 1.099 | | 1.542 | 0.305 | 0.044 | 0.540 | 1.123 | 2.013 | 0.001 | 0.000 | 0.089 | 0.051 |
| 24-May-16 | 3.833 | 0.808 | | 1.262 | 0.286 | 0.038 | 0.486 | 1.081 | 1.892 | 0.000 | 0.000 | 0.005 | 0.031 |
| 24-May-16 | 4.133 | 1.979 | | 2.464 | 0.340 | 0.044 | 0.892 | 1.419 | 2.695 | 0.013 | 0.001 | 0.391 | 0.189 |
| 24-May-16 | 4.317 | 1.928 | | 2.426 | 0.330 | 0.044 | 0.667 | 1.302 | 2.343 | 0.001 | 0.000 | 0.063 | 0.079 |
| 24-May-16 | 6.000 | 1.134 | | 1.603 | 0.328 | 0.044 | 0.631 | 1.283 | 2.286 | 0.001 | 0.000 | 0.078 | 0.045 |

Simulated Rainfall

| | | Base Cations | | | | | | | Dissolved Metals | | | | |
|---------------------------------|------------|-----------------|------------------------------|-------------------|----------------|----------------|-------------------|-------------------|-------------------|-----------------------------|-----------------------------|-----------------------------|-----------------------------|
| Date | Time (hrs) | NO3- (meq/L) | PO4 ³⁻ (meq/L) | Sum AA (meq/L) | Na+ (meq/L) | K + (meq/L) | Mg ^2+ (meq/L) | Ca ^2+ (meq/L) | Sum BC (meq/L) | Al ³⁺ (meq/L) | Cu ²⁺ (meq/L) | Fe ²⁺ (meq/L) | Mn ²⁺ (meq/L) |
| CONC = shotcrete | | | | | | | | | | | | | |
| 7-Jun-16 | 0.000 | 2.892 | | 3.541 | 2.221 | 5.181 | 0.361 | 1.336 | 9.099 | 0.019 | 0.001 | 0.008 | 0.007 |
| 7-Jun-16 | 0.833 | 1.911 | | 2.488 | 0.443 | 0.202 | 0.600 | 0.957 | 2.202 | 0.003 | 0.001 | 0.001 | 0.000 |
| 7-Jun-16 | 1.667 | 0.275 | | 0.784 | 0.552 | 0.633 | 0.462 | 1.188 | 2.836 | 0.006 | 0.000 | 0.001 | 0.001 |
| 7-Jun-16 | 2.383 | 0.292 | | 0.796 | 0.461 | 0.611 | 0.495 | 1.231 | 2.798 | 0.004 | 0.000 | 0.001 | 0.000 |
| 7-Jun-16 | 3.117 | 0.273 | | 0.763 | 0.533 | 0.858 | 0.482 | 1.189 | 3.063 | 0.006 | 0.000 | 0.001 | 0.001 |
| 7-Jun-16 | 3.833 | 0.229 | | 0.698 | 0.427 | 0.451 | 0.512 | 1.249 | 2.639 | 0.004 | 0.000 | 0.000 | 0.000 |
| 7-Jun-16 | 4.133 | 0.211 | | 0.708 | 0.466 | 0.493 | 0.531 | 1.316 | 2.806 | 0.005 | 0.000 | 0.000 | 0.000 |
| LINER = geoliner | | | | | | | | | | | | | |
| 7-Jun-16 | 0.000 | 0.201 | | 0.700 | 0.358 | 0.053 | 0.529 | 1.313 | 2.253 | 0.002 | 0.001 | 0.001 | 0.000 |
| 7-Jun-16 | 0.833 | 0.387 | | 0.925 | 0.746 | 0.916 | 0.438 | 1.234 | 3.334 | 0.008 | 0.001 | 0.002 | 0.001 |
| 7-Jun-16 | 1.667 | 0.413 | | 0.939 | 0.373 | 0.076 | 0.543 | 1.328 | 2.321 | 0.002 | 0.000 | 0.000 | 0.000 |
| 7-Jun-16 | 2.383 | 0.203 | | 0.695 | 0.359 | 0.047 | 0.544 | 1.348 | 2.297 | 0.002 | 0.000 | 0.000 | 0.000 |
| 7-Jun-16 | 3.117 | 0.213 | | 0.708 | 0.375 | 0.049 | 0.557 | 1.361 | 2.341 | 0.002 | 0.000 | 0.000 | 0.000 |
| 7-Jun-16 | 3.833 | 0.208 | | 0.699 | 0.368 | 0.046 | 0.556 | 1.362 | 2.331 | 0.002 | 0.000 | 0.000 | 0.000 |
| 7-Jun-16 | 4.133 | 0.206 | | 0.692 | 0.371 | 0.048 | 0.556 | 1.364 | 2.340 | 0.002 | 0.000 | 0.000 | 0.000 |
| POOL = water supply tank | | | | | | | | | | | | | |
| 23-May-16 | 0.000 | 0.199 | | 0.615 | 0.304 | 0.045 | 0.582 | 0.000 | 0.931 | 0.002 | 0.000 | 0.000 | 0.000 |
| 23-May-16 | 4.317 | 0.202 | | 0.609 | 0.031 | 0.005 | 0.062 | 0.165 | 0.263 | 0.000 | 0.000 | 0.000 | 0.000 |
| 24-May-16 | 0.000 | 0.183 | | 0.582 | 0.259 | 0.037 | 0.377 | 0.976 | 1.649 | 0.002 | 0.000 | 0.000 | 0.000 |
| 24-May-16 | 4.317 | 0.223 | | 0.758 | 0.036 | 0.006 | 0.049 | 0.117 | 0.208 | 0.000 | 0.000 | 0.000 | 0.000 |
| 7-Jun-16 | 0.000 | 0.174 | | 0.567 | 0.297 | 0.040 | 0.467 | 1.173 | 1.978 | 0.001 | 0.000 | 0.000 | 0.000 |
| 7-Jun-16 | 0.833 | 0.181 | | 0.573 | 0.301 | 0.040 | 0.475 | 1.191 | 2.007 | 0.002 | 0.000 | 0.000 | 0.000 |
| 7-Jun-16 | 1.667 | 0.187 | | 0.611 | 0.338 | 0.042 | 0.514 | 1.258 | 2.153 | 0.002 | 0.000 | 0.000 | 0.000 |
| 7-Jun-16 | 2.383 | 0.204 | | 0.632 | 0.362 | 0.048 | 0.520 | 1.269 | 2.199 | 0.002 | 0.000 | 0.000 | 0.000 |
| 7-Jun-16 | 3.117 | 0.202 | | 0.635 | 0.340 | 0.042 | 0.516 | 1.265 | 2.162 | 0.002 | 0.000 | 0.000 | 0.000 |
| 7-Jun-16 | 3.833 | 0.193 | | 0.626 | 0.343 | 0.043 | 0.519 | 1.275 | 2.179 | 0.002 | 0.000 | 0.000 | 0.000 |

PYRITE GEOLOGY WATER CHEMISTRY: SUMMARY SHEET MIDDLEBROOK PIKE EXPERIMENT

Simulated Rainfall

| Date | Time (hrs) | Calculated | | | | |
|--------------------------------|------------|-----------------------------|-----------------------------|-----------------------------|-----------------------------|----------------|
| | | Si ²⁺ (meq/L) | Zn ²⁺ (meq/L) | Cd ²⁺ (meq/L) | Ni ²⁺ (meq/L) | ANC (meq/L) |
| GRASS = soil/vegetation | | | | | | |
| 23-May-16 | 0.000 | 0.236 | 0.007 | 0.000 | 0.002 | 1.306 |
| 23-May-16 | 0.833 | 0.631 | 0.008 | 0.000 | 0.004 | 3.554 |
| 23-May-16 | 1.667 | 0.738 | 0.005 | 0.000 | 0.003 | 3.492 |
| 23-May-16 | 2.383 | 0.710 | 0.003 | 0.000 | 0.002 | 4.307 |
| 23-May-16 | 3.117 | 0.671 | 0.002 | 0.000 | 0.002 | 4.031 |
| 23-May-16 | 3.833 | 0.575 | 0.003 | 0.000 | 0.002 | 4.061 |
| 23-May-16 | 4.133 | 0.707 | 0.002 | 0.000 | 0.002 | 3.759 |
| 23-May-16 | 4.317 | 0.737 | 0.003 | 0.000 | 0.002 | 3.897 |
| 24-May-16 | 0.000 | 1.106 | 0.038 | 0.000 | 0.023 | 2.655 |
| 24-May-16 | 0.833 | 0.443 | 0.013 | 0.000 | 0.006 | 2.031 |
| 24-May-16 | 1.667 | 0.389 | 0.009 | 0.000 | 0.005 | 1.881 |
| 24-May-16 | 2.383 | 0.342 | 0.006 | 0.000 | 0.003 | 2.226 |
| ROCK = pyrite rock | | | | | | |
| 23-May-16 | 0.000 | 0.224 | 0.188 | 0.000 | 0.105 | -0.317 |
| 23-May-16 | 0.833 | 0.177 | 0.061 | 0.000 | 0.035 | -3.556 |
| 23-May-16 | 1.667 | 0.184 | 0.054 | 0.000 | 0.032 | 0.016 |
| 23-May-16 | 2.383 | 0.182 | 0.047 | 0.000 | 0.028 | 0.082 |
| 23-May-16 | 3.117 | 0.181 | 0.044 | 0.000 | 0.027 | 0.345 |
| 23-May-16 | 3.833 | 0.162 | 0.033 | 0.000 | 0.021 | 0.060 |
| 23-May-16 | 4.133 | 0.159 | 0.028 | 0.000 | 0.019 | 0.396 |
| 23-May-16 | 4.317 | 0.140 | 0.014 | 0.000 | 0.012 | 0.077 |
| 24-May-16 | 0.000 | 0.331 | 0.110 | 0.000 | 0.057 | 1.991 |
| 24-May-16 | 0.833 | 0.220 | 0.041 | 0.000 | 0.023 | -1.419 |
| 24-May-16 | 1.667 | 0.177 | 0.015 | 0.000 | 0.009 | -0.602 |
| 24-May-16 | 2.383 | 0.166 | 0.006 | 0.000 | 0.004 | 0.406 |
| 24-May-16 | 3.117 | 0.164 | 0.004 | 0.000 | 0.004 | 0.784 |
| 24-May-16 | 3.833 | 0.150 | 0.001 | 0.000 | 0.002 | 0.820 |
| 24-May-16 | 4.133 | 0.187 | 0.026 | 0.000 | 0.013 | 1.051 |
| 24-May-16 | 4.317 | 0.165 | 0.005 | 0.000 | 0.005 | 0.235 |
| 24-May-16 | 6.000 | 0.172 | 0.005 | 0.000 | 0.004 | 0.988 |

Simulated Rainfall

| Date | Time (hrs) | Calculated | | | | |
|---------------------------------|------------|-----------------|------------------|------------------|------------------|----------------|
| | | Si2+ (meq/L) | Zn^2+ (meq/L) | Cd^2+ (meq/L) | Ni^2+ (meq/L) | ANC (meq/L) |
| CONC = shotcrete | | | | | | |
| 7-Jun-16 | 0.000 | 0.891 | 0.000 | 0.000 | 0.000 | 6.484 |
| 7-Jun-16 | 0.833 | 0.234 | 0.002 | 0.000 | 0.000 | -0.046 |
| 7-Jun-16 | 1.667 | 0.356 | 0.001 | 0.000 | 0.000 | 2.416 |
| 7-Jun-16 | 2.383 | 0.280 | 0.001 | 0.000 | 0.000 | 2.288 |
| 7-Jun-16 | 3.117 | 0.320 | 0.001 | 0.000 | 0.000 | 2.628 |
| 7-Jun-16 | 3.833 | 0.248 | 0.001 | 0.000 | 0.000 | 2.195 |
| 7-Jun-16 | 4.133 | 0.268 | 0.001 | 0.000 | 0.000 | 2.373 |
| LINER = geoliner | | | | | | |
| 7-Jun-16 | 0.000 | 0.193 | 0.002 | 0.000 | 0.000 | 1.752 |
| 7-Jun-16 | 0.833 | 0.469 | 0.001 | 0.000 | 0.000 | 2.889 |
| 7-Jun-16 | 1.667 | 0.199 | 0.002 | 0.000 | 0.000 | 1.586 |
| 7-Jun-16 | 2.383 | 0.193 | 0.002 | 0.000 | 0.000 | 1.800 |
| 7-Jun-16 | 3.117 | 0.199 | 0.002 | 0.000 | 0.000 | 1.837 |
| 7-Jun-16 | 3.833 | 0.198 | 0.002 | 0.000 | 0.000 | 1.835 |
| 7-Jun-16 | 4.133 | 0.201 | 0.002 | 0.000 | 0.000 | 1.854 |
| POOL = water supply tank | | | | | | |
| 23-May-16 | 0.000 | 0.107 | 0.002 | 0.000 | 0.000 | 0.428 |
| 23-May-16 | 4.317 | 0.000 | 0.000 | 0.000 | 0.000 | -0.345 |
| 24-May-16 | 0.000 | 0.151 | 0.002 | 0.000 | 0.000 | 1.222 |
| 24-May-16 | 4.317 | 0.008 | 0.000 | 0.000 | 0.000 | -0.541 |
| 7-Jun-16 | 0.000 | 0.169 | 0.002 | 0.000 | 0.000 | 1.584 |
| 7-Jun-16 | 0.833 | 0.172 | 0.002 | 0.000 | 0.000 | 1.610 |
| 7-Jun-16 | 1.667 | 0.183 | 0.002 | 0.000 | 0.000 | 1.730 |
| 7-Jun-16 | 2.383 | 0.183 | 0.002 | 0.000 | 0.000 | 1.755 |
| 7-Jun-16 | 3.117 | 0.183 | 0.002 | 0.000 | 0.000 | 1.714 |
| 7-Jun-16 | 3.833 | 0.185 | 0.002 | 0.000 | 0.000 | 1.743 |

Appendix I: Water Chemistry for Middlebrook Pike Experiment: Natural Rainfall

PYRITE GEOLOGY WATER CHEMISTRY: SUMMARY SHEET MIDDLEBROOK PIKE EXPERIMENT

Natural Rainfall

| Analysis Name 1 | Storm Number | Date Collected | Sample/ Tray | Conductivity (uS/cm) | pH units | ANC - meas (meq/L) | ANC - calc (meq/L) | Cl- (mg/L) | Acid Anions | | | | |
|---------------------|--------------|----------------|--------------|-------------------------|----------|--------------------|--------------------|------------|---------------|-------------|---------------|--|--|
| | | | | | | | | | SO4^2- (mg/L) | NO3- (mg/L) | PO4^3- (mg/L) | | |
| NRF1 160621 CONC-b | 1 | 6/21/2016 | CONC | 614.0 | 6.76 | | 5.614232 | 6.76517 | 6.3281 | | | | |
| NRF2 160624 CONC-b | 2 | 6/24/2016 | CONC | 369.7 | 7.74 | | | 1.99124 | 2.05567 | 46.98609 | | | |
| NRF3 160627 CONC-b | 3 | 6/27/2016 | CONC | 69.5 | 7.46 | | | 0.22506 | 0.86851 | 8.90029 | 0.7389 | | |
| NRF4 160628 CONC-b | 4 | 6/28/2016 | CONC | 181.1 | 7.16 | | | 1.02395 | 1.71907 | 30.05022 | | | |
| NRF5 160705 CONC | 5 | 7/5/2016 | CONC | 154.8 | 6.98 | | | 0.87845 | 1.0868 | 31.92817 | | | |
| NRF6 160707 CONC | 6 | 7/7/2016 | CONC | 99.5 | 7.41 | | | 0.68158 | 0.77936 | 15.11764 | | | |
| NRF7 160709 CONC | 7 | 7/9/2016 | CONC | 82.8 | 6.87 | | | 0.63436 | 0.80814 | 11.49353 | | | |
| NRF8 160715 CONC | 8 | 7/15/2016 | CONC | 180.1 | 7.16 | | | 0.83566 | 2.52484 | 29.07725 | 0.00304 | | |
| NRF9 160727 CONC | 9 | 7/27/2016 | CONC | 165.6 | 6.85 | | | 0.48602 | 2.77483 | 33.88348 | 0.00087 | | |
| NRF10 160729 CONC | 10 | 7/29/2016 | CONC | 69.7 | 6.99 | | | 0.38734 | 0.6357 | 12.83604 | 0.0119 | | |
| NRF11 160730 CONC | 11 | 7/30/2016 | CONC | 64.8 | 7.03 | | | 0.15243 | 0.46494 | 12.34996 | 0.00537 | | |
| NRF12 160805 CONC | 12 | 8/5/2016 | CONC | 67.5 | 6.87 | | | 0.17109 | 0.58384 | 11.20241 | 0.01098 | | |
| NRF14 160808 CONC | 14 | 8/8/2016 | CONC | 46.4 | 6.80 | | | 0.15914 | 0.56813 | 4.81829 | 0.00297 | | |
| NRF15 160809 CONC | 15 | 8/9/2016 | CONC | 75.4 | 6.84 | | | 0.21864 | 1.13982 | 10.84072 | 0.00391 | | |
| NRF16 160816 CONC | 16 | 8/16/2016 | CONC | 43.4 | 6.73 | | | 0.11906 | 0.58404 | 7.28097 | 0.0036 | | |
| NRF17 160818 CONC | 17 | 8/18/2016 | CONC | 42.5 | 6.85 | | | 0.18153 | 0.43056 | 10.0873 | 0.01525 | | |
| NRF18 160822 CONC | 18 | 8/22/2016 | CONC | 82.4 | 6.91 | | | 0.58134 | 0.82641 | 22.93531 | 0.0022 | | |
| NRF19 160902 CONC | 19 | 9/2/2016 | CONC | 106.9 | 6.88 | | | 2.05929 | 3.34523 | 30.80017 | 0.00745 | | |
| NRF20 160919 CONC-a | 20 | 9/19/2016 | CONC | 66.9 | 6.99 | | | 0.86123 | 0.80688 | 15.9823 | | | |
| NRF22 161121 CONC | 22 | 11/21/2016 | CONC | 199.3 | 7.62 | | | 131.392 | 30.8089 | 138.273 | | | |
| NRF23 161130 CONC | 23 | 11/30/2016 | CONC | 93.2 | 7.01 | | | 0.22672 | 0.24154 | 19.301 | | | |
| NRF24 161204 CONC | 24 | 12/4/2016 | CONC | 10.8 | 6.15 | | | 0.0342 | 0.0516 | 4.65104 | | | |
| NRF25 161206 CONC | 25 | 12/6/2016 | CONC | 7.5 | 5.95 | | | 0.03489 | 0.05351 | 1.83138 | | | |
| NRF26 161212 CONC | 26 | 12/12/2016 | CONC | 36.9 | 6.71 | | | 0.23505 | 0.36193 | 6.97396 | | | |
| | | | | GRASS = soil/vegetation | | | | | | | | | |
| NRF1 160621 GRASS-b | 1 | 6/21/2016 | GRASS | 2360 | 6.20 | | 9.638282 | 87.55716 | 90.82285 | | | | |
| NRF2 160624 GRASS-b | 2 | 6/24/2016 | GRASS | 1232.43237 | 6.48 | | | 46.57309 | 39.81029 | 56.44726 | 16.31211 | | |
| NRF3 160627 GRASS-b | 3 | 6/27/2016 | GRASS | 709.189209 | 5.86 | | | 29.20129 | 40.67187 | 59.335 | | | |
| NRF4 160628 GRASS-b | 4 | 6/28/2016 | GRASS | 1676.75671 | 6.24 | | | 58.80179 | 82.41462 | 76.79998 | 3.42756 | | |

CONC = shotcrete

Acid Anions

| Analysis Name 1 | Storm Number | Date Collected | Sample/ Tray | Conductivity (uS/cm) | pH units | ANC - meas (meq/L) | ANC - calc (meq/L) | Cl- (mg/L) | SO4^2- (mg/L) | NO3- (mg/L) | PO4^3- (mg/L) |
|----------------------|--------------|----------------|-------------------------|----------------------|----------|--------------------|--------------------|------------|---------------|-------------|---------------|
| | 5 | 7/5/2016 | GRASS = soil/vegetation | | | | | | | | |
| NRF6 160707 GRASS | 6 | 7/7/2016 | GRASS | 1224.1344 | 5.02 | -1.425642 | -1.425642 | 78.8146 | 93.7397 | 89.052 | 0.0007 |
| NRF7 160709 GRASS | 7 | 7/9/2016 | GRASS | 731.111084 | 5.93 | -6.949228 | -6.949228 | 193.4379 | 220.3916 | 302.5941 | 0.0008 |
| NRF8 160715 GRASS | 8 | 7/15/2016 | GRASS | 801.268738 | 6.58 | 0.65793085 | -0.898155 | 88.956 | 116.3458 | 88.9548 | 0.1011 |
| NRF9 160727 GRASS | 9 | 7/27/2016 | GRASS | 423.875427 | 6.42 | 0.59244955 | 1.370314 | 28.2718 | 21.4401 | 66.6938 | 0.0027 |
| NRF10 160729 GRASS | 10 | 7/29/2016 | GRASS | 334.371399 | 5.66 | 0.00556682 | 0.34462 | 26.0852 | 57.4008 | 59.4146 | 0.016 |
| NRF11 160730 GRASS | 11 | 7/30/2016 | GRASS | 446.435638 | 5.20 | -0.0179877 | -0.458007 | 33.5069 | 79.692 | 50.3464 | 0.0496 |
| NRF12 160805 GRASS | 12 | 8/5/2016 | GRASS | 358.316833 | 5.63 | 0.00449146 | -17.53196 | 219.1913 | 420.4243 | 351.7899 | 3.9964 |
| NRF14 160808 GRASS | 14 | 8/8/2016 | GRASS | 269.207916 | 5.45 | -0.0136938 | 0.234067 | 15.9438 | 44.1896 | 44.4574 | 0.0469 |
| NRF15 160809 GRASS | 15 | 8/9/2016 | GRASS | 420.693054 | 4.96 | -0.0296566 | 0.536082 | 20.8648 | 56.0223 | 91.8962 | 0.0261 |
| NRF16 160816 GRASS | 16 | 8/16/2016 | GRASS | 180.693069 | 5.44 | -0.0112255 | 0.379939 | 6.1176 | 24.3264 | 35.6084 | 0.0296 |
| NRF17 160818 GRASS | 17 | 8/18/2016 | GRASS | 159.007523 | 4.90 | -0.0095564 | 0.43417 | 9.0644 | 36.5964 | 49.0334 | 0.2971 |
| NRF18 160822 GRASS | 18 | 8/22/2016 | GRASS | 243.218048 | 4.80 | -0.0140315 | 1.417363 | 12.3728 | 49.3509 | 89.1059 | 0.5239 |
| NRF19 160902 GRASS | 19 | 9/2/2016 | GRASS | 215.548874 | 5.25 | -0.0071012 | 0.957684 | 9.1706 | 23.0188 | 85.7362 | 0.0871 |
| NRF20 160919 GRASS-a | 20 | 9/19/2016 | GRASS | 144.571426 | 4.80 | -0.0213955 | 0.548806 | 6.1426 | 23.6651 | 66.265 | |
| NRF22 161121 GRASS | 22 | 11/21/2016 | GRASS | 432.691742 | 6.38 | 0.14599395 | 0.179933 | 68.7467 | 14.6023 | 275.07 | |
| NRF23 161130 GRASS | 23 | 11/30/2016 | GRASS | 117.744362 | 5.55 | 0.00182235 | 0.689625 | 4.0369 | 15.8494 | 37.3785 | |
| NRF24 161204 GRASS | 24 | 12/4/2016 | GRASS | 117.503761 | 6.14 | 0.0293068 | 0.613787 | 4.70511 | 15.5742 | 46.6571 | |
| NRF25 161206 GRASS | 25 | 12/6/2016 | GRASS | 138.556396 | 4.71 | -0.0169368 | 0.318336 | 5.04314 | 15.5928 | 87.877 | |
| NRF26 161212 GRASS | 26 | 12/12/2016 | GRASS | 115.759399 | 4.80 | -0.0159086 | 0.421292 | 2.52683 | 7.53436 | 71.832 | |
| | | | LINER = geoliner | | | | | | | | |
| NRF1 160621 LINER-b | 1 | 6/21/2016 | LINER | 55.2999992 | 6.81 | 0.269335 | 0.269335 | 3.69073 | 4.17241 | 9.64101 | |
| NRF2 160624 LINER-b | 2 | 6/24/2016 | LINER | 22.7027035 | 6.75 | | | 0.43018 | 1.64106 | 1.15883 | |
| NRF3 160627 LINER-b | 3 | 6/27/2016 | LINER | 7.39157057 | 5.24 | | | 0.1001 | 0.75365 | 0.66976 | |
| NRF4 160628 LINER-b | 4 | 6/28/2016 | LINER | 10.6658525 | 6.15 | | | 0.18857 | 1.78368 | 0.69017 | |
| NRF5 160705 LINER | 5 | 7/5/2016 | LINER | 14.8486462 | 6.37 | 0.101643 | 0.101643 | 0.42212 | 1.29586 | 1.39462 | |
| NRF6 160707 LINER | 6 | 7/7/2016 | LINER | 8.40148735 | 6.37 | 0.044051 | 0.044051 | 0.36263 | 0.79737 | 0.98213 | |
| NRF7 160709 LINER | 7 | 7/9/2016 | LINER | 8.48645878 | 5.82 | 0.060248 | 0.060248 | 0.2268 | 0.83015 | 0.58729 | |
| NRF8 160715 LINER | 8 | 7/15/2016 | LINER | 43.8062286 | 7.00 | 0.36237893 | 0.38402 | 0.45899 | 1.61066 | 1.66212 | 0.01536 |
| NRF9 160727 LINER | 9 | 7/27/2016 | LINER | 27.2895031 | 6.38 | 0.15202095 | 0.234865 | 0.28858 | 2.12727 | 1.04778 | 0.00093 |
| NRF10 160729 LINER | 10 | 7/29/2016 | LINER | 0.75609863 | 6.41 | 0.0441035 | 0.068868 | 0.1848 | 0.4037 | 0.3611 | 0.00034 |
| NRF11 160730 LINER | 11 | 7/30/2016 | LINER | 5.97864485 | 6.11 | 0.02475905 | 0.052683 | 0.0755 | 0.49398 | 0.41 | 0.00967 |

Acid Anions

| Analysis Name 1 | Storm Number | Date Collected | Sample/ Tray | Conductivity (uS/cm) | pH units | ANC - meas (meq/L) | ANC - calc (meq/L) | Cl- (mg/L) | SO4 ²⁻ (mg/L) | NO3- (mg/L) | PO4 ³⁻ (mg/L) | |
|--------------------|--------------|----------------|--------------|----------------------|------------|--------------------|--------------------|------------|--------------------------|-------------|--------------------------|------------------|
| | | | | | | | | | | | | LINER = geoliner |
| NRF12 | 160805 | LINER | 8/5/2016 | LINER | 6.93966961 | 6.27 | 0.02995178 | 0.071735 | 0.11088 | 0.51561 | 0.30175 | 0.05404 |
| NRF14 | 160808 | LINER | 8/8/2016 | LINER | 3.57608199 | 5.92 | 0.00951593 | 0.037548 | 0.05868 | 0.48399 | 0.13965 | 0.04145 |
| NRF15 | 160809 | LINER | 8/9/2016 | LINER | 5.8248806 | 6.02 | 0.02212968 | 0.041995 | 0.0792 | 1.03403 | 0.15748 | 0.00422 |
| NRF16 | 160816 | LINER | 8/16/2016 | LINER | 5.59423447 | 6.17 | 0.02727201 | 0.04763 | 0.07547 | 0.63149 | 0.19962 | 0.09296 |
| NRF17 | 160818 | LINER | 8/18/2016 | LINER | 6.32334375 | 6.15 | 0.01781111 | 0.038084 | 0.09746 | 0.45007 | 0.31573 | 0.02845 |
| NRF18 | 160822 | LINER | 8/22/2016 | LINER | 9.62853622 | 6.30 | | 0.050836 | 0.187 | 0.76781 | 0.6564 | 0.00192 |
| NRF19 | 160902 | LINER | 9/2/2016 | LINER | 51.6390991 | 6.89 | 0.36874601 | 0.298262 | 1.87603 | 3.53607 | 2.11808 | 0.00358 |
| NRF20 | 160919 | LINER-a | 9/19/2016 | LINER | 15.0916004 | 6.68 | 0.11000025 | 0.183599 | 0.20406 | 0.0405 | 0.5106 | |
| NRF22 | 161121 | LINER | 11/21/2016 | LINER | | | | 2.525085 | 27.0008 | 0.99235 | 7.61023 | |
| NRF23 | 161130 | LINER | 11/30/2016 | LINER | 11.2089491 | 6.51 | | 0.028663 | 0.58175 | 0.39569 | 3.46033 | |
| NRF24 | 161204 | LINER | 12/4/2016 | LINER | 2.81300211 | 5.99 | 0.00899264 | 0.015534 | 0.16345 | 0.1229 | 0.46509 | |
| NRF25 | 161206 | LINER | 12/6/2016 | LINER | 4.15027523 | 6.15 | 0.01435639 | -0.008121 | 0.19913 | 0.08424 | 0.34022 | |
| NRF26 | 161212 | LINER | 12/12/2016 | LINER | 7.44786882 | 6.24 | 0.02352318 | 0.026324 | 0.23099 | 0.37657 | 1.02077 | |
| ROCK = pyrite rock | | | | | | | | | | | | |
| NRF1 | 160621 | ROCK-b | 6/21/2016 | ROCK | 1740 | 3.24 | 12.91692 | | 5.40746 | 6.05609 | | |
| NRF2 | 160624 | ROCK-b | 6/24/2016 | ROCK | 1436.75671 | 3.20 | | | 1.7924 | 2.30808 | | |
| NRF3 | 160627 | ROCK-b | 6/27/2016 | ROCK | 294.054047 | 3.55 | | | 0.27409 | 0.92238 | 48.97087 | |
| NRF4 | 160628 | ROCK-b | 6/28/2016 | ROCK | 1232.43237 | 3.18 | | | 1.02902 | 2.24559 | | 1.13383 |
| --- | | | | | | | | | | | | |
| NRF6 | 160707 | ROCK | 7/7/2016 | ROCK | 1206.56335 | 3.16 | | -0.676464 | 0.9496 | 1.8093 | 392.072 | 0.2295 |
| NRF7 | 160709 | ROCK | 7/9/2016 | ROCK | 857.20929 | 3.24 | | 3.994018 | 0.7792 | 1.5677 | 287.9695 | 0.0277 |
| NRF8 | 160715 | ROCK | 7/15/2016 | ROCK | 1138.06226 | 3.74 | -0.075 | 6.67759 | 0.9286 | 1.7369 | 446.6091 | 0.1043 |
| NRF9 | 160727 | ROCK | 7/27/2016 | ROCK | 1233.10266 | 3.20 | -0.325 | -1.625891 | 0.5363 | 1.788 | 428.915 | 0.099 |
| NRF10 | 160729 | ROCK | 7/29/2016 | ROCK | 866.781982 | 3.31 | -0.2 | 0.734006 | 0.3895 | 0.6843 | 333.7801 | 0.0753 |
| NRF11 | 160730 | ROCK | 7/30/2016 | ROCK | 819.702942 | 3.30 | -0.21 | 3.345513 | 0.4089 | 1.0897 | 258.0309 | 0.053 |
| NRF12 | 160805 | ROCK | 8/5/2016 | ROCK | 790.990112 | 3.31 | -0.2 | 1.254134 | 4.0558 | 6.269 | 307.091 | 0.7741 |
| NRF14 | 160808 | ROCK | 8/8/2016 | ROCK | 583.069336 | 3.43 | -0.16 | 1.109857 | 0.6467 | 0.8456 | 207.0121 | 0.2642 |
| NRF15 | 160809 | ROCK-a | 8/9/2016 | ROCK | 886.039612 | 3.29 | -0.22 | 0.232261 | 1.1371 | 1.837 | 295.5458 | 0.1727 |
| NRF16 | 160816 | ROCK | 8/16/2016 | ROCK | 554.356445 | 3.44 | -0.155 | 0.034275 | 0.2708 | 1.447 | 191.374 | 0.3136 |
| NRF17 | 160818 | ROCK | 8/18/2016 | ROCK | 663.067688 | 3.21 | -0.2966151 | 2.122315 | 20.6841 | 0.7936 | 355.518 | 0.0849 |
| NRF18 | 160822 | ROCK | 8/22/2016 | ROCK | 1280.21057 | 3.00 | -0.6 | 12.24955 | 0.4609 | 0.9827 | 550.5404 | 0.0753 |
| NRF19 | 160902 | ROCK | 9/2/2016 | ROCK | 854.345886 | 3.13 | -0.35 | 4.247233 | 1.8874 | 2.0826 | 386.4143 | 0.1152 |

Acid Anions

| Analysis Name 1 | Storm Number | Date Collected | Sample/ Tray | Conductivity (uS/cm) | pH units | ANC - meas (meq/L) | ANC - calc (meq/L) | Cl- (mg/L) | SO4 ²⁻ (mg/L) | NO3- (mg/L) | PO4 ³⁻ (mg/L) |
|---------------------|--------------|----------------|--------------|----------------------|----------|--------------------|--------------------|------------|--------------------------|-------------|--------------------------|
| | | | | | | | | | | | |
| NRF20 160919 ROCK-a | 20 | 9/19/2016 | ROCK | 163.218048 | 3.58 | -0.1498004 | -0.345117 | 0.10377 | 0.19101 | 94.069 | |
| NRF22 161121 ROCK | 22 | 11/21/2016 | ROCK | 1713.29321 | 3.00 | -0.55 | -18.80789 | 2.25927 | 6.85548 | 1580.574 | |
| NRF23 161130 ROCK | 23 | 11/30/2016 | ROCK | 1005.3233 | 3.09 | -0.42 | -2.476343 | 0.78822 | 1.48687 | 950.288 | |
| NRF24 161204 ROCK | 24 | 12/4/2016 | ROCK | 532.541382 | 3.29 | -0.3072091 | 2.125978 | 0.16087 | 0.78336 | 402.541 | |
| NRF25 161206 ROCK | 25 | 12/6/2016 | ROCK | 347.8797 | 3.38 | -0.1780055 | 0.773986 | 0.27452 | 0.56808 | 232.634 | |
| NRF26 161212 ROCK-b | 26 | 12/12/2016 | ROCK | 303.96991 | 3.39 | -0.1675125 | 0.64671 | 0.29228 | 0.50108 | 165.217 | |
| --- | 1 | 6/21/2016 | | | | | | | | | |
| --- | 2 | 6/24/2016 | | | | | | | | | |
| --- | 3 | 6/27/2016 | | | | | | | | | |
| --- | 4 | 6/28/2016 | | | | | | | | | |
| --- | 5 | 7/5/2016 | | | | | | | | | |
| --- | 6 | 7/7/2016 | | | | | | | | | |
| --- | 7 | 7/9/2016 | | | | | | | | | |
| --- | 8 | 7/15/2016 | | | | | | | | | |
| NRF9 160727 RAIN | 9 | 7/27/2016 | RAIN | 53.125721 | 6.66 | 0.23607969 | 0.434432 | 1.85503 | 2.91275 | 4.93489 | 0.00095 |
| NRF10 160729 RAIN | 10 | 7/29/2016 | RAIN | 1.23170841 | 6.12 | 0.02955578 | 0.101294 | 0.1269 | 0.50466 | 1.18243 | 0.00069 |
| NRF11 160730 RAIN | 11 | 7/30/2016 | RAIN | 14.483716 | 6.43 | 0.07475274 | 0.147367 | 0.04687 | 0.63215 | 0.83645 | 0.01753 |
| NRF12 160805 RAIN | 12 | 8/5/2016 | RAIN | 10.9951954 | 6.27 | 0.03120972 | 0.082372 | 0.02607 | 0.78366 | 0.63925 | 0.01565 |
| NRF14 160808 RAIN | 14 | 8/8/2016 | RAIN | 11.8312874 | 6.43 | 0.04772752 | 0.138872 | 0.05528 | 0.52884 | 0.23147 | 0.01866 |
| NRF15 160809 RAIN | 15 | 8/9/2016 | RAIN | 15.2237053 | 6.42 | 0.06021989 | 0.120993 | 0.03565 | 1.15239 | 0.25884 | 0.01237 |
| NRF16 160816 RAIN | 16 | 8/16/2016 | RAIN | 9.37106323 | 6.23 | 0.03168127 | 0.128653 | 0.06106 | 0.58057 | 0.51088 | 0.01336 |
| NRF17 160818 RAIN | 17 | 8/18/2016 | RAIN | 13.6707478 | 6.23 | 0.03961918 | 0.127352 | 0.0691 | 0.67779 | 1.42658 | 0.055 |
| NRF18 160822 RAIN | 18 | 8/22/2016 | RAIN | 21.5639095 | 6.33 | 0.05724778 | 0.120231 | 0.22743 | 1.68284 | 1.54781 | 0.01925 |
| NRF19 160902 RAIN | 19 | 9/2/2016 | RAIN | 75.4586487 | 6.96 | 0.41198784 | 1.210281 | 2.9456 | 14.09757 | 11.85447 | 0.00408 |
| NRF20 160919 RAIN-a | 20 | 9/19/2016 | RAIN | | | | 0.248023 | 0.14428 | 1.39996 | 1.25647 | |
| NRF22 161121 RAIN | 22 | 11/21/2016 | RAIN | 24.0902252 | 6.79 | 0.14582506 | 3.494509 | 12.1473 | 88.7762 | 47.6 | |
| --- | 23 | 11/30/2016 | | | | | | | | | |
| NRF24 161204 RAIN | 24 | 12/4/2016 | RAIN | 3.68678856 | 5.89 | 0.00363026 | -0.003659 | 0.05744 | 0.37581 | 0.49922 | |
| NRF25 161206 RAIN | 25 | 12/6/2016 | RAIN | 5.19122076 | 5.92 | 0.00767107 | -0.005784 | 0.39371 | 0.241 | 0.66911 | |
| NRF26 161212 RAIN | 26 | 12/12/2016 | RAIN | 4.34022903 | 6.06 | 12.0880893 | 0.039117 | 0.13173 | 0.21036 | 0.41256 | |

| Storm Number | Sample/ Tray | Base Cations | | | | | Dissolved Metals | | | | | | | | | | | | | | | | | | |
|--------------------|-----------------|---------------|--------------|----------------------------|----------------------------|--------------|------------------|--------------|--------------|--------------|--------------|--------------|--------------|--|--|--|--|--|--|--|--|--|--|--|--|
| | | Na+ (mg/L) | K+ (mg/L) | Mg ²⁺ (mg/L) | Ca ²⁺ (mg/L) | Al (mg/L) | Cu (mg/L) | Fe (mg/L) | Mn (mg/L) | Si (mg/L) | Zn (mg/L) | Cd (mg/L) | Ni (mg/L) | | | | | | | | | | | | |
| LINER = geoliner | | | | | | | | | | | | | | | | | | | | | | | | | |
| 12 | LINER | 0.371597 | 0.220222 | 0.122897 | 1.008865 | 0.011602 | 0.000378 | 0.177542 | 0.002386 | 0.025828 | 0.007042 | 0 | 0.000583 | | | | | | | | | | | | |
| 14 | LINER | 0.331983 | 0.067654 | 0.075835 | 0.532092 | 0.003867 | 0.000651 | 0.066472 | 0.003422 | 0.008586 | 0.008725 | 0 | 2.82E-05 | | | | | | | | | | | | |
| 15 | LINER | 0.344846 | 0.170638 | 0.125071 | 0.700015 | 0.003961 | 4.39E-05 | 0.003928 | 0.005236 | 0.039699 | 0.006037 | 0 | 0 | | | | | | | | | | | | |
| 16 | LINER | 0.372377 | 0.188078 | 0.080725 | 0.778106 | 0.001943 | 0.00038 | 0.002049 | 0.004253 | 0.026363 | 0.005809 | 0 | 0 | | | | | | | | | | | | |
| 17 | LINER | 0.350956 | 0.128147 | 0.082536 | 0.567466 | 0.001433 | 0 | 0.00227 | 0.001669 | 0.029246 | 0.003595 | 0 | 0 | | | | | | | | | | | | |
| 18 | LINER | 0.360567 | 0.081294 | 0.169532 | 0.941379 | 0.006159 | 0.000435 | 0.036027 | 0.005261 | 0.022647 | 0.005852 | 0 | 0.000215 | | | | | | | | | | | | |
| 19 | LINER | 0.347854 | 0.667149 | 0.459798 | 7.513653 | 0.019792 | 0.002172 | 0.013416 | 0.010277 | 0.109905 | 0.081591 | 0 | 0 | | | | | | | | | | | | |
| 20 | LINER | 0.146949 | 0.099756 | 0.287088 | 3.098123 | 0.010133 | 0.001334 | 0.050554 | 0.00304 | 0.10444 | 0.012424 | 0.0013 | 0.000923 | | | | | | | | | | | | |
| 22 | LINER | 0.727401 | 5.028546 | 1.586033 | 62.10518 | 0.013011 | 0.005999 | 0.033135 | 0.004291 | 0.406981 | 0.094137 | 0.001147 | 0.000712 | | | | | | | | | | | | |
| 23 | LINER | 0.058322 | 0.012101 | 0.090045 | 1.973911 | 0 | 0.000393 | 0 | 0.001774 | 0.003988 | 0 | 0 | 0 | | | | | | | | | | | | |
| 24 | LINER | 0.059717 | 0.085253 | 0.009477 | 0.46166 | 0.003971 | 0.002555 | 0.017319 | 0.002726 | 0.009356 | 0 | 0.001075 | | | | | | | | | | | | | |
| 25 | LINER | 0.083614 | 0 | 0 | 0.002762 | 0.002044 | 0.015268 | 0.001368 | | 0.004192 | 0 | 0.000444 | | | | | | | | | | | | | |
| 26 | LINER | 0.042969 | 0.002011 | 0.067935 | 0.97682 | 0.000989 | 0.001441 | 0.011783 | 0.002478 | 0.004538 | 0 | 0 | 0 | | | | | | | | | | | | |
| ROCK = pyrite rock | | | | | | | | | | | | | | | | | | | | | | | | | |
| 1 | ROCK | 3.839908 | 3.968077 | 35.66902 | 52.71912 | 46.32671 | 1.398215 | | 45.40168 | 3.077125 | 6.297525 | 0 | 2.81132 | | | | | | | | | | | | |
| 2 | ROCK | | | | | | | | | | | | | | | | | | | | | | | | |
| 3 | ROCK | | | | | | | | | | | | | | | | | | | | | | | | |
| 4 | ROCK | | | | | | | | | | | | | | | | | | | | | | | | |
| 5 | ROCK | | | | | | | | | | | | | | | | | | | | | | | | |
| 6 | ROCK | 0.314746 | 0.6121 | 17.79009 | 12.85554 | 24.5471 | 0.645928 | | 17.1065 | 1.292949 | 2.374941 | 0 | 1.334167 | | | | | | | | | | | | |
| 7 | ROCK | | 0.544492 | 11.21477 | 6.43542 | 13.75159 | 0.381491 | 148.4558 | 12.47733 | 0.670452 | 1.541893 | 0 | 0.838491 | | | | | | | | | | | | |
| 8 | ROCK | 3.260131 | 0.692026 | 27.2145 | 48.03558 | 20.25336 | 0.455586 | 172.9658 | 14.25658 | 1.04708 | 1.603623 | 0 | 1.234955 | | | | | | | | | | | | |
| 9 | ROCK | 3.089653 | 0.199151 | 9.751806 | 14.7933 | 8.434192 | 0.195214 | 70.34808 | 3.982088 | 0.285119 | 0.613547 | 0 | 0.54289 | | | | | | | | | | | | |
| 10 | ROCK | 3.006577 | 0.136893 | 12.2039 | 10.99358 | 12.53932 | 0.237065 | 77.55281 | 5.734315 | 0.386081 | 0.666502 | 0 | 0.617715 | | | | | | | | | | | | |
| 11 | ROCK | 0.359602 | 0.160296 | 15.8737 | 13.38341 | 13.0977 | 0.26907 | 104.6397 | 7.044023 | 0.480208 | 0.775254 | 0 | 0.620965 | | | | | | | | | | | | |
| 12 | ROCK | 3.14313 | 0.809213 | 14.38856 | 9.66663 | 12.95615 | 0.267781 | 80.8625 | 6.163183 | 0.608876 | 0.62589 | 0 | 0.586138 | | | | | | | | | | | | |
| 14 | ROCK | 3.296538 | 0.782166 | 9.809672 | 5.784525 | 10.12979 | 0.155798 | 52.76719 | 4.269786 | 0.405144 | 0.457319 | 0 | 0.401821 | | | | | | | | | | | | |
| 15 | ROCK | 3.013274 | 0.398691 | 11.15768 | 5.87459 | 10.66494 | 0.193995 | 62.09984 | 6.420899 | 0.50709 | 0.618295 | 0 | 0.497285 | | | | | | | | | | | | |
| 16 | ROCK | 3.071201 | 0.150875 | 7.075079 | 4.155928 | 6.817229 | 0.137614 | 37.26577 | 2.779691 | 0.321695 | 0.334884 | 0 | 0.28858 | | | | | | | | | | | | |
| 17 | ROCK | 2.945831 | 0.246112 | 19.65834 | 7.251543 | 19.06898 | 0.333947 | 106.0607 | 8.727159 | 0.655154 | 0.880168 | 0 | 0.720146 | | | | | | | | | | | | |
| 18 | ROCK | 3.054048 | 0.523546 | 45.73868 | 14.63228 | 47.29316 | 0.997017 | 284.269 | 22.41891 | 1.424479 | 2.394405 | 0 | 1.799145 | | | | | | | | | | | | |
| 19 | ROCK | 2.944074 | 0.443991 | 27.34733 | 24.55974 | 19.75775 | 0.462097 | 117.4688 | 11.59705 | 0.742434 | 1.207312 | 0 | 0.92432 | | | | | | | | | | | | |

| Storm Number | Sample/ Tray | Base Cations | | | | Dissolved Metals | | | | | | | | | | |
|--------------|--------------|--------------------|-----------|-------------------------|-------------------------|------------------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|--|--|--|
| | | Na+ (mg/L) | K+ (mg/L) | Mg ²⁺ (mg/L) | Ca ²⁺ (mg/L) | Al (mg/L) | Cu (mg/L) | Fe (mg/L) | Mn (mg/L) | Si (mg/L) | Zn (mg/L) | Cd (mg/L) | Ni (mg/L) | | | |
| | | ROCK = pyrite rock | | | | | | | | | | | | | | |
| 20 | ROCK | 0.158501 | 0.129737 | 2.772098 | 3.917612 | 2.687545 | 0.070936 | 10.4099 | 1.233184 | 0.203304 | 0.188959 | 0.000936 | 0.138948 | | | |
| 22 | ROCK | 0.504326 | 0.799184 | | 101.3095 | 2.560751 | | 35.9686 | 1.49032 | 5.130391 | | 0 | 3.845403 | | | |
| 23 | ROCK | 0.349126 | 0.144248 | 51.32655 | 23.30369 | 59.21431 | 1.592778 | 17.49483 | 1.032532 | 2.725746 | | 0 | 1.982335 | | | |
| 24 | ROCK | 0.01586 | 0.001323 | 18.77498 | 11.04354 | 20.72401 | 0.655818 | 103.3565 | 0.992004 | 1.366541 | | 0 | 0.80723 | | | |
| 25 | ROCK | 0.151851 | 0.042656 | 9.594442 | 7.10817 | 9.113009 | 0.255727 | 57.05916 | 6.613928 | 0.740303 | 0.696683 | 0 | 0.391666 | | | |
| 26 | ROCK | 0.117812 | 0.052994 | 7.35809 | 5.681438 | 6.127255 | 0.250974 | 42.00211 | 5.037224 | 0.413363 | 0.552389 | 0 | 0.319786 | | | |
| | | RAIN = rainwater | | | | | | | | | | | | | | |
| 1 | | | | | | | | | | | | | | | | |
| 2 | | | | | | | | | | | | | | | | |
| 3 | | | | | | | | | | | | | | | | |
| 4 | | | | | | | | | | | | | | | | |
| 5 | | | | | | | | | | | | | | | | |
| 6 | | | | | | | | | | | | | | | | |
| 7 | | | | | | | | | | | | | | | | |
| 8 | | | | | | | | | | | | | | | | |
| 9 | RAIN | 1.440875 | 0.821527 | 0.942017 | 8.045391 | 0.110978 | 0.055931 | 0.993071 | 0.081916 | 0.145034 | 0.012654 | | 0.007841 | | | |
| 10 | RAIN | 0.29173 | 0.070591 | 0.222089 | 1.014446 | 0.129416 | 0.009072 | 0.870628 | 0.062435 | 0.033668 | 0.009789 | | 0.004856 | | | |
| 11 | RAIN | 0.331128 | 0.07507 | 0.203683 | 2.694166 | 0.013742 | 0.001688 | 0.087958 | 0.009861 | 0.04076 | 0.003781 | | 9.19E-05 | | | |
| 12 | RAIN | 0.32825 | 0.061846 | 0.14762 | 1.533492 | 0.00484 | 0.000487 | 0.022199 | 0.010354 | 0.051621 | 0.005009 | | 8.9E-05 | | | |
| 14 | RAIN | 0.305912 | 0.07288 | 0.094566 | 2.610109 | 0.004526 | 0.00118 | 0.007111 | 0.00241 | 0.018574 | 0.003628 | | | | | |
| 15 | RAIN | 0.287812 | 0.054855 | 0.143092 | 2.426111 | 0.004932 | | 0.020695 | 0.003053 | 0.028379 | 0.004879 | | | | | |
| 16 | RAIN | 0.327961 | 0.03827 | 0.183092 | 2.127724 | 0.026379 | 0.001217 | 0.222885 | 0.022406 | 0.03454 | 0.005338 | | 0.001382 | | | |
| 17 | RAIN | 0.317877 | 0.034259 | 0.25103 | 1.943892 | 0.086119 | 0.002099 | 0.587255 | 0.059592 | 0.033752 | 0.007137 | | 0.004471 | | | |
| 18 | RAIN | 0.400698 | 0.194189 | 0.262899 | 2.691584 | 0.010616 | 0.001424 | 0.033928 | 0.007995 | 0.081504 | 0.004417 | | | | | |
| 19 | RAIN | 0.474291 | 5.405494 | 2.169304 | 27.48096 | 0.124082 | 0.034679 | 0.451934 | 0.079458 | 0.444757 | 0.032353 | | 0.002399 | | | |
| 20 | RAIN | 0.233869 | 0.236458 | 0.394589 | 4.930591 | 0.00462 | 0.001938 | 0.003314 | 0.000848 | 0.072182 | 0.013074 | 0.000957 | 0.000362 | | | |
| 22 | RAIN | 3.202364 | 5.514781 | 5.579394 | 110.627 | 0.023299 | 0.025315 | 0.008031 | 0.000924 | 2.510287 | 0.040334 | 0.001045 | 0.000296 | | | |
| 23 | | | | | | | | | | | | | | | | |
| 24 | RAIN | | 0.026354 | 0.01866 | 0.212922 | 0.000774 | 0.001173 | 0.011499 | 0.003109 | 0 | 0.010853 | | 0.000214 | | | |
| 25 | RAIN | 0.239919 | 0.134501 | 0.011581 | 0.104914 | 0.002411 | 0.002252 | 0.00931 | 0.001687 | 0 | 0.010562 | 1.02E-05 | 0.000442 | | | |
| 26 | RAIN | 0.027054 | 0.481731 | 0.057722 | 0.699416 | | 0.000708 | 0.010306 | 0.004262 | 0 | 0.002535 | 6.22E-05 | 3.08E-06 | | | |

| Storm Number | Sample/ Tray | Pb (m/L) | Acid Anions | | | | | Base Cations | | | | | Sum BC (meq/L) |
|--------------|--------------------|----------|-------------|----------------|--------------|----------------|----------------|--------------|------------|----------------|----------------|----------|----------------|
| | | | Cl- (meq/L) | SO4^2- (meq/L) | NO3- (meq/L) | PO4^3- (meq/L) | Sum AA (meq/L) | Na+ (meq/L) | K+ (meq/L) | Mg ^2+ (meq/L) | Ca ^2+ (meq/L) | | |
| | LINER = geoliner | | | | | | | | | | | | |
| 12 | LINER | 0.000499 | 0.003132 | 0.010742 | 0.004867 | 0.001707 | 0.020448 | 0.016156 | 0.005647 | 0.010115 | 0.050443 | 0.082361 | |
| 14 | LINER | 0.000498 | 0.001658 | 0.010083 | 0.002252 | 0.001309 | 0.015302 | 0.014434 | 0.001735 | 0.006242 | 0.026605 | 0.049015 | |
| 15 | LINER | 0.00014 | 0.002237 | 0.021542 | 0.00254 | 0.000133 | 0.026453 | 0.014993 | 0.004375 | 0.010294 | 0.035001 | 0.064663 | |
| 16 | LINER | 0 | 0.002132 | 0.013156 | 0.00322 | 0.002936 | 0.021443 | 0.01619 | 0.004823 | 0.006644 | 0.038905 | 0.066562 | |
| 17 | LINER | 0 | 0.002753 | 0.009376 | 0.005092 | 0.000898 | 0.01812 | 0.015259 | 0.003286 | 0.006793 | 0.028373 | 0.053711 | |
| 18 | LINER | 0.000685 | 0.005282 | 0.015996 | 0.010587 | 6.06E-05 | 0.031926 | 0.015677 | 0.002084 | 0.013953 | 0.047069 | 0.078783 | |
| 19 | LINER | 0.001145 | 0.052995 | 0.073668 | 0.034163 | 0.000113 | 0.160939 | 0.015124 | 0.017106 | 0.037843 | 0.375683 | 0.445757 | |
| 20 | LINER | 0.001247 | 0.005764 | 0.000844 | 0.008235 | 0 | 0.014844 | 0.006389 | 0.002558 | 0.023629 | 0.154906 | 0.187482 | |
| 22 | LINER | 0.001672 | 0.762734 | 0.020674 | 0.122746 | 0 | 0.906154 | 0.031626 | 0.128937 | 0.130538 | 3.105259 | 3.39636 | |
| 23 | LINER | 0 | 0.016434 | 0.008244 | 0.055812 | 0 | 0.080489 | 0.002536 | 0.00031 | 0.007411 | 0.098696 | 0.108953 | |
| 24 | LINER | 0 | 0.004617 | 0.00256 | 0.007501 | 0 | 0.014679 | 0.002596 | 0.002186 | 0.00078 | 0.023083 | 0.028645 | |
| 25 | LINER | 0 | 0.005625 | 0.001755 | 0.005487 | 0 | 0.012868 | 0.003635 | 0 | 0 | 0 | 0.003635 | |
| 26 | LINER | 0 | 0.006525 | 0.007845 | 0.016464 | 0 | 0.030834 | 0.001868 | 5.16E-05 | 0.005591 | 0.048841 | 0.056352 | |
| | ROCK = pyrite rock | | | | | | | | | | | | |
| 1 | ROCK | 0.039411 | 0.152753 | 0.126169 | 0 | 0 | 0.278922 | 0.166953 | 0.101746 | 2.935722 | 2.635956 | 5.840375 | |
| 2 | ROCK | | | | | | | | | | | | |
| 3 | ROCK | | | | | | | | | | | | |
| 4 | ROCK | | | | | | | | | | | | |
| 5 | ROCK | | | | | | | | | | | | |
| 6 | ROCK | 0.012466 | 0.026825 | 0.037694 | 6.323742 | 0.007247 | 6.395508 | 0.013685 | 0.015695 | 1.464205 | 0.642777 | 2.136361 | |
| 7 | ROCK | 0.009122 | 0.022011 | 0.03266 | 4.644669 | 0.000875 | 4.700216 | 0 | 0.013961 | 0.923026 | 0.321771 | 1.258759 | |
| 8 | ROCK | 0.010603 | 0.026232 | 0.036185 | 7.203373 | 0.003294 | 7.269083 | 0.141745 | 0.017744 | 2.239876 | 2.401779 | 4.801144 | |
| 9 | ROCK | 0.02039 | 0.01515 | 0.03725 | 6.917984 | 0.003126 | 6.97351 | 0.134333 | 0.005106 | 0.802618 | 0.739665 | 1.681722 | |
| 10 | ROCK | 0.00863 | 0.011003 | 0.014256 | 5.38355 | 0.002378 | 5.411187 | 0.130721 | 0.00351 | 1.004436 | 0.549679 | 1.688346 | |
| 11 | ROCK | 0.005648 | 0.011551 | 0.022702 | 4.161789 | 0.001674 | 4.197715 | 0.015635 | 0.00411 | 1.306477 | 0.669171 | 1.995393 | |
| 12 | ROCK | 0.013977 | 0.114571 | 0.130604 | 4.953081 | 0.024445 | 5.222701 | 0.136658 | 0.020749 | 1.184244 | 0.483332 | 1.824982 | |
| 14 | ROCK | 0.01019 | 0.018268 | 0.017617 | 3.338905 | 0.008343 | 3.383133 | 0.143328 | 0.020056 | 0.80738 | 0.289226 | 1.25999 | |
| 15 | ROCK | 0.012275 | 0.032121 | 0.038271 | 4.766868 | 0.005454 | 4.842714 | 0.131012 | 0.010223 | 0.918327 | 0.293729 | 1.353292 | |
| 16 | ROCK | 0.002433 | 0.00765 | 0.030146 | 3.086677 | 0.009903 | 3.134376 | 0.13353 | 0.003869 | 0.582311 | 0.207796 | 0.927507 | |
| 17 | ROCK | 0.001667 | 0.584297 | 0.016533 | 5.734161 | 0.002681 | 6.337672 | 0.12808 | 0.006311 | 1.617971 | 0.362577 | 2.114938 | |
| 18 | ROCK | 0.017801 | 0.01302 | 0.020473 | 8.879684 | 0.002378 | 8.915554 | 0.132785 | 0.013424 | 3.7645 | 0.731614 | 4.642323 | |
| 19 | ROCK | 0.021411 | 0.053316 | 0.043388 | 6.232489 | 0.003638 | 6.33283 | 0.128003 | 0.011384 | 2.250809 | 1.227987 | 3.618184 | |

| Storm Number | Sample/ Tray | Pb (m/L) | Acid Anions | | | | | Base Cations | | | | | Sum BC (meq/L) |
|--------------|--------------------|----------|-------------|----------------|--------------|----------------|----------------|--------------|------------|----------------|----------------|----------|----------------|
| | | | Cl- (meq/L) | SO4^2- (meq/L) | NO3- (meq/L) | PO4^3- (meq/L) | Sum AA (meq/L) | Na+ (meq/L) | K+ (meq/L) | Mg ^2+ (meq/L) | Ca ^2+ (meq/L) | | |
| | ROCK = pyrite rock | | | | | | | | | | | | |
| 20 | ROCK | 0.002647 | 0.002931 | 0.003979 | 1.517242 | 0 | 1.524153 | 0.006891 | 0.003327 | 0.228156 | 0.195881 | 0.434255 | |
| 22 | ROCK | 0.038443 | 0.063821 | 0.142823 | 25.49313 | 0 | 25.69977 | 0.021927 | 0.020492 | 0 | 5.065474 | 5.107893 | |
| 23 | ROCK | | 0.022266 | 0.030976 | 15.32723 | 0 | 15.38047 | 0.015179 | 0.003699 | 4.224407 | 1.165184 | 5.40847 | |
| 24 | ROCK | | 0.004544 | 0.01632 | 6.492597 | 0 | 6.513461 | 0.00069 | 3.39E-05 | 1.545265 | 0.552177 | 2.098166 | |
| 25 | ROCK | | 0.007755 | 0.011835 | 3.752161 | 0 | 3.771751 | 0.006602 | 0.001094 | 0.789666 | 0.355408 | 1.15277 | |
| 26 | ROCK | | 0.008256 | 0.010439 | 2.66479 | 0 | 2.683486 | 0.005122 | 0.001359 | 0.605604 | 0.284072 | 0.896157 | |
| | RAIN = rainwater | | | | | | | | | | | | |
| 1 | | | | | | | | | | | | | |
| 2 | | | | | | | | | | | | | |
| 3 | | | | | | | | | | | | | |
| 4 | | | | | | | | | | | | | |
| 5 | | | | | | | | | | | | | |
| 6 | | | | | | | | | | | | | |
| 7 | | | | | | | | | | | | | |
| 8 | | | | | | | | | | | | | |
| 9 | RAIN | | 0.052402 | 0.060682 | 0.079595 | 0.00003 | 0.192709 | 0.062647 | 0.021065 | 0.077532 | 0.40227 | 0.563513 | |
| 10 | RAIN | 0.000665 | 0.003585 | 0.010514 | 0.019071 | 2.18E-05 | 0.033192 | 0.012684 | 0.00181 | 0.018279 | 0.050722 | 0.083495 | |
| 11 | RAIN | 0.000995 | 0.001324 | 0.01317 | 0.013491 | 0.000554 | 0.028539 | 0.014397 | 0.001925 | 0.016764 | 0.134708 | 0.167794 | |
| 12 | RAIN | 0.000501 | 0.000736 | 0.016326 | 0.01031 | 0.000494 | 0.027867 | 0.014272 | 0.001586 | 0.01215 | 0.076675 | 0.104682 | |
| 14 | RAIN | | 0.001562 | 0.011018 | 0.003733 | 0.000589 | 0.016902 | 0.013301 | 0.001869 | 0.007783 | 0.130505 | 0.153458 | |
| 15 | RAIN | 0.000594 | 0.001007 | 0.024008 | 0.004175 | 0.000391 | 0.029581 | 0.012514 | 0.001407 | 0.011777 | 0.121306 | 0.147003 | |
| 16 | RAIN | | 0.001725 | 0.012095 | 0.00824 | 0.000422 | 0.022482 | 0.014259 | 0.000981 | 0.015069 | 0.106386 | 0.136696 | |
| 17 | RAIN | 0.000882 | 0.001952 | 0.014121 | 0.023009 | 0.001737 | 0.040819 | 0.013821 | 0.000878 | 0.020661 | 0.097195 | 0.132555 | |
| 18 | RAIN | 0.000704 | 0.006425 | 0.035059 | 0.024965 | 0.000608 | 0.067056 | 0.017422 | 0.004979 | 0.021638 | 0.134579 | 0.178618 | |
| 19 | RAIN | 0.002155 | 0.083209 | 0.293699 | 0.191201 | 0.000129 | 0.568238 | 0.020621 | 0.138602 | 0.178544 | 1.374048 | 1.711815 | |
| 20 | RAIN | 0.001452 | 0.004076 | 0.029166 | 0.020266 | 0 | 0.053507 | 0.010168 | 0.006063 | 0.032476 | 0.24653 | 0.295237 | |
| 22 | RAIN | 0.002566 | 0.343144 | 1.849504 | 0.767742 | 0 | 2.96039 | 0.139233 | 0.141405 | 0.459209 | 5.531349 | 6.271197 | |
| 23 | | | | | | | | | | | | | |
| 24 | RAIN | | 0.001623 | 0.007829 | 0.008052 | 0 | 0.017504 | 0 | 0.000676 | 0.001536 | 0.010646 | 0.012858 | |
| 25 | RAIN | | 0.011122 | 0.005021 | 0.010792 | 0 | 0.026935 | 0.010431 | 0.003449 | 0.000953 | 0.005246 | 0.020079 | |
| 26 | RAIN | | 0.003721 | 0.004383 | 0.006654 | 0 | 0.014758 | 0.001176 | 0.012352 | 0.004751 | 0.034971 | 0.05325 | |

Dissolved Metals

| Storm Number | Sample/ Tray | Al ³⁺ (meq/L) | Fe ²⁺ (meq/L) | Mn ²⁺ (meq/L) | Zn ²⁺ (meq/L) | Cu ²⁺ (meq/L) | Cd ²⁺ (meq/L) | Ni ²⁺ (meq/L) | Si ²⁺ (meq/L) | ANC (meq/L) | |
|--------------|-------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|-------------|--|
| | GRASS = soil/vegetation | | | | | | | | | | |
| 5 | | | | | | | | | | | |
| 6 | GRASS | 0.058942 | 0.056524 | 0.105073 | 0.007184 | 0.001042 | 1.67E-08 | 0.001893 | 0.182588 | -1.42564 | |
| 7 | GRASS | 0.079225 | 0.035644 | 0.212057 | 0.0138 | 0.000565 | 6.82E-07 | 0.00411 | 0.370529 | -6.94923 | |
| 8 | GRASS | 0.035102 | 0.063985 | 0.090734 | 0.002831 | 0.000404 | 0 | 0.001588 | 0.067415 | -0.89816 | |
| 9 | GRASS | 0.060208 | 0.08193 | 0.09263 | 0.002601 | 0.000755 | 0 | 0.001242 | 0.066758 | 1.370314 | |
| 10 | GRASS | 0.063868 | 0.099556 | 0.062986 | 0.002664 | 0.000509 | 0 | 0.001233 | 0.110301 | 0.34462 | |
| 11 | GRASS | 0.077889 | 0.1187 | 0.067785 | 0.002443 | 0.00046 | 1E-06 | 0.001235 | 0.152751 | -0.45801 | |
| 12 | GRASS | 0.055359 | 0.075279 | 0.060258 | 0.001662 | 0.000416 | 0 | 0.000788 | 0.125928 | -17.532 | |
| 14 | GRASS | 0.045488 | 0.052957 | 0.042341 | 0.001932 | 0.000342 | 0 | 0.001074 | 0.095445 | 0.234067 | |
| 15 | GRASS | 0.144078 | 0.075671 | 0.086728 | 0.003992 | 0.000505 | 0 | 0.002446 | 0.190824 | 0.536082 | |
| 16 | GRASS | 0.050505 | 0.052324 | 0.043451 | 0.001396 | 0.000263 | 0 | 0.00079 | 0.092069 | 0.379939 | |
| 17 | GRASS | 0.078154 | 0.057851 | 0.050213 | 0.002292 | 0.000349 | 0 | 0.001294 | 0.130581 | 0.43417 | |
| 18 | GRASS | 0.176116 | 0.311779 | 0.099414 | 0.00438 | 0.001167 | 0 | 0.00292 | 0.258855 | 1.417363 | |
| 19 | GRASS | 0.015143 | 0.060251 | 0.146076 | 0.001514 | 0.000276 | 0 | 0.001099 | 0.061038 | 0.957684 | |
| 20 | GRASS | 0.088117 | 0.071935 | 0.043917 | 0.003039 | 0.00088 | 2.05E-05 | 0.001324 | 0.176091 | 0.548806 | |
| 22 | GRASS | 0.007713 | 0.017689 | 0.392343 | 0.010935 | 0.000388 | 1.91E-05 | 0.005419 | 0.123602 | 0.179933 | |
| 23 | GRASS | 0.034075 | 0.091279 | 0.051046 | 0.001677 | 0.000558 | 0 | 0.000665 | 0.142042 | 0.689625 | |
| 24 | GRASS | 0.084307 | 0.160969 | 0.0578 | 0.002796 | 0.000783 | 0 | 0.00146 | 0.152661 | 0.613787 | |
| 25 | GRASS | 0.103968 | 0.225586 | 0.085118 | 0.004704 | 0.000984 | 0 | 0.00243 | 0.176371 | 0.318336 | |
| 26 | GRASS | 0.114791 | 0.153425 | 0.057771 | 0.0049 | 0.000872 | 0 | 0.002459 | 0.097879 | 0.421292 | |
| | LINER = geoliner | | | | | | | | | | |
| 1 | LINER | 0.003 | 0.002869 | 0.002213 | 0.004106 | 0.000744 | 0 | 3.59E-05 | 0.009637 | 0.269335 | |
| 2 | LINER | | | | | | | | | | |
| 3 | LINER | | | | | | | | | | |
| 4 | LINER | | | | | | | | | | |
| 5 | LINER | 0.000559 | 0.001222 | 0.000615 | 0.000839 | 0.000399 | 0 | 2.52E-06 | 0.00563 | 0.101643 | |
| 6 | LINER | 0.00026 | 0.000854 | 0.001699 | 0.000228 | 2.53E-05 | 0 | 1.23E-05 | 0.004402 | 0.044051 | |
| 7 | LINER | 0.00679 | 0.014324 | 0.001604 | 0.000591 | 2.72E-05 | 0 | 9.03E-05 | 0.002605 | 0.060248 | |
| 8 | LINER | 0.001695 | 0.000106 | 6.94E-05 | 0.000131 | 6.18E-05 | 0 | 0 | 0.006735 | 0.38402 | |
| 9 | LINER | 0.004318 | 0.007592 | 0.001327 | 0.000408 | 9.4E-05 | 0 | 4.11E-05 | 0.006573 | 0.234865 | |
| 10 | LINER | 0.00074 | 0.001176 | 0.00035 | 0.000148 | 3.98E-05 | 0 | 0 | 0.002091 | 0.068868 | |
| 11 | LINER | 0.00044 | 0.000761 | 0.000178 | 0.000222 | 3.07E-05 | 0 | 0 | 0.001904 | 0.052683 | |

Dissolved Metals

| Storm Number | Sample/ Tray | Al ³⁺ (meq/L) | Fe ²⁺ (meq/L) | Mn ²⁺ (meq/L) | Zn ²⁺ (meq/L) | Cu ²⁺ (meq/L) | Cd ²⁺ (meq/L) | Ni ²⁺ (meq/L) | Si ²⁺ (meq/L) | ANC (meq/L) |
|--------------------|-----------------|-----------------------------|-----------------------------|-----------------------------|-----------------------------|-----------------------------|-----------------------------|-----------------------------|-----------------------------|----------------|
| LINER = geoliner | | | | | | | | | | |
| 12 | LINER | 0.00129 | 0.006358 | 8.69E-05 | 0.000215 | 1.19E-05 | 0 | 1.98E-05 | 0.001839 | 0.071735 |
| 14 | LINER | 0.00043 | 0.002381 | 0.000125 | 0.000267 | 2.05E-05 | 0 | 9.62E-07 | 0.000611 | 0.037548 |
| 15 | LINER | 0.00044 | 0.000141 | 0.000191 | 0.000185 | 1.38E-06 | 0 | 0 | 0.002827 | 0.041995 |
| 16 | LINER | 0.000216 | 7.34E-05 | 0.000155 | 0.000178 | 1.19E-05 | 0 | 0 | 0.001877 | 0.04763 |
| 17 | LINER | 0.000159 | 8.13E-05 | 6.08E-05 | 0.00011 | 0 | 0 | 0 | 0.002082 | 0.038084 |
| 18 | LINER | 0.000685 | 0.00129 | 0.000192 | 0.000179 | 1.37E-05 | 0 | 7.32E-06 | 0.001612 | 0.050836 |
| 19 | LINER | 0.002201 | 0.00048 | 0.000374 | 0.002495 | 6.84E-05 | 0 | 0 | 0.007825 | 0.298262 |
| 20 | LINER | 0.001127 | 0.001811 | 0.000111 | 0.00038 | 4.2E-05 | 2.31E-05 | 3.14E-05 | 0.007436 | 0.183599 |
| 22 | LINER | 0.001447 | 0.001187 | 0.000156 | 0.002879 | 0.000189 | 2.04E-05 | 2.43E-05 | 0.028977 | 2.525085 |
| 23 | LINER | 0 | 0 | 6.46E-05 | 0.000122 | 1.24E-05 | 0 | 0 | 0 | 0.028663 |
| 24 | LINER | 0.000442 | 0.00062 | 9.92E-05 | 0.000286 | 8.35E-05 | 0 | 3.66E-05 | 0 | 0.015534 |
| 25 | LINER | 0.000307 | 0.000547 | 4.98E-05 | 0.000128 | 6.43E-05 | 0 | 1.51E-05 | 0 | -0.00812 |
| 26 | LINER | 0.00011 | 0.000422 | 9.02E-05 | 0.000139 | 4.54E-05 | 0 | 0 | 0 | 0.026324 |
| ROCK = pyrite rock | | | | | | | | | | |
| 1 | ROCK | 5.151228 | 0 | 1.652773 | 0.192585 | 0.044004 | 0 | 0.095786 | 0.21909 | 12.91692 |
| 2 | ROCK | | | | | | | | | |
| 3 | ROCK | | | | | | | | | |
| 4 | ROCK | | | | | | | | | |
| 5 | ROCK | | | | | | | | | |
| 6 | ROCK | 2.729478 | 0 | 0.622734 | 0.072628 | 0.020328 | 0 | 0.045457 | 0.092058 | -0.67646 |
| 7 | ROCK | 1.529087 | 5.316708 | 0.454217 | 0.047153 | 0.012006 | 0 | 0.028569 | 0.047736 | 3.994018 |
| 8 | ROCK | 2.252042 | 6.194494 | 0.518987 | 0.04904 | 0.014338 | 0 | 0.042077 | 0.074552 | 6.67759 |
| 9 | ROCK | 0.937827 | 2.519405 | 0.144961 | 0.018763 | 0.006144 | 0 | 0.018497 | 0.0203 | -1.62589 |
| 10 | ROCK | 1.39429 | 2.777431 | 0.208748 | 0.020382 | 0.007461 | 0 | 0.021047 | 0.027489 | 0.734006 |
| 11 | ROCK | 1.456379 | 3.747506 | 0.256426 | 0.023708 | 0.008468 | 0 | 0.021157 | 0.034191 | 3.345513 |
| 12 | ROCK | 1.44064 | 2.895962 | 0.22436 | 0.01914 | 0.008427 | 0 | 0.019971 | 0.043352 | 1.254134 |
| 14 | ROCK | 1.126367 | 1.889773 | 0.155435 | 0.013985 | 0.004903 | 0 | 0.013691 | 0.028846 | 1.109857 |
| 15 | ROCK | 1.185872 | 2.224007 | 0.233742 | 0.018908 | 0.006105 | 0 | 0.016943 | 0.036105 | 0.232261 |
| 16 | ROCK | 0.758031 | 1.334614 | 0.10119 | 0.010241 | 0.004331 | 0 | 0.009832 | 0.022905 | 0.034275 |
| 17 | ROCK | 2.120346 | 3.798396 | 0.317698 | 0.026916 | 0.01051 | 0 | 0.024536 | 0.046647 | 2.122315 |
| 18 | ROCK | 5.258691 | 10.18064 | 0.816123 | 0.073223 | 0.031377 | 0 | 0.0613 | 0.101422 | 12.24955 |
| 19 | ROCK | 2.196933 | 4.206957 | 0.422171 | 0.036921 | 0.014543 | 0 | 0.031493 | 0.052861 | 4.247233 |

Dissolved Metals

| Storm Number | Sample/ Tray | Al ³⁺ (meq/L) | Fe ²⁺ (meq/L) | Mn ²⁺ (meq/L) | Zn ²⁺ (meq/L) | Cu ²⁺ (meq/L) | Cd ²⁺ (meq/L) | Ni ²⁺ (meq/L) | Si ²⁺ (meq/L) | ANC (meq/L) | |
|--------------|--------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|-------------|--|
| | ROCK = pyrite rock | | | | | | | | | | |
| 20 | ROCK | 0.298837 | 0.372814 | 0.044892 | 0.005779 | 0.002232 | 1.67E-05 | 0.004734 | 0.014475 | -0.34612 | |
| 22 | ROCK | 0 | 0 | 1.309378 | 0.156893 | 0.08059 | 0 | 0.131019 | 0.10611 | -18.8079 | |
| 23 | ROCK | 6.584245 | 0 | 0.63687 | 0.083356 | 0.050127 | 0 | 0.067541 | 0.073516 | -2.47634 | |
| 24 | ROCK | 2.304375 | 3.701551 | 0.374783 | 0.04179 | 0.020639 | 0 | 0.027504 | 0.07063 | 2.125978 | |
| 25 | ROCK | 1.013307 | 2.043483 | 0.240769 | 0.021305 | 0.008048 | 0 | 0.013345 | 0.052709 | 0.773986 | |
| 26 | ROCK | 0.681311 | 1.504239 | 0.183372 | 0.016893 | 0.007898 | 0 | 0.010896 | 0.029431 | 0.64671 | |
| | RAIN = rainwater | | | | | | | | | | |
| 1 | | | | | | | | | | | |
| 2 | | | | | | | | | | | |
| 3 | | | | | | | | | | | |
| 4 | | | | | | | | | | | |
| 5 | | | | | | | | | | | |
| 6 | | | | | | | | | | | |
| 7 | | | | | | | | | | | |
| 8 | | | | | | | | | | | |
| 9 | RAIN | 0.01234 | 0.035565 | 0.002982 | 0.000387 | 0.00176 | 0 | 0.000267 | 0.010326 | 0.434432 | |
| 10 | RAIN | 0.01439 | 0.03118 | 0.002273 | 0.000299 | 0.000286 | 0 | 0.000165 | 0.002397 | 0.101294 | |
| 11 | RAIN | 0.001528 | 0.00315 | 0.000359 | 0.000116 | 5.31E-05 | 0 | 3.13E-06 | 0.002902 | 0.147367 | |
| 12 | RAIN | 0.000538 | 0.000795 | 0.000377 | 0.000153 | 1.53E-05 | 0 | 3.03E-06 | 0.003675 | 0.082372 | |
| 14 | RAIN | 0.000503 | 0.000255 | 8.77E-05 | 0.000111 | 3.71E-05 | 0 | 0 | 0.001322 | 0.138872 | |
| 15 | RAIN | 0.000548 | 0.000741 | 0.000111 | 0.000149 | 0 | 0 | 0 | 0.002021 | 0.120993 | |
| 16 | RAIN | 0.002933 | 0.007982 | 0.000816 | 0.000163 | 3.83E-05 | 0 | 4.71E-05 | 0.002459 | 0.128653 | |
| 17 | RAIN | 0.009576 | 0.021032 | 0.002169 | 0.000218 | 6.61E-05 | 0 | 0.000152 | 0.002403 | 0.127352 | |
| 18 | RAIN | 0.00118 | 0.001215 | 0.000291 | 0.000135 | 4.48E-05 | 0 | 0 | 0.005803 | 0.120231 | |
| 19 | RAIN | 0.013797 | 0.016185 | 0.002893 | 0.000989 | 0.001091 | 0 | 8.17E-05 | 0.031667 | 1.210281 | |
| 20 | RAIN | 0.000514 | 0.000119 | 3.09E-05 | 0.0004 | 6.1E-05 | 1.7E-05 | 1.23E-05 | 0.005139 | 0.248023 | |
| 22 | RAIN | 0.002591 | 0.000288 | 3.36E-05 | 0.001233 | 0.000797 | 1.86E-05 | 1.01E-05 | 0.178732 | 3.494509 | |
| 23 | | | | | | | | | | | |
| 24 | RAIN | 8.61E-05 | 0.000412 | 0.000113 | 0.000332 | 3.69E-05 | 0 | 7.29E-06 | 0 | -0.00366 | |
| 25 | RAIN | 0.000268 | 0.000333 | 6.14E-05 | 0.000323 | 7.09E-05 | 1.81E-07 | 1.51E-05 | 0 | -0.00578 | |
| 26 | RAIN | 0 | 0.000369 | 0.000155 | 7.75E-05 | 2.23E-05 | 1.11E-06 | 1.05E-07 | 0 | 0.039117 | |